

Agricultural Water Abstraction Behaviour in Response to Policy and Climate Change

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Abstract

This thesis presents an assessment of farmers' behaviour in response to current changes in climate and policy in the UK. At present, many catchments in England are considered as over licensed or over abstracted and although water licence trading is regarded as a potential solution to the problem, and is currently possible, high transaction costs and institutional barriers deter farmers from trading. In response, two new water allocation systems have been proposed to provide farmers with the ability to adapt to climate and demand change pressures (i.e. basic and enhanced systems). A review of the current literature suggested farmers' behaviour very much influences the success of policy interventions. Therefore, this study sought to understand the behavioural intentions of farmers in England, and the underlying factors which drive their decision-making, under different climate and policy scenarios. Furthermore, this study examined whether farmers with different behavioural intentions lead to different patterns of abstraction behaviour at the system level, thus providing a means of assessing the current and proposed water allocation systems.

An empirical survey was conducted within the Great Ouse catchment in eastern England, UK, where freshwater availability for crop irrigation is considered highly vulnerable to climate change. The questionnaire, and subsequent interpretation of behaviours, was developed under the theoretical framework of the Theory of Planned Behaviour (TPB) (Ajzen, 1991). Farmers' preferred behavioural intentions were identified under different strategic (long-term) and in-season (short-term) water shortage and surplus scenarios. Furthermore, the TPB explained between 29-65 % of the variance in intention, based on Nagelkerke's R^2 , and was similar to the range found by meta-analytical reviews (i.e. 40-49 % based on R^2). In addition, attitude and subjective norm were found to be significant predictors of intention in three of the four scenarios. Overall, farmers believed they have greater volitional control with regards to decision-making in the long-term but less in the short-term. Furthermore, a behavioural farm typology based on farmers' preferred behavioural intentions was used in the development of an Agent-Based Model (ABM) to simulate system level patterns of abstraction behaviour which emerge from individual farm level decision-making. The scenario simulation results indicated the proposed enhanced water allocation system was likely to provide the greatest utility to balance the needs of licence users, at least farmers, whilst protecting the environment.

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List of Acronyms and Abbreviations

ABM	Agent-Based Model
ANOVA	Analysis of Variance
AP	Assessment Point
CAMS	Catchment Abstraction Management Strategies
CAP	Common Agricultural Policy
CO ₂	Carbon Dioxide
DEFRA	Department for Environment, Food and Rural Affairs
EA	Environment Agency
EFI	Environmental Flow Indicator
EU	European Union
FAO	Food and Agricultural Organisation
FL	Fully Licensed scenario
HOF	Hands Off Flow
HWUO	High Water Usage Options
IPCC	Intergovernmental Panel on Climate Change
IQR	Interquartile Range
IWN	Irrigation Water Need
LWUO	Low Water Usage Options
NQF	National Qualification Framework
NRW	Natural Resource Wales
ODD	Overview, Design concepts, and Details
Q30	High flow
Q50	Moderate flow
Q70	Below moderate flow
Q95	Low flow
RA	Recent Actual scenario
SPSS	Statistical Package for the Social Sciences
TIB	Theory of Interpersonal Behaviour
TPB	Theory of Planned Behaviour

TRA	Theory of Reasoned Action
UK	United Kingdom
UKIA	United Kingdom Irrigation Association
VIF	Variance Inflation Factor

1 Chapter 1

2 Introduction

3 This thesis presents an empirical investigation into farmer water abstraction behaviour
4 in response to policy and climate change in an area of eastern England, where supple-
5 mental irrigation is currently used to safeguard crop quality and to maintain expected
6 yields. The current water allocation system in England is considered by the regulatory
7 authorities as inadequate to provide users with the ability to adapt to climate change
8 whilst providing adequate protection for the environment. In response, the Department
9 for Environment, Food and Rural Affairs (DEFRA), who determine policy in England
10 and Wales, have proposed two new water allocation systems, basic and enhanced, which
11 aim to address the inadequacies of the current system. Despite important differences
12 between the proposed systems, the encouragement of water licence trading underpins
13 both. However, despite groups of stakeholders having been consulted, very little is
14 known whether either of these allocation systems will achieve their intended aims.
15 This thesis aims to address this issue and in doing so aid policy decision-makers in the
16 design and implementation of the proposed systems.

17

18 This chapter provides a general introduction to this thesis by discussing the im-
19 portance of this research within the wider context of water resource management and
20 climate change, with a particular focus on agricultural irrigation. Furthermore, it is
21 argued that the success of the proposed water allocation systems largely depend on
22 the response of users' on the ground. It is this unknown factor which underpins the
23 rationale for this research. Finally, this chapter highlights the main aim and subsequent
24 research questions, including an overview of the methodological approaches used, which
25 together form the main structure of this thesis.

1.1 Rationale of the study

1.1.1 A transition in freshwater resource management

Water is clearly an essential natural resource for life on this planet, yet it is also clear that human activities are directly threatening freshwater systems (Vörösmarty et al., 2010). In one of the first global syntheses investigating human and environmental perspectives on water security, Vörösmarty et al. (2010) reported on the level of impact that large-scale activities such as land cover change, urbanisation, industrialisation, as well as engineering projects such as reservoirs, irrigation and catchment transfers are having on the sustainability of these freshwater systems.

Although these human activities bring considerable benefits to society, they also bring unanticipated social, economical, and ecological costs (Gleick, 2003). For example: the construction of the Three Gorges Dam in China displaced more than one million people (World Commission on Dams, 2000); in 1999 27 % of all North American freshwater fauna populations were considered threatened with extinction (Ricciardi and Rasmussen, 1999); and many rivers around the world such as the Nile; Huang He (Yellow); and the Colorado, no longer maintain adequate environmental flows to their deltas during average years (Gleick, 2003).

Evidence gathered in the Fifth Assessment Report of the Intergovernmental Panel on Climate Change (IPCC) indicates that there is high agreement and robust evidence that freshwater related risks of climate change increase considerably with increasing greenhouse gas concentrations (Cisneros et al., 2014). In particular, for every degree Celsius of global warming approximately 7 % of the global population is projected to be exposed to a 20 % decrease in renewable freshwater resources. Furthermore, by the end of this century, under the Representative Concentration Pathway 8.5 scenario, climate change is likely to increase the frequency of droughts within the agricultural sector, due to less precipitation and increased evapotranspiration, in current dry regions such as southern and central Europe and the Mediterranean. Overall, climate change, coupled with increased demand and non-climatic human activities previously mentioned, are likely to exacerbate local variability within the hydrological system and increase the vulnerability of those areas to floods and droughts (Cisneros et al., 2014).

58 However, in the last few decades a transition in freshwater resource management
59 has occurred which attempts to tackle these unanticipated issues and potential risks
60 of climate change. This transition complements the large-scale, high-cost, centralised
61 infrastructure projects with low-cost, community-scale, decentralised systems. Largely
62 instigated by major water reform policies prescribed by the International Food Policy
63 Research Institute and the World Bank during the 1990s (Rosegrant and Binswanger,
64 1994) these transitional approaches include: technological investment at the farm and
65 catchment scale; encouragement of farmer abstraction groups to manage and improve
66 current allocation systems; and to apply economic tools such as water markets and
67 pricing. These water reforms aim to encourage: efficiency; productivity; and equitable
68 distribution with the long-term aim of providing a sustainable allocation system to
69 meet the demands of users without detriment to the environment (Gleick, 2003).

70

71 **1.1.2 Experiences of water markets**

72 The transition from large-scale, centralised, infrastructure projects to achieve continu-
73 ous supplies of water to meet demand for small-scale, community based, water market
74 mechanisms has struggled to achieve its intended aim (Garrick et al., 2013). Australia,
75 Chile, South Africa, and the United States have encouraged and permitted water rights
76 trading to improve allocation efficiency (Erfani et al., 2015). However, these mature
77 water markets have experienced major problems ensuring sustainable environmental
78 flows, with governments in Australia and the United States having to buy back water
79 entitlements to protect the environment. In Australia, the government is in the middle
80 of a \$3.1 billion programme to buy back water for the environment (Wheeler et al.,
81 2013). As a result of these unanticipated costs, Australia, South Africa and the United
82 States, as well as emerging water markets in countries such as China (Zhang et al.,
83 2014) and parts of the UK (DEFRA, 2016c), are in the process of reforming water
84 allocations to improve efficiency and protect the environment (Young, 2014).

85

86 In China, where water is readily available in the south but increasingly scarce in the
87 north, a pilot project in the city of Zhangye found that the main reasons farmers were
88 discouraged from saving water and trading surplus were due to high transaction costs in
89 some areas and where transaction costs were low reasons included: management; legal;

90 administrative; and fiscal barriers (Zhang, 2007; Zhang et al., 2009, 2014). In parts of
91 the UK similar institutional barriers which have discouraged water trading have been
92 recognised, which restrict the ability to increase water allocation efficiency and also
93 to meet water quality standards set by the European Union (EU) Water Framework
94 Directive (EA, 2008). The same barriers to successful water markets are apparent in
95 many developing countries (Thobani, 1998).

96

97 Garrick et al. (2013) supported these findings and summarised water markets as
98 systems which involve two factors: first, a system which balances demand for human
99 requirements and demand for sustainable flows to protect the environment; and second,
100 a system which establishes tradeable water rights to allocate water within and across
101 productive sectors. The experiences of mature water markets in Australia, Chile, South
102 Africa, and the United States indicate that the current policies in these countries do not
103 fully protect the environment and further institutional reforms are required (Bjornlund
104 and McKay, 2002). Therefore, a potential framework for the successful implementation
105 of a water market should include: a multiphase sequencing of reform; strategic invest-
106 ment in institutional tutorial transition costs; and institutional choices that preserve
107 future flexibility to adjust water rights and diversion limits to manage social and envi-
108 ronmental externalities (Garrick et al., 2013).

109

110 However, institutional barriers are not the only obstacle in successfully implement-
111 ing a water market or associated allocation mechanisms. The success of any new al-
112 location mechanism is also largely dependent on the actual users whom the allocation
113 reforms will directly affect. Several studies have shown that farmers' behaviour very
114 much influences how and to what success policy proposals are realised on the ground
115 (Moon and Cocklin, 2011; Home et al., 2014; Feola et al., 2015). Furthermore, Feola
116 et al. (2015) suggested that understanding farmers' behaviour is essential to identifying
117 where policy intervention may be required, or should be avoided, as well as where it
118 can be used to inform the design and implementation of reforms.

119

120 Bjornlund (2003) examined farmer participation in water markets in southeastern
121 Australia during the first ten years of operation starting in 1989. The main findings
122 of the study indicated that an increasing number of crop irrigation abstractors partic-

123 ipated in temporary licence trades (i.e. weekly, monthly, or seasonal) as more farmers
124 increasingly understood the operations and the advantages of a secure, reliable, fast,
125 and cheap water transfer. However, very few farmers participated in permanent wa-
126 ter licence trading due to: differential tax treatment; policy uncertainty; institutional
127 barriers such as administrative complexity and cost associated with permanent trades;
128 and crop irrigation abstractors' perception that water rights are an inherent part of
129 their property. The results of this study, and those previously mentioned suggest that
130 institutional barriers and high transaction costs deter farmers from trading; and im-
131 portantly the success of any new allocation method is largely dependent on the level of
132 participation by the users on the ground.

133

134 1.1.3 Agricultural irrigation and climate change

135 The use of freshwater for crop production by irrigation from rivers, reservoirs, lakes,
136 groundwater aquifers and inter-catchment transfers are all considered sources of blue
137 water; precipitation in regards to crop production is considered a green water source
138 (Rost et al., 2008). Agriculture in general is responsible for approximately 85 % of total
139 global water consumption of both blue and green freshwater resources (Shiklomanov
140 et al., 2003). Crop irrigation accounts for the largest human withdrawal and consump-
141 tion (i.e. withdrawal minus return flow to the river) of blue water globally, accounting
142 for almost 70 % of this resource (Shiklomanov et al., 2003). However, to put this in
143 perspective, 80 % of global croplands are rainfed, from which 60-70 % of the world's
144 food is produced (Falkenmark and Rockström, 2004). In many places, such as parts of
145 the UK, blue water is used to supplement green water to improve crop quality and not
146 solely to maintain expected yields.

147

148 Crop irrigation ultimately relies on freshwater resources being available for abstrac-
149 tion. Whether a farmer abstracts the water and to what extent they use that water for
150 crop irrigation is a human behavioural factor regarding crop management. Nonetheless,
151 crop water requirement is governed by a balance between atmospheric moisture deficit
152 and soil water supply. Therefore, any change in climate (i.e. precipitation, tempera-
153 ture, or radiation) will affect crop water requirements and also the amount of water
154 available for abstraction at the catchment scale (Cisneros et al., 2014).

155 Although the IPCC Fifth Assessment Report 2008 indicated that global trends in
156 precipitation from several datasets during the last century were statistically insignif-
157 icant (Bates et al., 2008), regional observations indicate that the most severe floods
158 and droughts since the 1950s have been experienced between 1990 and 2010 (Arndt
159 et al., 2010). The majority of regional changes in precipitation are considered to be
160 attributable to either internal variability regarding atmospheric circulation or to global
161 warming (Lambert et al., 2004; Stott et al., 2010). Despite the uncertainty regarding
162 the climatic drivers Zhang et al. (2007) estimated that twentieth century anthropogenic
163 forcing contributed considerably to observed changes in global and regional precipita-
164 tion. In addition, there has been a continual decrease in global and regional evapo-
165 transpiration since the middle of the last century which has been attributed to changes
166 in precipitation; diurnal temperature range; aerosol concentration; net solar radiation;
167 vapour pressure deficit; and wind speed (Fu et al., 2009; McVicar et al., 2010; Miralles
168 et al., 2011; Wang et al., 2011). All of these factors greatly influence crop water re-
169 quirements (Allen et al., 1998) particularly as precipitation and evaporation are the
170 main climatic drivers controlling freshwater resources (Cisneros et al., 2014).

171

172 A series of different projections regarding changes in irrigation demand and wa-
173 ter availability, based on different global climate models and greenhouse gas emission
174 scenarios, report a high confidence that irrigation demand will increase by more than
175 40 % across Europe, United States, and parts of Asia by 2080 (Cisneros et al., 2014).
176 However, some of these projections also indicate that irrigation demand could decrease
177 in parts of India, Pakistan, and other parts of Asia due to an increase in precipitation.
178 In addition, where poor soil is not a limiting factor, the physiological and structural re-
179 sponse of crops to elevated atmospheric carbon dioxide (CO₂) concentration (i.e. CO₂
180 fertilisation) might alleviate some of the adverse effects of climate change by reducing
181 irrigation demand (Konzmann et al., 2013). Overall, it is projected that irrigation de-
182 mand will exceed local freshwater availability in many areas (Wada et al., 2013).

183

184 1.1.4 The UK transition

185 In regards to the UK, DEFRA determine policy in England and Wales. These policies
186 are regulated by the Environment Agency (EA) and Natural Resource Wales (NRW) in

Table 1.1: Abstractions licences in England and Wales (2014)

Sector	Licences (%)	Licensed volume (%)	Licensed volume used (%)
Public water supply	8	18	60
Agriculture (spray irrigation only)	48	1	26
Agriculture (excluding spray irrigation)	14	<1	27
Electricity supply industry	3	69	27
Other industry	18	9	40
Fish farming, cress growing, and amenity ponds	3	3	54
Private water supply	5	<1	24
Other	1	<1	13
Total	100	100	35

Source (DEFRA, 2016a)

187 England and Wales respectively. Anyone wishing to abstract more than $20 \text{ m}^3 \text{ day}^{-1}$ of
188 water requires an abstraction licence. Table 1.1 highlights the percentage of: licences
189 allocated; total volume allocated; and licensed volume abstracted per sector in Eng-
190 land and Wales in 2014 (DEFRA, 2016a). Overall, only 35 % of the volume licensed
191 to be abstracted was actually used, but this was dependent on water availability in
192 addition to changes in demand. Furthermore, although the electricity supply industry
193 accounted for the majority of licensed volume, this sector is generally considered as a
194 non-consumptive user as nearly all of the water it abstracts is returned to the catch-
195 ment. Other non-consumptive sectors include fish farming, cress growing, amenity
196 ponds, and some other industry. Therefore, in regards to consumptive users, the public
197 water supply sector was by far the largest abstractor (i.e. 18 % of licensed volume)
198 with spray irrigation licence users only accounting for 1 % of total licensed volume.

199

200 The largest proportion of irrigated agriculture is concentrated in eastern England
201 (i.e. the Anglian region) employing over 50,000 people and contributing £3 billion an-
202 nually to the region's economy (Leathes et al., 2008). Furthermore, in 1995 50 % of
203 the total irrigated land in England and Wales was located in the Anglian region (Knox
204 et al., 2000) and in 2014 the area accounted for 36 % of all spray irrigation licences and
205 62 % of estimated spray irrigation abstractions (DEFRA, 2016a). In addition, Knox
206 et al. (2010) reported that since 1990 there has been a marked increase in irrigation of

207 high value crops such as potatoes and field vegetables, largely driven by supermarkets
208 demand for quality; consistency; and continuity of supply, which can only be guaran-
209 teed by irrigation. However, this region is one of the driest and most water stressed
210 regions in the UK, with nearly three quarters of the water licensed for irrigation in 2006
211 located in catchments under severe levels of water stress according to the UK Irrigation
212 Association (UKIA) (i.e. with 47 % over-licensed; 23 % over-abstracted; and 21 % with
213 no water available for further licensing) (UKIA, 2007). Furthermore, several studies
214 have shown that freshwater available for irrigation in this region will be vulnerable to
215 projected changes in climate (Gowing and Ejieji, 2001; Gibbons and Ramsden, 2008;
216 Henriques et al., 2008; Knox et al., 2009).

217

218 Therefore, the need to improve efficiency at the farm and catchment scale has be-
219 come increasingly important as water availability has become increasingly competitive
220 between users and the environment. Several initiatives have previously been imple-
221 mented towards alleviating some of this pressure such as: restoring sustainable abstrac-
222 tions (EA, 2010); and requiring spray irrigation licence users to demonstrate efficiency
223 as part of their licence application process (Knox et al., 2012). However, DEFRA and
224 the EA consider the current water allocation system in England as inadequate to pro-
225 vide users with the adaptive strategies they require to meet their needs whilst providing
226 sufficient protection to the environment (EA, 2008; DEFRA, 2011).

227

228 Consequently, DEFRA have proposed two new water allocation systems, basic and
229 enhanced, which aim to address the inadequacies of the current system, and which are
230 to be implemented by the early 2020s (see Table 1.2). Although both of the proposed
231 systems encourage water licence trading by removing seasonal restrictions (i.e. licences
232 which can only be used during certain months) and time limited licences (i.e. licences
233 which expire after a certain number of years), the enhanced system offers greater incen-
234 tives for trading. These include access to ‘bonus water’ (i.e. water which is available for
235 abstraction during high flow periods in addition to licensed volumes); gradual imple-
236 mentation of Hands off Flow (HOF) conditions rather than simply on or off as with the
237 current and basic systems (i.e. when the EA stop users from abstracting water during
238 certain low flows in order to protect the environment or other users); and pre-approved
239 trading rules to increase the speed and ease of trading.

Table 1.2: Current and proposed water allocation systems for England

Measure	Current allocation system	Proposed allocation systems	
		Basic	Enhanced
Abstraction $>20 \text{ m}^3 \text{ day}^{-1}$	Licence required	Permit required	Permit required
Daily, annual, and return licence limits	Yes (fixed)	Yes (fixed)	Yes (share)
Seasonal restrictions	Yes	No	No
Time limited licences	Yes	No	No
Hands off Flow (HOF) conditions for some licences	Yes	Yes	Yes
Low flow conditions for those without HOF conditions	No	Yes	Yes
‘Bonus water’ during high flows in addition to licence	No	No	Yes
Water licence trading	Yes	Yes	Yes
Pre-approved trades	No	No	Yes

Source (DEFRA, 2016c)

240 These additional incentives for water licence trading, under the enhanced system,
 241 are made possible as annual licence limits would not be fixed, as they are with the
 242 current and basic systems, but accounted for as shares of available water within a
 243 catchment or sub-catchment. The type of proposed system to be implemented depends
 244 on whether DEFRA consider a catchment as basic or enhanced. Classification of catch-
 245 ments is still ongoing but the main determining factor is likely to concern the potential
 246 for water licence trading related to the cost of implementing an enhanced catchment, as
 247 users are required to have smart meters installed in order for the regulatory authorities
 248 to calculate water use and availability accurately (DEFRA, 2016c).

249

250 Furthermore, licences will become permits under the proposed systems and consist
 251 of three parts: water account conditions; site specific conditions; and standard catch-
 252 ment rules. The water account conditions consist of daily, annual, and return licence
 253 limits and are stored online to facilitate trading. Site specific conditions relate to any
 254 conditions which are unique to a particular licence and for most users will be trans-
 255 ferred from their existing licence. Standard catchment rules will apply to everyone and
 256 relate to: common HOF conditions; trading rules; and low flow conditions. The latter
 257 are restrictions which can be imposed by the regulatory authority for those without
 258 HOF conditions, in order to protect the environment and other users.

1.2 The need for further research

England is currently in the process of a major transition in water resource management. The proposed water allocation systems are designed to address the inadequacies of the current system, largely by encouraging water licence trading in order to increase efficiency, whilst providing adequate protection to the environment. These issues are particularly apparent in regards to irrigation within the Anglian region, in eastern England. The success of the proposed water allocation systems, at least in regards to increasing efficiency within the irrigation sector, is very much dependent on how likely users are to engage with the new systems, in particular to trading. Although experiences of water markets elsewhere, and consultations with some users in the UK (DEFRA, 2014b, 2016b), may provide some understanding of potential engagement, no formal scientific approach has been made to attempt to measure, or understand farmers' behaviours with regards to the proposed water allocation systems.

Therefore, further research is required to understand farmers' behaviours, under different scenarios of water shortage and surplus, with regards to the proposed water allocation systems. In addition, an investigation which aims to understand whether a farm typology based on farmers' behaviours, rather than traditional economic or physical descriptors, could provide a valuable tool for policy decision-makers involved with the design and implementation of the proposals. Finally, an understanding of the potential system level patterns of abstraction behaviour which could emerge from individual farm scale decisions under the proposed water allocation systems is essential. This research intends to inform policy decision-makers involved with the proposals and, more generally, assess the potential effectiveness of the transition in freshwater resource management in England by addressing these issues.

1.3 Aim, research questions and methodological overview

The main aim of this thesis was to understand farmers' behavioural intentions under different climate and proposed water allocation system scenarios, in order to assess whether the proposals will achieve their intended aim. Three research questions were formulated to address this main aim.

290 **Research question 1:** *What are farmers' preferred behavioural intentions under*
291 *different scenarios of water shortage and surplus with regards to the proposed water*
292 *allocation systems in England? And which underlying predictors of intention most in-*
293 *fluence their decision-making?*

294

295 As previously mentioned, and discussed in more depth in the following chapter,
296 there is a clear need to understand farmers' preferred behavioural intentions as this
297 heavily influences the success of the proposals at the ground level. Therefore, an em-
298 pirical investigation was conducted, in a selected study area in the Anglian region, with
299 the main aim of understanding farmers' preferred behavioural intentions under different
300 scenarios of water shortage and surplus with regards to the proposed water allocation
301 systems. The investigation also collected data concerning farm attributes and social
302 demographics. The questionnaire, and subsequent interpretation of behaviours, was
303 developed under the theoretical framework of the Theory of Planned Behaviour (TPB)
304 (Ajzen, 1985).

305

306 **Research question 2:** *If farmers share similar behavioural intentions under dif-*
307 *ferent scenarios how can traditional farm typologies incorporate these preferred be-*
308 *haviours?*

309

310 Traditional farm typologies tend to concentrate on particular farm attributes such
311 as farm income, size, or output rather than on preferred behavioural intentions. This
312 study presents a farm typology based on the preferred behavioural intentions of farmers
313 identified. Furthermore, statistical analyses were conducted to examine similarities and
314 differences between farm types in regards to: farm attributes; social demographics; li-
315 cence characteristics; and past abstractions from 2008 to 2012. This study offers one of
316 the first typologies designed for policy decision-makers to understand which farm types
317 are more likely to engage with the proposed water allocation systems and therefore aid
318 in deciding which catchments could be considered basic or enhanced.

319

320 **Research question 3:** *What potential system level patterns of abstraction be-*
321 *haviour emerge from individual farm scale decisions under different policy and climate*
322 *scenarios?*

323 This final research question addresses the challenge of developing a model to assess
324 potential patterns of catchment level water abstraction behaviour which emerge from
325 individual farm scale decisions under different water allocation systems (i.e. current
326 and proposed) and climate scenarios. While sophisticated hydrological models (Gosling
327 et al., 2011) and crop models exist (Allen et al., 1998), few successfully integrate farmer
328 decision making (Feola et al., 2015) and none do so in the context of the proposed water
329 allocation systems in England. Therefore, this study presents the development of an
330 Agent-Based Model (ABM) that utilises the farm typology that has been developed.
331 The model is capable of modelling the preferred behaviours of individual farmers and
332 exploring what system level patterns of abstraction behaviour emerge under the differ-
333 ent policy and climate scenarios. Furthermore, particular farm indicators are used to
334 present the outputs from the model and measure the system level patterns of abstrac-
335 tion behaviour.

336

337 1.4 Thesis structure

338 This chapter has provided a general introduction outlining the main rationale and ar-
339 eas for further research which this thesis aims to address. Chapter 2 presents a critical
340 review of the main bodies of literature regarding measuring individual behaviour; incor-
341 porating preferred behaviours into farm typologies; and modelling individual behaviour
342 at the farm and catchment scales. As a result, Chapter 2 highlights the main research
343 gaps this thesis aims to address. Chapter 3 introduces the study area which was the
344 focus of this investigation. Chapter 4, 5, and 6, are associated with each of the research
345 questions previously discussed and therefore present the methodologies and results of
346 these individual studies. Finally, Chapter 7 presents a discussion of the main results
347 of this research placing the findings within the broader context of agricultural water
348 resource management and climate change in the UK. Furthermore, this final chapter
349 presents the main conclusions and policy applications of this research along with some
350 of the limitations of the methods and potential areas for further research.

351

352 Chapter 2

353 Literature review

354 The previous chapter discussed the proposed water allocation systems which are to be
355 introduced in England. In particular, Chapter 1 discussed the importance, from a pol-
356 icy decision-making perspective, of understanding how users on the ground (i.e. those
357 who the proposals are going to directly effect) are likely to respond and engage with
358 the proposed water allocation systems. That is, at least with the parts that they will
359 have control over such as deciding whether or not to trade, rather than catchment rules
360 which are enforced to protect the environment from over abstraction. Furthermore, this
361 thesis is particularly interested in how the proposed water allocation systems will effect
362 irrigation farmers, and therefore the term term farmers refers to spray irrigation licence
363 users only, not farmers who abstract for other purposes (see Table 1.1). This chapter
364 critically reviews the main bodies of literature related to each of the three research
365 questions presented previously (see Section 1.3). In particular, this chapter examines:
366 farmers' behaviours with regards to behavioural approaches, limitations, and exten-
367 sions; farm typologies including limitations and incorporating behaviours; and lastly
368 approaches used to model farmers' behaviours, and how these models can be used to
369 inform policy decision-making. Finally, this chapter highlights the main research gaps
370 this thesis aims to address.

371

372 2.1 Farmer behaviour

373 The behavioural approach within the fields of agricultural and policy decision-making
374 is not a recent concept and is thought to have originally begun with the 'satisficing'
375 concept during the 1950s (Simon, 1957). This concept acknowledged that people do not

necessarily indulge in economically optimal decision-making but instead may optimise social, intrinsic, or expressive goals. Despite the growing necessity and increasing application of the behavioural approach over the following decades, particularly in policy driven studies (Wolpert, 1964; Gasson, 1973), it has often been regarded as an additional component of rational, or neo-classical economic models (Beedell and Rehman, 2000), rather than a separate entity partly because the main criticism of it is that it overemphasises the role that attitude has on determining behaviour and partly due to the lack of any real theoretical foundation (Burton, 2004). However, with advancements in the field of social psychology, such as the Theory of Planned Behaviour (TPB) (Ajzen, 1985), the number of studies incorporating farmer decision-making rapidly increased (Burton, 2004; Poppenborg and Koellner, 2013) providing a practical and theoretical basis for the growing amount of research on farmers' attitudes and intentions (Beedell and Rehman, 2000).

Morris and Potter (1995) provided a valuable definition of behavioural approaches in agricultural studies as those that: a) seek to understand the behaviour of individual decision-makers, usually the farmers or land managers, directly responsible for the land; b) focus on psychological constructs such as attitudes, values, and goals but also commonly gather additional relevant data such as farm structure, economic situation, successional status; and c) employ largely quantitative methodologies, in particular psychometric scales such as Likert-type scaling procedures (Likert, 1932), for investigating psychological constructs such as those of the TPB.

2.1.1 The Theory of Planned Behaviour (TPB)

The TPB is an extension of the Theory of Reasoned Action (TRA) (Fishbein and Ajzen, 1975) and was required due to the original model's limitations in dealing with behaviours where individuals fail to have complete volitional control in regards to performing, or not performing, a particular behaviour (Ajzen, 1991). Furthermore, it is one of the most frequently cited and influential models for the prediction of human social behaviour (Ajzen, 2011) and has regularly been supported with empirical evidence (Armitage and Conner, 2001). A central factor in the theory is an individuals' intention to perform a given behaviour. Intentions are assumed to capture the motivational

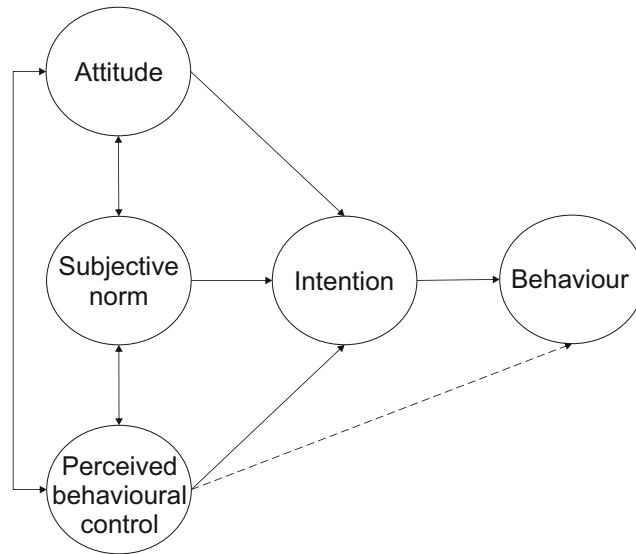


Figure 2.1: The Theory of Planned Behaviour (adapted from Ajzen (1991))

408 factors that influence a behaviour and therefore the stronger the intention the more
409 likely the behaviour will be performed.

410

411 Figure 2.1 illustrates the main constructs of the TPB with the direct antecedents
412 of intention including attitude (i.e. an individuals' overall positive or negative evalu-
413 ation of performing, or not performing, a particular behaviour); subjective norm (i.e.
414 the social pressure an individual might feel towards performing, or not performing, a
415 particular behaviour); and perceived behavioural control (i.e. the degree of volitional
416 control an individual feels they have over performing, or not performing, a particular
417 behaviour). Consequently, the strength of an individuals' intention to perform, or not
418 perform, a particular behaviour depends on the level, or strength, of that individuals'
419 attitude; subjective norm; and perceived behavioural control (Armitage and Conner,
420 2001). Furthermore, the constructs of attitude; subjective norm; and perceived be-
421 havioural control are based on associated salient behavioural; normative; and control
422 beliefs about the particular behaviour, respectively (Ajzen, 1991).

423

424 The relative importance of these three constructs in regards to predicting intention
425 varies depending on the type of behaviours and nature of the situation (Ajzen, 1991).
426 Furthermore, perceived behavioural control is also considered a direct antecedent of

427 behaviour based on the rationale that however strong an individuals' intention is to
428 perform, or not perform, a particular behaviour, the actual performance of said be-
429 haviour can depend on personal and environmental barriers. Therefore, this direct re-
430 lationship between perceived behavioural control and actual behaviour becomes more
431 important as volitional control decreases (Ajzen, 1991), to the extent that perceived
432 behavioural control should be able to directly predict actual behaviour (Armitage and
433 Conner, 2001). Conversely, where an individual has complete volitional control per-
434 ceived behavioural control should have very little or no influence on their intention to
435 perform, or not perform, a particular behaviour.

436

437 Poppenborg and Koellner (2013) investigated whether attitudes toward ecosys-
438 tem services determine agricultural land use practices. The study examined farmer's
439 decision-making processes, based on the TPB, with respect to land use in a South Ko-
440 rean watershed. Decisions between a variety of cropping patterns and practices were
441 compared among farmers as a function of their attitude towards particular ecosys-
442 tem services including: biomass production; prevention of soil erosion; improvement
443 of water quality; and conservation of plants and animals. The study found that deci-
444 sions to plant perennial crops are most often accompanied by positive attitudes toward
445 ecosystem services whereas no differences were found between organic and conventional
446 farming. Furthermore, positive attitudes toward ecosystem services were most likely
447 held by farmers with high income indicating that a farmer's financial means were a key
448 determinant of farmers' environmental attitudes.

449

450 Similarly, Wauters et al. (2010) investigated the adoption of soil conservation prac-
451 tices of farmers in Belgium based on the TPB. The results of the study showed that
452 the most important explanatory factor was attitude towards soil conservation prac-
453 tices. They concluded that future interventions directed at promoting erosion control
454 measures should be directed towards changing farmers attitudes, although further in-
455 vestigation was required to understand the negative attitudes of farmers.

456

457 **2.1.2 Limitations of the TPB**

458 Despite the popularity of the TPB, particularly within the health sciences (Sniehotta
459 et al., 2014), it has also had considerable criticism most notably concerning efficacy (i.e.
460 the predictive ability of the model) and sufficiency (i.e. the assumption that the effect
461 of all other biological, social, environmental, economic, cultural, and unconscious in-
462 fluences are hypothesised to be accounted for by the TPB constructs) (McCarty, 1981;
463 Hardeman et al., 2002; Chatzisarantis et al., 2005; Sniehotta et al., 2014).

464

465 In regards to efficacy, several meta-analytical reviews reported that although the
466 TPB explained between 40-49 % of the variance in intention, it only explained be-
467 tween 26-36 % of the variance in actual behaviour (Sheeran et al., 1999; Armitage
468 and Conner, 2001; Hagger et al., 2002; Trafimow et al., 2002; Schulze and Wittmann,
469 2003; McEachan et al., 2011). This common occurrence has been described as the
470 intention-behaviour gap (Sniehotta et al., 2005), or the issue of inclined abstainers (i.e.
471 despite an individual's intention they fail to act out the behaviour) (Orbell and Sheeran,
472 1998). The intention-behaviour gap indicates low correlations between the TPB con-
473 structs and therefore a failure to account for the majority of variability in behaviour
474 (Sniehotta et al., 2014).

475

476 Ajzen (2011) argued that even with carefully considered measures (i.e. well de-
477 signed salient beliefs), the most one can reasonably expect in regards to correlations
478 between constructs are coefficients of approximately .60 (i.e. the weaker the correlation
479 the less the constructs can explain the variance in intention and therefore behaviour).
480 This statement corresponds with several meta-analytical reviews, for example, multi-
481 ple mean correlations between attitudes, subjective norms, and perceived behavioural
482 control with intention ranged from .59 to .66 (Cheung and Chan, 2000; Armitage and
483 Conner, 2001; Schulze and Wittmann, 2003); and .53 between intention and behaviour
484 (Sheeran, 2002). Furthermore, several studies have shown that the predictive ability
485 of the TPB model is often reduced depending on: the length of the study (i.e. the
486 predictive ability of the model decreased as the length of the study increased due to ex-
487 ternal events intervening and changing individuals' salient beliefs) (Sheeran et al., 1999;
488 Conner et al., 2000); and whether measures were taken objectively or self reported (i.e.
489 the predictive ability of the model decreased where outcome measures were measured

objectively) (Ajzen, 2011; McEachan et al., 2011). Finally, Ajzen (2011) suggested that where studies which occurred over a short time period reported a large intention-behaviour gap (Kor and Mullan, 2011), the cause was often due to actual control not being accurately represented by perceived behavioural control. In addition, intention-behaviour gap has also been attributed to behaviours which require different social or psychological drivers, such as efficiency and curtailment behaviours (i.e. behaviours which are one-off or occur regularly respectively) (Gardner and Stern, 1996; Russell and Fielding, 2010). However, beyond methodological failures and different types of behaviours being assessed, a low intention-behaviour relation could simply indicate the limits of this particular behavioural approach (Ajzen, 2011).

500

In regards to sufficiency, several studies have criticised the TPB model for its exclusive focus on rational reasoning (Sniehotta et al., 2013) and ignoring unconscious influences such as habits (Conner and Armitage, 1998; Abraham and Sheeran, 2003; Gardner et al., 2011); affects; emotions; and irrationality (Sutton, 1994; Conner and Armitage, 1998; Richard et al., 1998; Rapaport and Orbell, 2000; McEachan et al., 2011; Wolff et al., 2011; Carraro and Gaudreau, 2013; Conner et al., 2013; Sheeran et al., 2013). However, Ajzen (2011) argued that there is no assumption in the TPB that the salient beliefs are formed in a rational or unbiased fashion, and in fact are more likely to be irrational or inaccurate as the individual beliefs about a behaviour are often incomplete or self serving. Therefore, regardless of being rational or irrational, attitudes, subjective norms, and perceived behavioural control follow on from beliefs and it is only in this sense that behaviour is said to be planned (Geraerts and McNally, 2008).

514

515 **2.1.3 Adding past behaviour as a predictor variable**

Nonetheless, several studies, predominantly within the health sciences, provided evidence to suggest that additional predictor constructs, such as habit which was used within Triandis' Theory of Interpersonal Behaviour (TIB) (Triandis, 1979), can and have been used to address both of the main limitations (Conner and Armitage, 1998). Additional predictor variables have included: past behaviour (Kor and Mullan, 2011; Norman and Cooper, 2011); similarity (Rivis et al., 2011); uncertainty avoidance (Wolff

522 et al., 2011); and self-concept (Hassandra et al., 2011). The possibility of adding addi-
523 tional predictors, or constructs, was in fact originally suggested (Fishbein and Ajzen,
524 1975; Ajzen, 1991) which led to the updated TPB from the TRA by adding perceived
525 behavioural control. However, Ajzen (2011) argued that the addition of further predic-
526 tors, or constructs, should be made with caution and after empirical exploration.

527

528 Adding past behaviour as a means of improving the predictive measure of intention,
529 has been supported by considerable empirical evidence (Conner and Armitage, 1998;
530 Abraham and Sheeran, 2003; Ajzen, 2011; Kor and Mullan, 2011; Norman and Cooper,
531 2011). For example, in the study by Abraham and Sheeran (2003), the amount of
532 variance in physical activity explained by the TPB increased from 36 % to 53 % with
533 the addition of past physical activity. However, Ajzen (2011) argued that past be-
534 haviour should not be included as an additional construct since it constitutes a causal
535 antecedent of intention, and therefore fails to meet the original criteria for the TPB.
536 In particular, it is difficult to argue that the performance of a behaviour in the past
537 directly causes an individual's current intention; rather it is usually considered a proxy
538 for habit when performed routinely in a stable environment (Ajzen, 2011; Norman and
539 Cooper, 2011).

540

541 Nevertheless, in three meta-analytic syntheses the addition of past behaviour raised
542 the proportion of explained variance in intention by between 10 % and 13 % (Albar-
543 racin et al., 2001; Sandberg and Conner, 2008; Rise et al., 2010). Ajzen (2011) explained
544 that one possible conclusion to this is that intentions may not only be determined by
545 attitudes, subjective norms, and perceived behavioural control but by one or more
546 additional variables, and these additional variables are captured, at least partly, by
547 measures of past behaviour.

548

549 **2.2 Farm typologies**

550 The use of farm typologies, particularly with regards to policy decision-making, is not
551 a new concept and has been used for several decades (Kostrowicki, 1977; Whatmore
552 et al., 1987; Landais, 1998; Andreoli and Tellarini, 2000; Daskalopoulou and Petrou,
553 2002; Köbrich et al., 2003; Tavernier and Tolomeo, 2004). Most commonly they are

554 used to assist policy decision-makers with the design, implementation, or assessment
555 of the performance of agricultural or environmental schemes at the farm and regional
556 scales (Andersen et al., 2007).

557

558 Farm typologies are often designed based on some type or combination of farm char-
559 acteristic, such as: farm size; income; capital; labour; production pattern; soil quality;
560 managerial ability; or output. A farm is the key level at which decisions are made in
561 relation to the management of farmland and natural resources, and a typology offers
562 policy decision-makers a tool to assess farms by these particular criteria or indicators,
563 as well as providing a means to better understand the underlying drivers behind farm
564 management decisions (Andersen et al., 2007).

565

566 The EU farm typology, for example, is used to inform policy decision-makers about
567 multiple agricultural issues such as the performance of particular schemes linked with
568 the EU Common Agricultural Policy (CAP) (Andersen et al., 2007). However, the ra-
569 tional behind the EU farm typology is exclusively economic, with the main distribution
570 of farms categorised by farm income. Although farm income can be argued to be one
571 of the main drivers behind farm management decisions and policies, CAP reforms are
572 no longer simply about production and economy; rather they have shifted towards the
573 environment and landscape. Andersen et al. (2007) therefore suggested that the EU
574 farm typology should be updated to include an environmentally based extension, which
575 would classify farms into farm types that are more homogeneous in their environmental
576 performance rather than simply by their income.

577

578 **2.2.1 Limitations of farm typologies**

579 Although farm typologies clearly offer policy decision-makers a useful tool for categoris-
580 ing farms, and making more informed policy decisions, their main criticism is that they
581 fail to fully capture true farmer behaviour and are therefore of limited value in support-
582 ing policy formulation (Guillem et al., 2012). In regards to the EU farm typology, the
583 suggestion proposed by Andersen et al. (2007) to include environmental performance
584 indicators, relied on further statistical measures of environmental pressures caused by
585 farm management practices, rather than any assessment of true farmer behaviour.

586 Therefore, to address this main criticism, and partly due to the lack of success
587 or changes to such environmental policy schemes (Brotherton, 1991; Morris and Pot-
588 ter, 1995; Wilson, 1996; Winter, 2000; Burton, 2004), policy decision-makers realised
589 they required more sophisticated methods of anticipating farmers motivation to comply
590 with new policy approaches, thus, providing a catalyst for research in this field (Austin
591 et al., 1998; Burton, 2004; Emtage et al., 2006; Guillem et al., 2012). In addition,
592 Emtage et al. (2006) reviewed a number of studies which used landholder typologies
593 in the development of natural resource management programmes in Australia. They
594 concluded that policy decision-makers needed to: better understand the variation in
595 socio-economic circumstances and values of individuals; understand how this variation
596 affects their attitudes and behaviour; and how the differences lead to variation in the
597 impacts of policy reforms across the community or target area.

598

599 **2.2.2 Incorporating decision-making into farm typologies**

600 As previously discussed, over the last few decades considerable research has focussed
601 on understanding farmer's intentions and behaviours (Austin et al., 1998; Garforth and
602 Rehman, 2006; Gorton et al., 2008; Barnes et al., 2009; Sutherland et al., 2011) by
603 linking behavioural theories from the field of social psychology with the field of agri-
604 cultural decision-making (Burton, 2004). Guillem et al. (2012) indicated that studies
605 which have incorporated farmer decisions have improved the relevance of policy formu-
606 lation and have been a major factor in increasing the successful development and use of
607 farmer typologies (Schmitzberger et al., 2005; Emtage et al., 2006, 2007; Gorton et al.,
608 2008; Pike, 2008; Barnes et al., 2011; Poppenborg and Koellner, 2013).

609

610 There are two main approaches used to develop farm typologies based on a range
611 of farm characteristics: the data-driven approach; and the conceptual approach. The
612 data-driven approach often utilises multi-variate techniques, such as cluster analysis,
613 to search for patterns in the data which can be identified as farm types (Köbrich et al.,
614 2003; Acosta et al., 2014). The conceptual approach usually pre-defines the types based
615 on a particular research aim and subsequently conducts statistical tests to search for
616 any similarities or differences in farm characteristics (Phillip et al., 2010).

617

2.3 Modelling farmer behaviour

Traditional neo-classical economic models have for many years been used to model farmer decision-making at different spatial scales (Willock et al., 1999; Edwards-Jones, 2006). Whilst these types of models provided a useful tool for policy decision-making, they have also received criticism for assuming individuals only make rational economic decisions, whilst in reality, individual decisions are often based on a combination of psychological constructs such as attitude, subjective norms, and perceived behavioural control (Ajzen, 1991; Edwards-Jones, 2006). Therefore, the traditional models of farmer decision-making have been increasingly supplemented since the 1990s by simulation models, which integrate theoretical frameworks from psychology in order to fully understand farmer behaviour at the farm or individual scale. Such models include: system dynamic modelling (Guerrin, 2001; Keating et al., 2003; Darnhofer et al., 2011); discrete event modelling (Sokhansanj et al., 2006); and, most commonly, Agent-Based Modelling (ABM) (Railsback, 2001; Janssen, 2002; Strand et al., 2002; Edwards-Jones, 2006; Grimm et al., 2006; Janssen and Ostrom, 2006; Grimm et al., 2010).

2.3.1 Agent-Based Models (ABM)

Janssen and Ostrom (2006) defined ABM as the computational study of social agents, such as farmers, as evolving systems of autonomous interacting agents. The development of ABM can be traced back to early research by Neumann and Burks (1966), who provided a technical methodology for modelling multiple interacting agents during their work on cellular automata. This modelling approach was popularised during the 1970s after a study by Gardner (1970), who illustrated how following simple rules of local interaction could lead to the emergence of complex global patterns. However, cellular automaton models were limited in their ability to model the heterogeneity of agents beyond their specific location and history (Janssen and Ostrom, 2006). Therefore, early studies focussing on ABM, although theoretical, such as work on segregation (Schelling, 1971) and prisoner's dilemma strategies (Axelrod and Hamilton, 1981), showed how simple rules of interaction could explain more complex spatial patterns and levels of cooperation at the larger system scale (Janssen and Ostrom, 2006).

649 ABM are now widely used within the fields of ecology (DeAngelis et al., 1992;
650 Shugart et al., 1992; Van Winkle et al., 1993; Grimm, 1999; Gimblett, 2002; Huse
651 et al., 2002; DeAngelis and Mooij, 2005; Grimm and Railsback, 2013); social sciences
652 (Epstein and Axtell, 1996; Gilbert and Troitzsch, 2005; Epstein, 2006; Gilbert, 2007;
653 Billari and Prskawetz, 2012); economics (Tesfatsion, 2002; Fagiolo et al., 2007); and ge-
654 ography, particularly land use change, (dAquino et al., 2002; Parker et al., 2003; Evans
655 and Kelley, 2004; Brown et al., 2005; Matthews et al., 2007). The increase in popular-
656 ity of these models has partly been due to the advancements in computer science over
657 the last two decades, but also in their ability to consider aspects often ignored in an-
658 alytical, or neo-classical economic models, such as variability among individuals; local
659 interactions; complete life cycles; and individual behaviour adapting to the individual's
660 changing internal and external environment (Grimm et al., 2006).

661

662 2.3.2 Limitations of ABM

663 The main limitation of ABM lies within their development, largely criticised for being
664 far more complex in structure than typical analytical models (Grimm et al., 2006).
665 Furthermore, ABM are also often more difficult to analyse, understand and communi-
666 cate than traditional analytical models (Grimm, 1999). In regards to communication,
667 Grimm et al. (2006) argues that analytical models are easier to communicate as they
668 are formulated in the general language of mathematics; the description is often com-
669 plete, unambiguous and accessible to the reader, unlike ABM descriptions which are
670 often difficult to understand, incomplete, ambiguous, and therefore less accessible. As
671 a consequence of this, the results are often very difficult to reproduce, which largely un-
672 dermines ABM as a suitable modelling approach since science is based on reproducible
673 observations (Hales et al., 2003). As a potential solution to this limitation, Grimm and
674 Railsback (2005) suggested a standardised protocol for describing ABMs which was
675 successfully applied by several leading researchers in this field (Grimm et al., 2006) and
676 later revised after considerable use (Grimm et al., 2010).

677

678 The Overview, Design concepts, and Details (ODD) protocol was suggested as a
679 means of standardising the way researchers communicate their ABM to one another to
680 overcome one of the main criticisms of this modelling approach (Grimm et al., 2010).

681 The ODD protocol is subdivided into seven elements: purpose; entities, state vari-
682 ables, and scale; process overview and scheduling; design concepts (further divided into
683 several elements); initialisation; input data; and sub-models. Although the number of
684 studies using the ODD protocol has increased rapidly since it was first published, it has
685 also received criticism since some elements may be redundant for particular or simple
686 models. However, Grimm et al. (2010) argued that although this can be the case, it is
687 the price of having a hierarchical structure and the majority of times redundancy can
688 be avoided by keeping detail short and precise and any detail provided in the design
689 concept element can be left out of the sub-model element. Overall, ODD provides a
690 method of reviewing and categorising different types of ABM in a systematic approach
691 and has increasingly been used by researchers within the field of ABM (Grimm et al.,
692 2010).

693

694 **2.3.3 Use of ABM within the field of farmer decision-making**

695 In addition, although ABM has often been considered a promising quantitative method-
696 ology for social science research (Parker et al., 2003), it is only in the last few years
697 that researchers are combining ABM with empirical methods (Janssen and Ostrom,
698 2006). For example, Acosta et al. (2014) developed an ABM to understand the influ-
699 ence of global environmental change drivers and land manager decisions on the future
700 of the Montado, a multifunctional semi-wooded area used for grazing and cereal cul-
701 tivation in Portugal. The model description followed the standardised ODD protocol
702 suggested by Grimm et al. (2010). A farm typology, based on cluster analysis of em-
703 pirically derived farm attributes represented the different agents, along with particular
704 behavioural strategies which were driven by global economic and climatic parameters.
705 Overall, the ABM developed indicated that farmers would continue to abandon their
706 land if the future global economic environment, characterised by rapid industrialisa-
707 tion and urbanisation, should persist. However, agriculture remained the dominant
708 land use, indicating some resilience to change from local farmers.

709

710 Similarly, other studies have applied ABM to deal with complex human-environmental
711 systems such as Galán et al. (2009). In this study an ABM was used to integrate differ-
712 ent social sub-models: models of urban dynamics; water consumption; and technological

713 and opinion diffusion, which together were linked with a geographical information sys-
714 tem. The completed ABM represented a computational environment which was able to
715 simulate and compare different water demand scenarios for different agent types. In an-
716 other study, Schlüter and Pahl-Wostl (2007) developed an ABM which was designed to
717 compare the resilience of different institutional settings of water management to changes
718 in the variability and uncertainty of water availability for irrigation and environmental
719 water requirements. The results of this study indicated that under fluctuating inflows
720 and compliance with restrictions, a centralised management approach provided more
721 sustained levels of water availability when irrigation was the only user. However, as
722 demands from other sectors were introduced a decentralised management approach to
723 water allocation provided a more sustained level of water availability. Therefore, this
724 study highlights the ability of ABM to explore system level patterns of water use and
725 behaviour based on individual decision making at the farm scale.

726

727 **2.4 Further remarks and research gaps**

728 This chapter has presented a critical review of the main bodies of literature highlighting
729 three important research gaps pertaining to the methods which are to be used to ad-
730 dress the three research questions presented in Section 1.3. Firstly, despite the number
731 of studies addressing the issue of understanding farmers' behaviours having increased,
732 it appears no studies have attempted to understand farmers' preferred behavioural in-
733 tentions under different scenarios of water shortage and surplus with regards to the
734 proposed new water allocation systems in England. The TPB demonstrated its abil-
735 ity to provide a reliable method of successfully measuring intention and its predictors,
736 particularly with the addition of past behaviour. Therefore, Chapter 4 presents an
737 empirical investigation in a case study catchment in England, under the theoretical
738 framework of the TPB, aimed at addressing this research gap.

739

740 Secondly, traditional farm typologies fail to incorporate true farmer behaviour and
741 are therefore of limited value in supporting policy formulations (Guillem et al., 2012).
742 Furthermore, where studies have measured farmers' behaviours, relatively few have used
743 this data to develop farm typologies (Beedell and Rehman, 2000; Wauters et al., 2010;
744 Poppenborg and Koellner, 2013). In addition, it appears no studies have developed

745 a farm typology based on farmers' preferred behavioural intentions' under different
746 scenarios of water shortage and surplus with regards to the proposed water allocation
747 systems in England. Therefore, Chapter 5 presents a farm typology aimed at address-
748 ing this research gap.

749

750 Finally, although several studies have used ABM to simulate farmer decision-making
751 in regards to irrigation and land use (Bousquet et al., 1999; Barreteau et al., 2001;
752 Berger, 2001; Becu et al., 2003; Hare and Deadman, 2004; Janssen and Ostrom, 2006;
753 Matthews et al., 2007; Schreinemachers and Berger, 2011), no studies have explicitly
754 used ABM as a method of simulating farmers' behavioural intentions under different
755 scenarios of water shortage and surplus with regards to the proposed water allocation
756 systems in England. Furthermore, an ABM which incorporated a farm typology based
757 on farmers' preferred behavioural intentions could be used to: a) see what patterns
758 of abstraction behaviour emerge at the system level based on decisions made at the
759 farm level; b) see how different climate and proposed water allocation system scenarios
760 affect decisions made at the farm level; and c) see how the results of an ABM can
761 be used to inform policy-decision makers involved in the process of implementing the
762 proposed water allocation systems. Therefore, Chapter 6 presents the development and
763 simulation scenario results of an ABM aimed at addressing these issues.

764

765 Chapter 3

766 The Great Ouse catchment, 767 Anglian region, UK

768 The Great Ouse catchment represents approximately a third of the Anglian region in
769 eastern England, UK, and was selected as the study area due to two important fac-
770 tors including: the importance of available freshwater resources for agricultural spray
771 irrigation (see Section 1.1.4); and the fact that the Anglian region, although represent-
772 ing more than 50 % of total irrigated land (Knox et al., 2000), and 62 % of estimated
773 spray irrigation abstractions in England and Wales (DEFRA, 2016a), is one of the most
774 water-stressed regions in the UK (Knox et al., 2009) (see Figure 3.1).

775

776 3.1 Biophysical environment

777 The Great Ouse catchment is approximately 8,596 km² with elevation in the range of
778 -2 and 241 metres above mean sea level (m.a.s.l.). On average, the area received 655
779 mm of precipitation annually, and 51 mm monthly, between 1973 and 2012 (MetOffice,
780 2016). This was 499 mm less annually, and 42 mm less monthly, than the UK in general.
781 In addition, 2011 was the driest year and 2012 was the wettest with annual rainfalls
782 of 454 mm and 810 mm respectively. Furthermore, on average, the area has been 1
783 °C warmer than the UK in general with 1986 being the coldest year whilst 2011 was
784 the hottest year with mean temperatures of 9 °C and 11 °C respectively (see Figure 3.2).

785

786 The Great Ouse meanders from the uplands in the south west to the tidal reaches
787 near Earith where it continues to the mouth of the river (i.e. the Wash) at King's

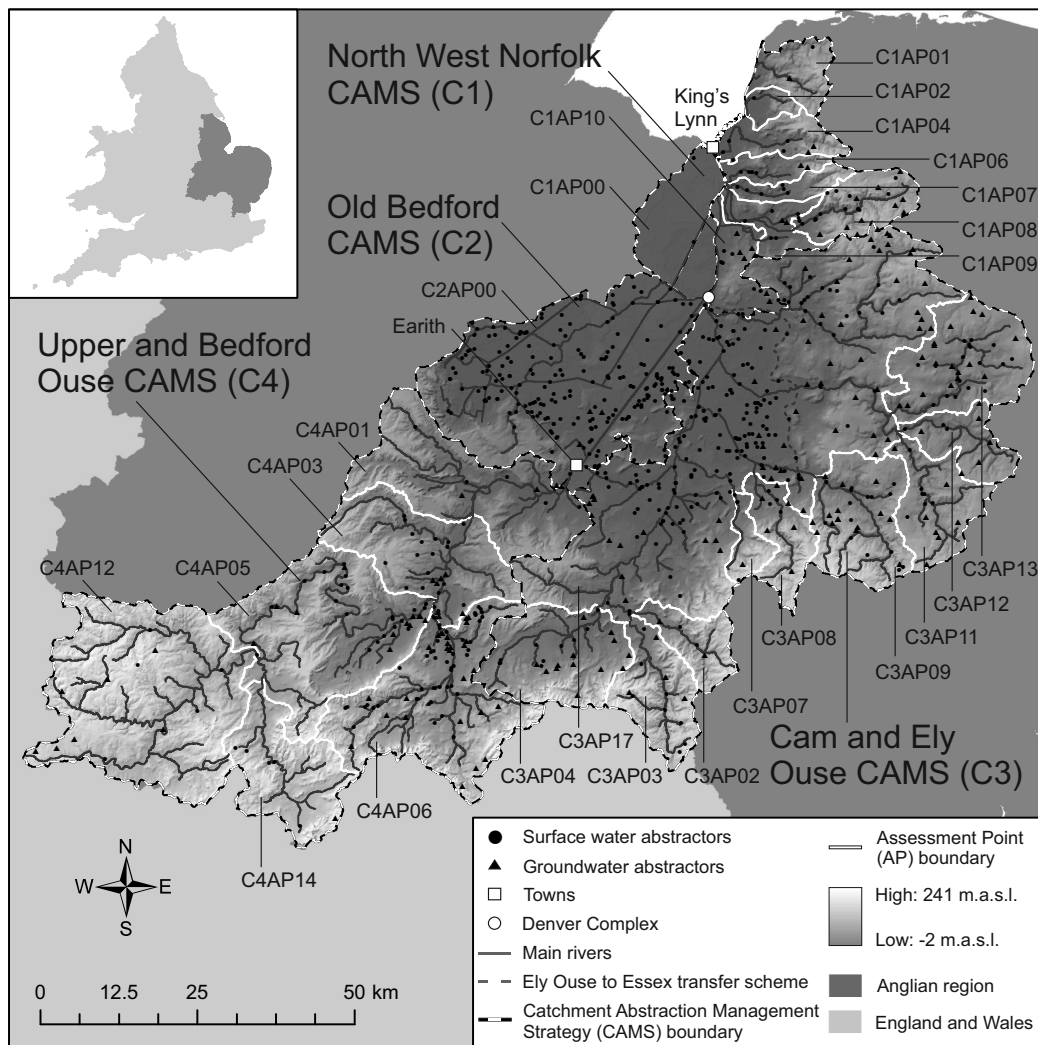


Figure 3.1: The Great Ouse catchment, Anglian region, UK

788 Lynn. Several major tributaries join the tidal river between Earith and King's Lynn,
 789 most notably the Ely Ouse in the south east. Much of the catchment in the north
 790 west is below sea level and consists of irrigation drainage channels (i.e. the Fens). The
 791 main aquifers in the study area include: the Great Oolite aquifer in the south west;
 792 the Woburn Sands aquifer in the south; the Chalk aquifer in the south and east of the
 793 catchment; and the Sandringham Sands aquifer in the north east (EA, 2013b,c,a,d).

794

795 The catchment is predominantly used for crop production, with 39 % classified as
 796 grade one or two agricultural land and a further 40 % classified as grade three (EA,
 797 2011). There are currently 836 spray irrigation farmers operating in the study area

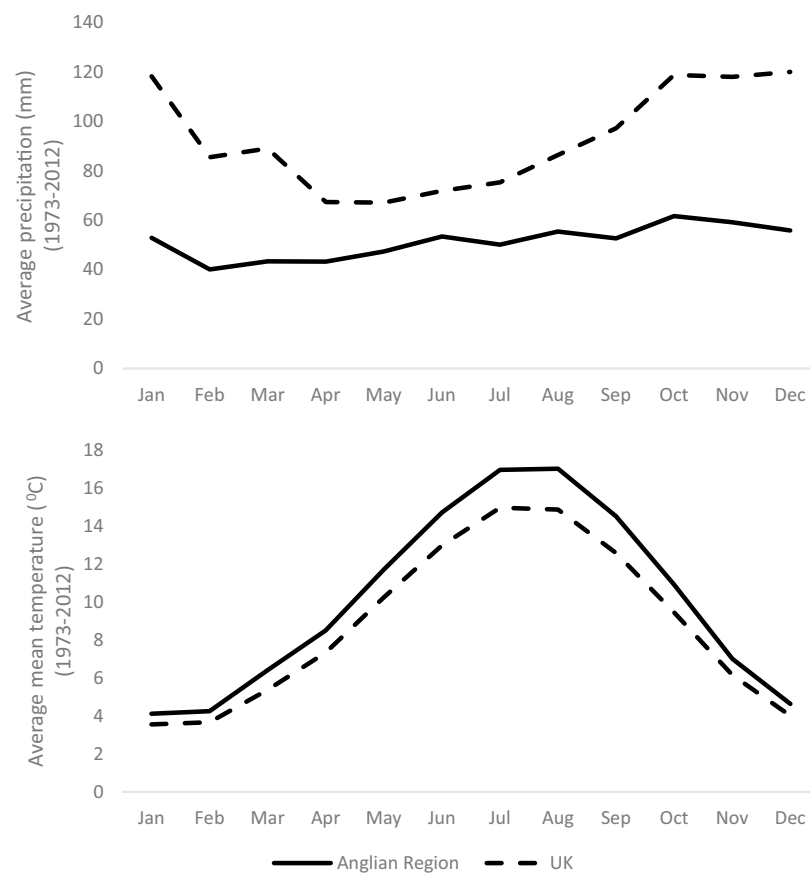


Figure 3.2: Difference in average monthly precipitation and mean temperature between the study area and the UK (1973 to 2012). (Top) Line graph indicating the difference in average monthly precipitation between the study area and the UK. (Bottom) Line graph indicating the difference in average monthly mean temperature between the study area and the UK

798 who, along with the remaining spray irrigation farmers in the larger Anglian region,
 799 rely on available freshwater for supplemental irrigation in order to supply high quality
 800 produce, including over 30 % of potatoes and 25 % of all fruit and vegetables, to UK
 801 supermarkets (UKIA, 2007). However, water available for spray irrigation in the study
 802 area is increasingly under pressure from climate change (see Section 1.1.4), demand
 803 from other sectors, most notably public water supply, and growing protection for the
 804 environment (Knox et al., 2009). Despite this increasing pressure, Weatherhead and
 805 Rivas Casado (2007) found that the total water applied each year for spray irrigation
 806 in England is increasing by 2.1 % per annum, as is the total area being irrigated (i.e.
 807 0.9 %).

3.2 Water management

Water availability in England is managed by the Environment Agency (EA) who, through their Catchment Abstraction Management Strategies (CAMS), determine water availability for further licensed abstractions. The study area is divided into four CAMS areas: North West Norfolk (EA, 2013b); Old Bedford (EA, 2013c); Cam and Ely Ouse (EA, 2013a); and Upper and Bedford Ouse (EA, 2013d), which are further divided into 50 Assessment Point (AP) areas. However, for reasons discussed in Section 3.3 only 26 AP areas were used in this study.

Three factors are used to determine water availability within an AP area: an Environmental Flow Indicator (EFI) which refers to a proportion of natural flow allocated to the environment; a Fully Licensed (FL) scenario which refers to the quantity of water which could be abstracted (i.e. a hypothetical scenario); and a Recent Actual (RA) scenario which refers to the quantity of water abstracted over the previous six years. A flow deficit occurs when water available after a RA scenario is below the EFI and therefore no water is available and the AP area is considered over abstracted. A risk of a flow deficit occurs when water available after a FL scenario is below the EFI and therefore restricted water is available and the AP area is considered over licensed. If there is no risk of a flow deficit then the AP area is considered as having water available (EA, 2010). However, the EA also use four flow levels, derived from flow duration curves which show the percentage of time that a particular flow level is equalled or exceeded, to account for natural annual flow variation including: Q30 (i.e. high flow); Q50; Q70; and Q95 (i.e. low flow). Therefore, despite the risk of a flow deficit water may still be available for further licensing but only during high flows.

In regards to the Great Ouse catchment, AP areas within the low lying fens are assessed differently by the EA with water available for further licensing in the Winter and not in the Summer, as river flows are largely controlled by a system of pumps and drains, most notably at the Denver Complex, managed by the Internal Drainage Board and the Middle Level Commissioners (EA, 2010). The Denver Complex controls inundation by the sea for much of the low lying fens in the Winter and provides water for spray irrigation in the Summer by using a large combination of sluices. In addition,

841 the Ely Ouse to Essex transfer scheme augments river flows and reservoir levels in south
 842 Essex by transferring excess water from the Ely Ouse during high flow conditions (i.e.
 843 inter-catchment transfer).

844

845 3.3 River flows

846 River flow data for each AP area was derived from the National River Flow Archive
 847 gauge station network (CEH, 2016). However, only 16 of the 50 AP areas used by
 848 the EA corresponded with available gauge station data. Furthermore, river flow data
 849 for these 16 AP areas consisted of missing data and were only available from 1973 to
 850 2012. Therefore, where gauge station records were incomplete average values for the
 851 AP area were used. For this study, any upstream AP area which did not correspond
 852 with available river flow data was amalgamated with the downstream AP area which
 853 did have river flow data available. Furthermore, due to the tidal nature of the river,
 854 10 AP areas were essentially all downstream catchments and none had river flow data
 855 available. Therefore, for these 10 AP areas the watershed area ratio method was used
 856 to derive river flows (Gianfagna et al., 2015) using either an upstream AP area or a
 857 similar sized neighbouring AP area:

$$858 \quad Q_{ungauged} = Q_{gauged} \left(\frac{A_{ungauged}}{A_{gauged}} \right) \quad (3.1)$$

859 where Q equals river flow (i.e. discharge) and A equals AP area (i.e. km²). Although
 860 this method provides a reasonable estimation for river flows in ungauged AP areas
 861 for this study, it does assume that river flow is proportional to area which may not
 862 always be accurate. In addition, for the purposes of this study, the two AP areas which
 863 represent the low lying fens (i.e. C1AP00 and C2AP00) derived their river flows using
 864 the watershed area ration method, from C3AP17 in the Cam and Ely Ouse CAMS
 865 area. River flow in this AP area largely influences the availability of freshwater at the
 866 Denver Complex, and therefore also for the low lying fens. Figures 3.3, 3.4, and 3.5
 867 illustrate, respectively, a schematic of the 26 AP area network used in this study, flow
 868 duration curves for each AP area, and average monthly river flows for each AP area.
 869 Furthermore, Table 3.1 summarises average river flows for each AP area at the flow
 870 exceedence levels used by the EA to determine water availability.

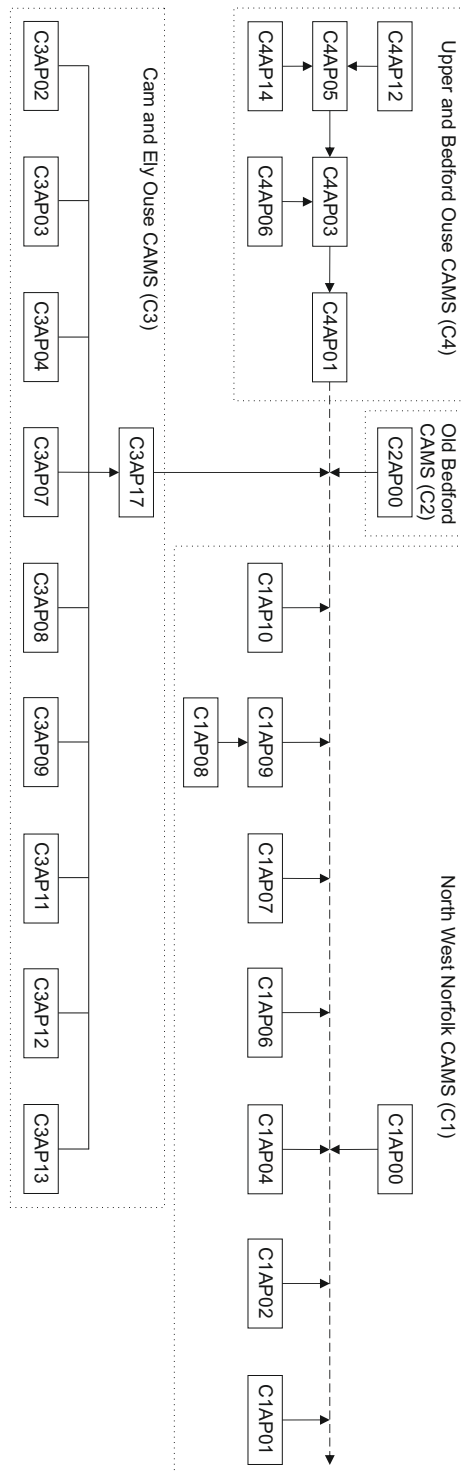


Figure 3.3: Schematic diagram illustrating river flow between Assessment Point (AP) areas within each of the four Catchment Abstraction Management Strategy (CAMS) areas (the dash line relates to the tidal river whilst dotted lines relate to CAMS boundaries)

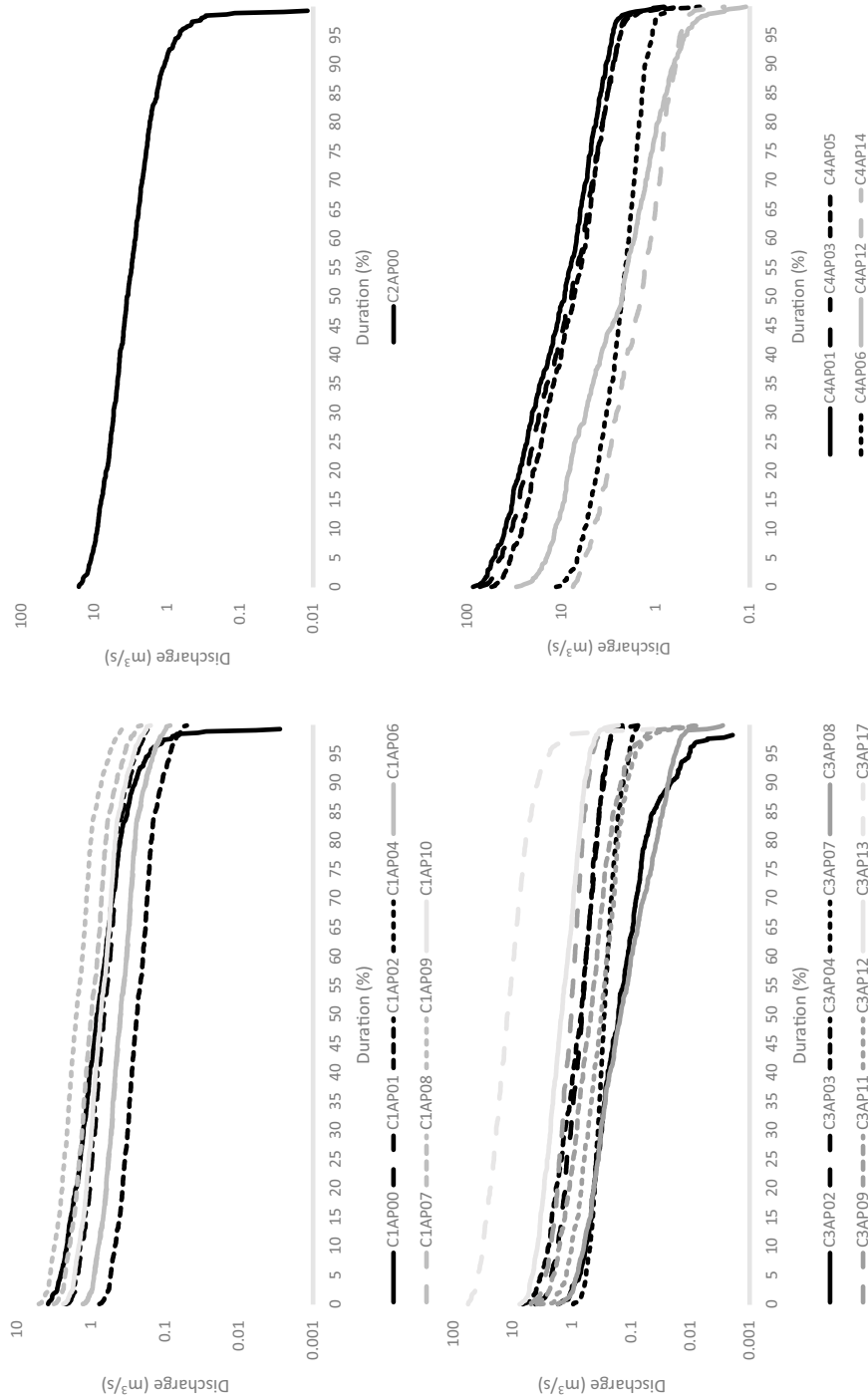


Figure 3.4: Flow duration curves of each AP area within the study area (1973 to 2012). (Top left) Flow duration curves of AP areas within the North West Norfolk CAMS area. (Top right) Flow duration curve of the AP area within the Old Bedford CAMS area. (Bottom left) Flow duration curves of AP areas within the Cam and Ely Ouse CAMS area. (Bottom right) Flow duration curves of AP areas within the Upper and Bedford Ouse CAMS area

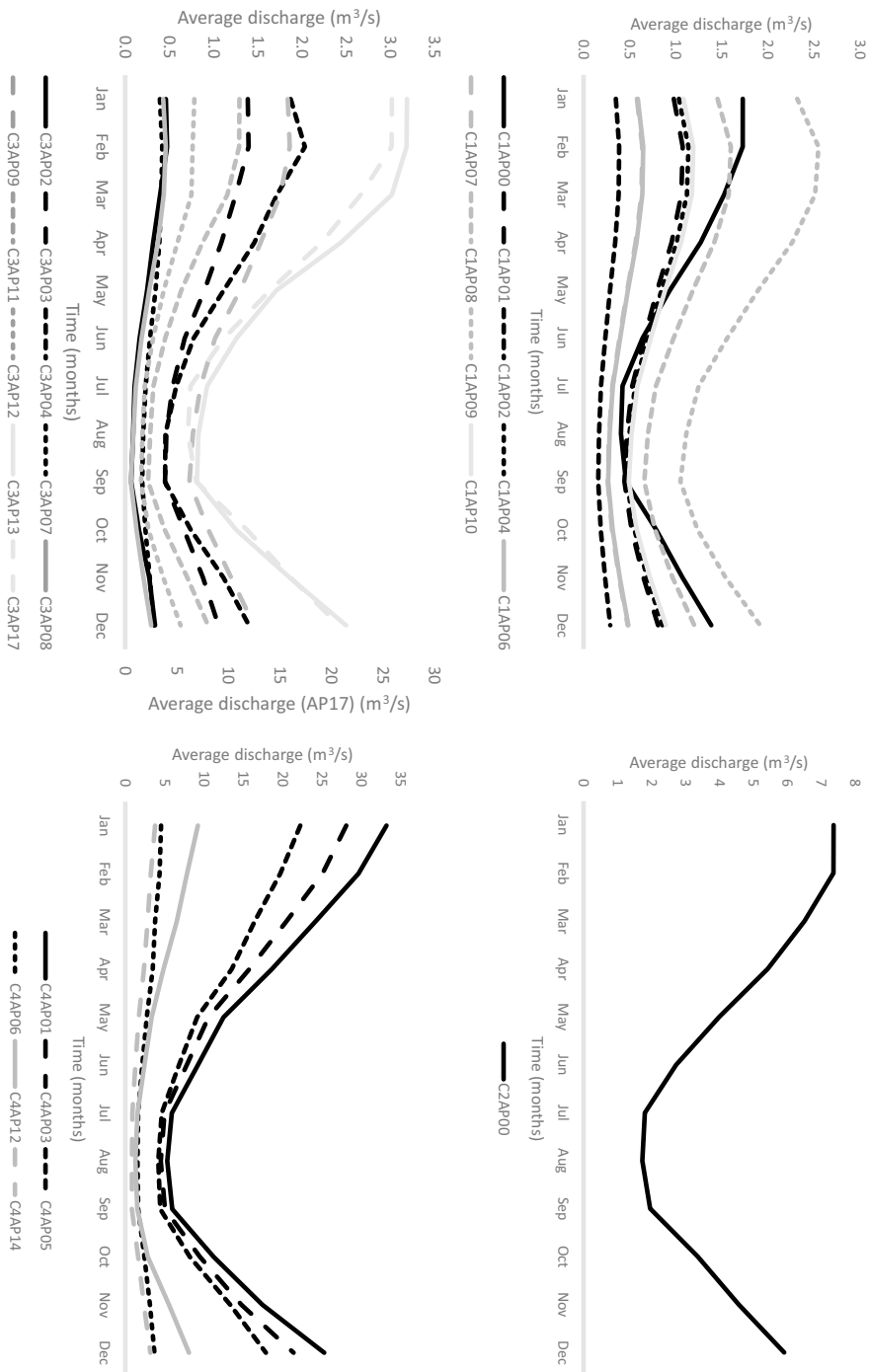


Figure 3.5: Average monthly river flows of each AP area within the study area (1973 to 2012). (Top left) Average monthly river flows of AP areas within the North West Norfolk CAMS area. (Top right) Average monthly river flow of the AP area within the Old Bedford CAMS area. (Bottom left) Average monthly river flows of AP areas within the Cam and Ely Ouse CAMS area. (Bottom right) Average monthly river flows of AP areas within the Upper and Bedford Ouse CAMS area

Table 3.1: Average river flows for each Assessment Point (AP) area (1973 to 2012)

Assessment Point (AP)	Gauge station	Upstream area (km ²)	Q30 flow (m ³ /s)	Q50 flow (m ³ /s)	Q70 flow (m ³ /s)	Q95 flow (m ³ /s)
North West Norfolk CAMS (C1)						
- C1AP01	-	103	0.89	0.66	0.49	0.24
- C1AP02	-	37	0.32	0.24	0.18	0.09
- C1AP04	-	109	0.94	0.70	0.52	0.25
- C1AP06	-	62	0.54	0.40	0.29	0.14
- C1AP07	-	61	0.53	0.39	0.29	0.14
- C1AP08	33007	153	1.33	0.99	0.72	0.35
- C1AP09	-	91	2.11	1.57	1.15	0.56
- C1AP10	-	114	0.99	0.74	0.54	0.26
- C1AP00	-	229	1.27	0.83	0.52	0.16
Old Bedford CAMS (C2)						
- C2AP00	-	977	5.43	3.56	2.21	0.68
Cam and Ely Ouse CAMS (C3)						
- C3AP02	33053	114	0.32	0.16	0.08	0.01
- C3AP03	33024	198	1.05	0.69	0.48	0.25
- C3AP04	33021	303	1.31	0.70	0.47	0.26
- C3AP07	33050	61	0.36	0.29	0.22	0.11
- C3AP08	33023	102	0.33	0.15	0.06	0.02
- C3AP09	33014	272	1.47	1.04	0.78	0.44
- C3AP11	33013	206	0.82	0.48	0.30	0.10
- C3AP12	33011	129	0.54	0.32	0.20	0.09
- C3AP13	33019	316	2.31	1.50	0.95	0.50
- C3AP17	33035	3,430	19.06	12.48	7.76	2.38
Upper and Bedford Ouse CAMS (C4)						
- C4AP01	-	3,038	20.54	9.30	5.43	2.90
- C4AP03	33026	2,570	17.37	7.87	4.59	2.45
- C4AP05	33039	1,660	14.59	7.36	4.47	2.37
- C4AP06	33022	541	3.33	2.27	1.69	1.09
- C4AP12	33037	800	5.53	2.27	1.28	0.46
- C4AP14	33015	277	2.59	1.42	0.90	0.55

If gauge station is not present then river flow was calculated using equation 3.1
Source (CEH, 2016)

871 Generally, the flow duration curves are relatively flat between high and low flow
872 conditions for all AP areas indicating the presence of surface storage (i.e. reservoirs) or
873 groundwater storage (i.e. aquifers) which tend to equalize flows (Searcy, 1959). How-
874 ever, in some AP areas, duration curves are steep below low flow conditions, indicating
875 almost zero flows. Furthermore, average monthly flows indicate that lower flows gen-
876 erally occur during the Summer, when spray irrigation demand is highest, and higher
877 flows during the Winter, when spray irrigation demand is lowest.

878 3.4 Summary

879 Overall, Figure 3.5 illustrates the current water availability status of the 24 AP areas
880 defined in this study, excluding the two AP areas which are assessed differently due
881 to their location within the low lying fens (i.e. C1AP00 and C2AP00). At high flows
882 (i.e. Q30) 83 % of the catchment has water available for further licensing with the
883 remainder having restricted water available. However, at moderate flows (i.e. Q50)
884 only 13 % of the catchment has water available for further licensing with 54 % having
885 restricted water available, and 33 % having no water available. At below moderate
886 flows (i.e. Q70) only 4 % of the catchment has water available for further licensing
887 with 13 % having restricted water available, and 83 % having no water available. Fi-
888 nally, at low flows (i.e. Q95) only 4 % of the catchment has water available for further
889 licensing with the remainder having no water available. Therefore, under the current
890 water allocation system there is little opportunity for issuing new abstraction licences
891 within the Great Ouse catchment except at high flows. Furthermore, climate change
892 and demand increases from other sectors is expected to exacerbate the vulnerability of
893 freshwater availability for spray irrigation in this catchment.

894

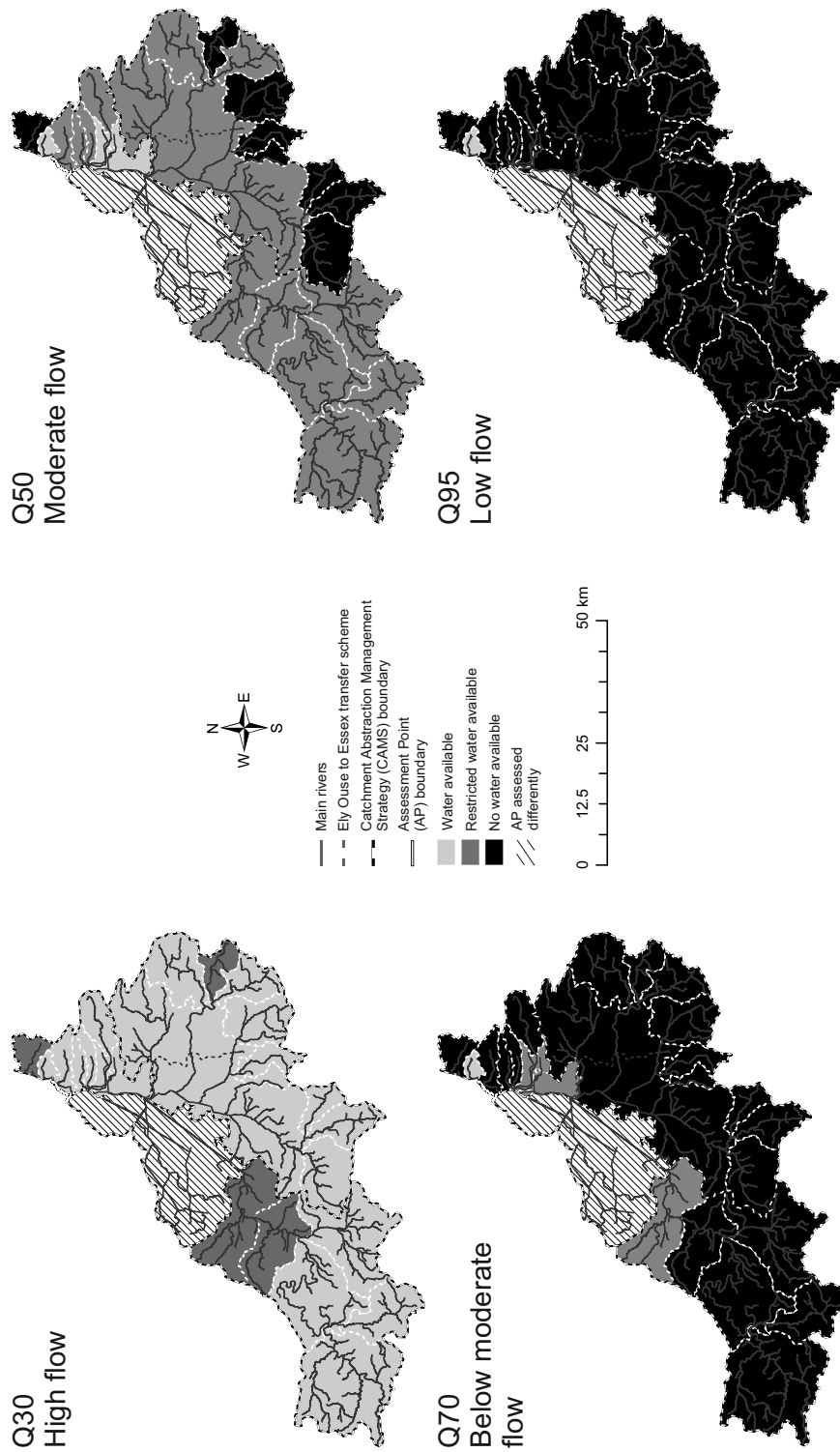


Figure 3.6: Water availability for further licensed abstractions within the study area from high to low flows (2013). (Top left) Water availability at Q30 (i.e. high flow). (Top right) Water availability at Q50 (i.e. moderate flow). (Bottom left) Water availability at Q70 (i.e. below moderate flow). (Bottom right) Water availability at Q95 (i.e. low flow)

895 Chapter 4

896 Farmers' behaviours

897 This chapter presents empirical data which identifies farmers' preferred behavioural
898 intentions under different scenarios of water shortage and surplus with regards to the
899 proposed water allocation systems in England. In this study, the term farmers' refer
900 to spray irrigation licence users' only. After an initial pilot study, a questionnaire was
901 sent to 826 farmers' in the study area, of which 11% responded. The questionnaire, and
902 subsequent analysis, was based on the theoretical framework of the Theory of Planned
903 Behaviour (TPB) (Ajzen, 1985). The results of this study are intended to inform policy
904 decision-makers involved in the design and implementation of the proposed water allo-
905 cation systems by: 1) understanding how the proposed water allocation systems might
906 be received by farmers' on the ground; and 2) identifying which underlying predictors
907 of intention most influence farmers' decision-making. Therefore, this chapter addresses
908 the first of the three research questions presented in the introductory chapter to this
909 thesis:

910

911 **Research question 1:** *What are farmers' preferred behavioural intentions under*
912 *different scenarios of water shortage and surplus with regards to the proposed water*
913 *allocation systems in England? And which underlying predictors of intention most in-*
914 *fluence their decision-making?*

915

916 4.1 Introduction

917 As discussed in Chapter 1 of this thesis, major water allocation reforms are currently
918 being proposed in England to address the climate and demand change pressures, which

919 the current system is failing to address (DEFRA, 2011). Furthermore, evidence has
920 been presented which highlights how farmer behaviour very much influences how and
921 to what success policy proposals are realised on the ground (Moon and Cocklin, 2011;
922 Home et al., 2014; Feola et al., 2015) (see Chapter 2). Policy decision-makers therefore
923 increasingly require more sophisticated methods of anticipating farmers' motivation to
924 comply with new policy approaches (Austin et al., 1998; Burton, 2004; Emtage et al.,
925 2006; Guillem et al., 2012). The TPB (Ajzen, 1985) was critically reviewed and has
926 been identified as a relatively reliable method of successfully measuring intention and
927 its predictors (see Chapter 2). Furthermore, evidence has been provided which has
928 shown that the addition of past behaviour as a predictor variable can increase the pre-
929 dictive power of the TPB model (Albarracin et al., 2001; Sandberg and Conner, 2008;
930 Rise et al., 2010). In addition, the TPB also provides a means by which the relative
931 importance (weights) of each of the underlying constructs (predictors) of intention can
932 be estimated. Therefore, this chapter presents empirical data, from the selected study
933 area (see Chapter 3), which identifies farmers' preferred behavioural intentions under
934 different scenarios of water shortage and surplus with regards to the proposed water
935 allocation systems in England.

936

937 **4.2 Materials and Methods**

938 The methodological procedure used in this study can be categorised into four stages:
939 1) scenario and behaviour selection; 2) questionnaire design; 3) survey procedure; and
940 4) statistical analysis.

941

942 **4.2.1 Scenario and behaviour selection**

943 In total there were four scenarios and 18 behaviours selected (see Table 4.1). These
944 included seven strategic (long-term)/water shortage behaviours, four strategic (long-
945 term)/water surplus behaviours, four in-season (short-term)/water shortage behaviours,
946 and three in-season (short-term)/water surplus behaviours. The strategic and in-season
947 scenarios related to farm level water management decisions made once every several
948 growing seasons, and regularly during each season, respectively. The water shortage
949 and surplus scenarios related to decisions when a farmer is unable to meet crop wa-

Table 4.1: Scenarios and behaviours (options) selected

Low Water Usage Options (LWUO)	High Water Usage Options (HWUO)
Strategic (long-term)/water shortage scenario	
- Grow the same crops but over a smaller area	- Increase storage capacity
- Grow less water intensive crops	- Buy more water for the duration of the growing season
- Increase application efficiency	- Apply for a larger abstraction licence
- Change nothing	
Strategic (long-term)/water surplus scenario	
- Sell surplus water for the duration of the growing season	- Grow the same crops but over a larger area
- Change nothing	- Grow more water intensive crops
In-season (short-term)/water shortage scenario	
- Only use your maximum abstraction licence to spread water evenly between all crops	- Buy more water to meet crop water requirements
- Only use your maximum abstraction licence to irrigate your most valuable crops	
- Restrict application (i.e. deficit irrigation)	
In-season (short-term)/water surplus scenario	
- Sell surplus water to maximise profits	- Abstract surplus water for storage
- Just use your abstraction licence to meet crop water requirements and leave the remainder of your licence unused	

950 ter requirements, or has a surplus of water after they meet crop water requirements,
951 respectively. The scenarios and behaviours were selected after consultation with sev-
952 eral farmers during multiple UKIA meetings held in and around the selected study
953 area between October 2012 and December 2013. Furthermore, an irrigation specialist
954 with experience in the selected study area reviewed the scenarios and behaviours (K.
955 Weatherhead, personal communication, November 2013) as did the head of water ab-
956 straction reform at DEFRA (H. Leveson-Gower, personal communication, December
957 2013). This selection approach was favoured over holding focus groups as many farmers
958 commented on their time restraints, and the UKIA meetings provided an opportunity
959 to meet several farmers simultaneously.

961 Under a strategic (long-term)/water shortage scenario farmers require additional
962 water to meet crop water requirements and have the following options available: grow
963 the same crops but over a smaller area; grow less water intensive crops; increase storage
964 capacity; increase application efficiency; buy more water for the duration of the growing
965 season; apply for a larger abstraction licence; or change nothing. Growing the same
966 crops but over a smaller area; growing less water intensive crops; increasing applica-
967 tion efficiency; or changing nothing were considered, for the purposes of this study, as
968 Low Water Usage Options (LWUO), as any one of these behaviours ultimately means
969 abstracting the same or less volume of water. Furthermore, farmers who favour these
970 options are expected to experience a reduction in crop quality or yield. In contrast,
971 increasing storage capacity; buying more water for the duration of the growing sea-
972 son; or applying for a larger abstraction licence were considered as High Water Usage
973 Options (HWUO) as any one of these behaviours ultimately means abstracting more
974 water. Farmers who favour these options are expected to reduce the risk of a reduction
975 in crop quality or yield.

976

977 Under a strategic (long-term)/water surplus scenario farmers have a surplus of wa-
978 ter after meeting crop water requirements and have the following options available:
979 grow the same crops but over a larger area; grow more water intensive crops; sell sur-
980 plus water for the duration of the growing season; or change nothing. The latter two
981 behaviours were considered as LWUO whilst the first two were considered as HWUO
982 due to the decrease and increase in water used, compared to an average year, respec-
983 tively. Farmers who favour selling surplus water, a LWUO, are expected to increase
984 their income similar to those farmers who favour the HWUO.

985

986 Under an in-season (short-term)/water shortage scenario farmers have fewer options
987 available to meet crop water requirements including: only using their maximum ab-
988 straction licence to spread water evenly between all crops; only using their maximum
989 abstraction licence to irrigate their most valuable crops; restricting application (i.e.
990 deficit irrigation); or buying more water to meet crop water requirements. The first
991 three behaviours were considered as LWUO whilst the latter was considered a HWUO
992 for the same reasons presented under a strategic (long-term)/water shortage scenario.

993

994 Lastly, under an in-season (short-term)/water surplus scenario farmers also have
995 fewer options available including: selling surplus water to maximise profits; just using
996 their abstraction licence to meet crop water requirements and leaving the remainder of
997 their licence unused; or abstracting surplus water for storage. The first two behaviours
998 were considered as LWUO whilst the latter was considered a HWUO for the same rea-
999 sons presented under a strategic (long-term)/water surplus scenario.

1000

1001 4.2.2 Questionnaire design

1002 The questionnaire (see Appendix A) was designed into four sections: farm attributes
1003 (section A); farmers' behavioural intentions under strategic (long-term) and in-season
1004 (short-term) scenarios (sections B and C respectively); and social demographics (sec-
1005 tion D). In total, there were 108 questions: five associated with section A; 98 associated
1006 with sections B and C; and five associated with section D. The design of the question-
1007 naire, as with the scenarios and behaviours selected, was developed after consultation
1008 with several farmers during multiple UKIA meetings held in and around the selected
1009 study area between October 2012 and December 2013. Furthermore, feedback for the
1010 final phrasing of the questions was provided by an irrigation specialist familiar with the
1011 study area (K. Weatherhead, personal communication, November 2013) and the head
1012 of water abstraction reform at DEFRA (H. Leveson-Gower, personal communication,
1013 December 2013).

1014

1015 Section A of the questionnaire attempted to identify farmers' three main irrigated
1016 crops (as several farmers reported irrigating more than one crop over equal areas); the
1017 irrigated area of each of these crops; the predominant soil type each crop was grown
1018 on (as this influenced irrigation water demand); the irrigation technique used for each
1019 crop (to assess the efficiency of each farm); and the overall water storage capacity of
1020 the farm (to assess farm storage reserves when licensed abstractions were unavailable).
1021 Example categories of crops included in the first question, and subsequently used in
1022 later analysis, were based on those used by Weatherhead and Rivas Casado (2007) and
1023 included: early potatoes; main crop potatoes; sugar beet; orchard fruit; small fruit;
1024 vegetables; grass; cereals; and other.

1025

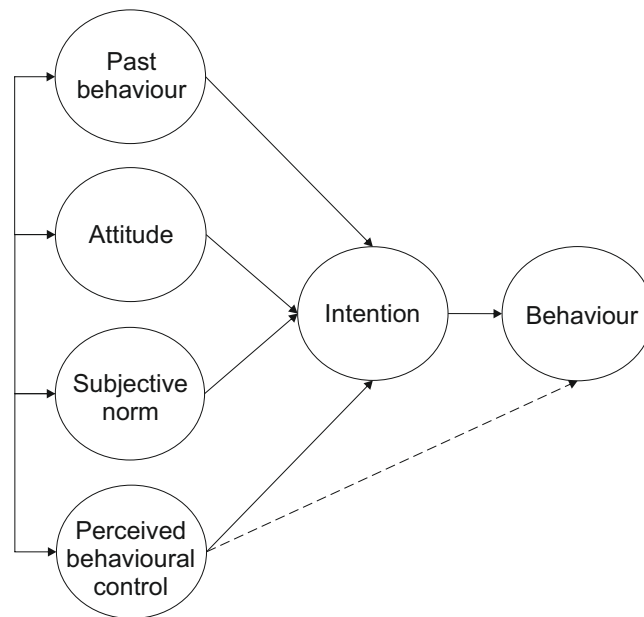


Figure 4.1: The Theory of Planned Behaviour including past behaviour (adapted from Ajzen (1991))

1026 Sections B and C, which were both divided into further water shortage and surplus
 1027 scenarios, related to the scenarios and behaviours previously discussed. In addition,
 1028 as the questionnaire was structured based on the theoretical framework of the TPB
 1029 (Ajzen, 1985) (see Chapter 2), each behaviour consisted of five questions related to
 1030 overall intention as well as the underlying constructs (or predictors) of intention in-
 1031 cluding: attitude; subjective norm; perceived behavioural control; and past behaviour
 1032 (see Figure 4.1). A Likert psychometric scale of one to seven (one being negative and
 1033 seven being positive) was used to record farmers' responses. In regards to subjective
 1034 norms, farmers were also asked to rank which social group most influenced their de-
 1035 cisions. These included: a group interested in water saving and efficiency; a group
 1036 interested in crop type and quality; and a group consisting of family, friends, and
 1037 neighbours. Lastly, in regards to past behaviour, an additional question was included
 1038 asking farmers whether they had actually experienced the scenario.

1039

1040 Section D of the questionnaire attempted to identify farmers' gender; age; level of
 1041 education (measured by National Qualification Framework level for England, Wales
 1042 and Northern Ireland); total gross annual farm business income; and whether they had
 1043 used their abstraction licence within the last ten years.

1044 4.2.3 Survey procedure

1045 The farmers selected for the survey were chosen due to their location within the se-
1046 lected study area (see Chapter 3). The EA provided data regarding the addresses and
1047 abstraction point locations for the total population of farmers in the study area (a total
1048 of 836) from the National Abstraction Licence Database. A pilot survey was conducted
1049 by post, on the 14th January 2014, with a sample of 10 randomly selected farmers. The
1050 pilot survey included: the questionnaire; a cover letter; and a consent form. However,
1051 this randomly selected sample had to be selected from the total population within the
1052 study area, as the EA did not permit additional access to farmers' addresses outside
1053 the study area, thus only a small sample were chosen as these farmers were then ex-
1054 cluded from the final survey. Farmers selected for the pilot survey had two weeks to
1055 complete and return the questionnaire before a reminder was sent. The questionnaire
1056 and reminder were posted with first class return envelopes to avoid costs to farmers;
1057 however no incentives were offered for completing a questionnaire.

1058

1059 In the initial pilot survey a second perceived behavioural control question, asking
1060 farmers to report on the difficulty in performing a behaviour, was included in order to
1061 examine the internal control aspect of self-efficacy, such as ability or motivation, rather
1062 than simply external control, such as resources (Manstead and Eekelen, 1998):

1063

1064 *How difficult or easy would it be to... (e.g. increase application efficiency?)*

1065

1066 was used in addition to:

1067

1068 *How much do you disagree or agree with the following statements? (e.g. Increasing*
1069 *application efficiency or not is entirely up to you)*

1070

1071 The pilot survey response rate was 30 % which was towards the higher end compared
1072 to that reported in the survey literature (i.e. 20 % to 30 %) (Yammarino et al., 1991;
1073 Pennings et al., 2002). The pilot survey provided valuable feedback, most noticeably
1074 highlighting the redundancy of the additional perceived behavioural control question
1075 as none of the respondents noted any difference between these two questions with many
1076 commenting negatively on the length and repetitiveness of the questionnaire, and in

1077 particular those two questions. Therefore, in response to the pilot survey, and to reduce
1078 the risk of a low response rate in the final survey, the additional question was removed
1079 based on the knowledge that this would not cause problems to the way perceived be-
1080 havioural control was used as a measure of intention. The final questionnaire was then
1081 distributed by post on the 12th February 2014 to the remaining population of farmers
1082 in the study area, a total of 826, offering them one month to respond. No incentives
1083 were offered or reminders sent due to available resources.

1084

1085 **4.2.4 Statistical analysis**

1086 All statistical analyses were performed using SPSS v22. In addition, the EA provided
1087 additional licence characteristic data and past abstraction data for those who responded
1088 to the questionnaire, rather than the total population, due to EA data restrictions. Li-
1089 cence characteristic data included: licence type (i.e. direct; storage; or direct and
1090 storage); licence source (i.e. surface or groundwater); whether licences were time lim-
1091 ited or not; and annual and daily licensed volumes. Past abstraction data consisted
1092 of monthly abstractions from the start of 2008 to the end of 2012. Furthermore, the
1093 statistical analysis for this study was conducted and is presented in two parts. The first
1094 part focussed on farm characteristics including: farm attributes; social demographics;
1095 licence characteristics; and past abstractions (see Appendix B). The second part fo-
1096 cussed on farmers' behavioural intentions (see Appendix C).

1097

1098 **Farm characteristics**

1099 In regards to farm characteristics, relative frequency tables and histograms were used
1100 to highlight the results of the respondents. Furthermore, where comparative data were
1101 available, the results were compared to the characteristics of the total surveyed pop-
1102 ulation and farmers in England and Wales in general. In addition, two-tailed tests
1103 of association were conducted to explore whether associations existed between: stor-
1104 age capacities and irrigated areas; annual and daily licence limits; annual abstractions
1105 and precipitation; and monthly abstractions and precipitation. Although for the first
1106 two tests it could be expected that positive associations were more likely, a negative
1107 association could not be ruled out, particularly with regards to the role of seasonal

1108 restrictions on licences resulting in large daily licence limits for a relatively small an-
1109 nual licence. Conversely for the latter two tests it could be expected that negative
1110 associations were more likely, as one would expect abstractions to increase as precip-
1111 itation decreases. However, additional water available may have meant that farmer's
1112 abstract more for storage, whilst less water available may coincide with HoF restric-
1113 tions. Therefore, the null hypotheses for these tests were that there was no significant
1114 association between variables. The appropriate parametric (Pearson's correlation), or
1115 non-parametric (Spearman's rank correlation), test of association was used depending
1116 on whether the data were normally distributed or not.

1117

1118 The Shapiro-Wilk test was used to test whether quantitative data was normally
1119 distributed (Shapiro and Wilk, 1972). The null hypothesis for the Shapiro-Wilk test
1120 was that the data are normally distributed. Therefore, if the *p-value* was less than the
1121 chosen alpha level, which for this test was .05, then the null hypothesis was rejected
1122 suggesting that the data were not from a normal distribution. However, as this test
1123 is biased by sample size, normal probability plots were presented for additional veri-
1124 fication. Normal probability plots are graphical techniques used to assess whether or
1125 not a data set is approximately normally distributed. The data are plotted against a
1126 theoretical normal distribution in a way that the data points should form a straight
1127 line. Departure from the line indicates the data are not normally distributed (Cham-
1128 bers et al., 1983).

1129

1130 Therefore, with regards to the tests of association, if the *p-value* was less than the
1131 chosen alpha level, which for these tests were .05, then the null hypothesis was rejected
1132 suggesting that the variables were associated. Furthermore, a Pearson's correlation
1133 coefficient (r), or a Spearman's rank correlation coefficient (r_s), of plus or minus 0.7 or
1134 greater was considered a strong correlation, plus or minus 0.5 a moderate correlation,
1135 and plus or minus 0.3 a weak correlation.

1136

1137 **Behavioural intentions**

1138 In regards to farmers' behavioural intentions, relative frequency tables were used to
1139 summarise the results of the respondents, whilst median and IQR values highlighted

1140 the variability of responses for each Likert item. As discussed, each behaviour com-
1141 prised of five Likert items, relating to each of the TPB constructs, and each Likert item
1142 comprised of a seven point Likert item scale ranging from negative to positive. In order
1143 to identify respondents preferred behavioural intentions for each scenario, the scores
1144 on each Likert item scale were rescaled from minus three to three and multiplied by
1145 the number of responses. The total sum of these rescaled Likert items were summed
1146 for each behaviour. This provided an overall single value which could be used to rank
1147 behaviours from most preferred to least preferred, where a positive or negative value
1148 corresponded to a positive or negative overall intention to perform a behaviour.

1149

1150 In addition, sign tests were conducted to test whether behaviours were statistically
1151 preferred to the behaviour ranked immediately below. The null hypothesis for the sign
1152 test was that the median of the differences between the two behaviours was zero. In
1153 addition, the sign test does not make any assumptions about distribution, and was
1154 therefore considered appropriate for use with ordinal data. If the *p-value* was less than
1155 the chosen alpha level, which for this test was .05, then the null hypothesis was re-
1156 jected suggesting that there was a difference in medians between the two behaviours.
1157 The frequency of positive and negative responses, between the two behaviours, was
1158 used to determine whether the difference in medians increased or decreased (Dixon
1159 and Mood, 1946). However, rather than perform the sign test on each construct, for
1160 each behaviour, only intention was used for simplicity and ease of interpretation and
1161 therefore despite the ranking, which takes into consideration all of the constructs, an
1162 increase in intention might occur for a lower ranked behaviour.

1163

1164 Studies employing the TPB framework typically use multiple linear regression to
1165 examine the relationships between the response and predictor variables (i.e. intention
1166 and underlying constructs). Despite this method requiring data measured on a con-
1167 tinuous scale (Hankins et al., 2000), these studies usually measure underlying salient
1168 beliefs for each construct using multiple Likert items. The Likert items for each con-
1169 struct are then assessed for internal consistency and averaged to form an overall Likert
1170 scale for each construct. These Likert scales are often treated as continuous to meet
1171 the assumptions required for multiple linear regression (Armitage and Conner, 2001).
1172 However, as this study forewent measuring underlying salient beliefs, in order to iden-

1173 tify respondents' preferred behaviours from a wider range of behaviours, binary logistic
1174 regression was used. This regression technique uses the maximum likelihood method,
1175 rather than ordinary least squares, to estimate the probability of a binary response
1176 based on one or more categorical predictors (Peng et al., 2002).

1177

1178 This method is to be favoured over ordinal logistic regression, despite the loss of in-
1179 formation incurred by dichotomising ordinal scales based on median values (Roozenbeek
1180 et al., 2011), due to ease of interpretation and data limitations caused by small sample
1181 sizes (i.e. with the data collected in this study complete or quasi-complete separation
1182 occurred for each preferred behavioural intention when using ordinal logistic regression
1183 thus leading to erroneous results) (Webb et al., 2004). Furthermore, as this study was
1184 interested in the respondents' preferred behaviours under each of the four scenarios,
1185 binary logistic regressions were only conducted for the highest ranked behaviours. Peng
1186 et al. (2002) suggested that in order to evaluate the effectiveness of a binary logistic
1187 regression model, it is important to examine: the overall model; statistical tests of in-
1188 dividual predictors; goodness-of-fit statistics; and validations of predicted probabilities.

1189

1190 However, prior to conducting the binary logistic regression, tests for multi-collinearity
1191 of the dichotomous predictor variables were conducted based on inspection of the cor-
1192 relation matrix and variance inflation factors (VIF). Although no formal cutoff values
1193 exist, correlation coefficients greater than 0.8 and VIF values greater than 2.5, when
1194 analysing logistic regression models, were considered signs of multi-collinearity follow-
1195 ing Midi et al. (2010). Moreover, plots of standardised Pearson residuals are presented
1196 to highlight any possible outliers exceeding the absolute value of 2.58 which was the
1197 cutoff value proposed by Jöreskog and Sörbom (1989). If these occurred, then binary
1198 logistic regression was conducted with and without the outliers in order to compare
1199 results. Few other assumptions are required for this regression technique other than
1200 the response variable (i.e. intention) has to be a dichotomous variable, the categories
1201 must be mutually exclusive, and as maximum likelihood coefficients are large sample
1202 estimates, a minimum of 50 cases per predictor variable was recommended (Peng et al.,
1203 2002). Thus, despite dichotomising the response and predictor variables based on their
1204 median values, the inclusion of past behaviour as a predictor variable still resulted in
1205 complete or quasi-complete separation of each of the preferred behavioural intentions

1206 due to the low response rates obtained for this particular construct, and was therefore
1207 not included in the following analyses.

1208

1209 The null hypothesis for the binary logistic regressions was that the regression coeffi-
1210 cients (β), or predictor variables, equalled zero. That is, the predictor variables did not
1211 improve the predictive ability of the null model (i.e. the constant only model). Overall
1212 model evaluation was assessed using the likelihood ratio test. This test uses chi-square
1213 (X^2) to compare the difference between the likelihood ratio of the full model (i.e the
1214 model including predictors) and the null model based on the -2log likelihood. If the
1215 *p-value* was less than the chosen alpha level, which for this test was .05, then the null
1216 hypothesis was rejected suggesting that the alternative model (i.e. the model including
1217 predictors) was a better model for predicting respondents' behavioural intention.

1218

1219 If the null model was rejected, then the Wald chi-square statistic was used to test
1220 the significance of the individual regression coefficients (β) (i.e. the individual predictor
1221 variables which are expressed in log odds). If the *p-value* for an individual predictor
1222 variable was less than the chosen alpha level, which for this test was .05, then the null
1223 hypothesis was rejected suggesting that the predictor variable provides a significant
1224 contribution to the model. For ease of interpretation, the exponent of the regression
1225 coefficients ($\text{Exp}(\beta)$), which is the odds ratio, was used to interpret the results. For
1226 example, a one unit increase in the predictor variable increases the odds of a farmers'
1227 preferred behavioural intention by the value of $\text{Exp}(\beta)$.

1228

1229 In addition, the Hosmer-Lemeshow inferential goodness-of-fit statistic was used to
1230 assess the null hypothesis that there was no difference between observed and predicted
1231 values. If the *p-value* was greater than the chosen alpha level, which for this test was
1232 .05, then the null hypothesis could not be rejected suggesting that the model estimates
1233 fit the data at an acceptable level (Hosmer and Lemeshow, 1980). Furthermore, the
1234 two descriptive measures of goodness-of-fit which were reported included R^2 indices
1235 based on Cox and Snell (1989) and Nagelkerke (1991). However, it is important to note
1236 that although these approximate, or pseudo R^2 values, are variations of the R^2 concept
1237 used in linear regression, which explains the proportion of variation in the dependent
1238 variable explained by the predictor variables, they are not directly comparable (Peng

1239 et al., 2002). Lastly, classification tables are presented to validate the predicted prob-
1240 abilities.

1241

1242 **4.3 Results and discussion**

1243 The section is presented in two parts in relation to the statistical analysis conducted.
1244 The first part discusses the respondents farm characteristics which includes: farm at-
1245 tributes; social demographics; licence characteristics; and past abstractions between
1246 2008 to 2012. Farm attributes and social demographics were obtained from the survey
1247 (section A and D respectively) whilst licence characteristics and past abstractions, for
1248 those who responded to the survey, were obtained from the EA. The second part dis-
1249 cusses the respondents' behavioural intentions, for strategic (long-term) and in-season
1250 (short-term) scenarios, which were obtained from the survey (sections B and C re-
1251 spectively). Summary statistics with regards to farm characteristics and behavioural
1252 intentions are presented in Appendix B and Appendix C respectively. Overall, not
1253 including the random sample of farmers who participated in the pilot survey, 11 % of
1254 the remaining population of farmers in the study area responded, providing a sample
1255 size of 90 farmers.

1256

1257 **4.3.1 Farm characteristics**

1258 **Farm attributes**

1259 Table 4.2 provides a summary of the respondents' farm attributes including: the per-
1260 centage of crops which were irrigated; the percentage of predominant soil types which
1261 were irrigated; and the percentage of irrigation application mechanisms used. Main
1262 crop potatoes (41 % of respondents) and vegetables (33 %) were the most widely ir-
1263 rigated crops. Sand and loam were the most widely irrigated predominant soil types,
1264 36 % and 35 % respectively. Whilst rainguns were the most widely used irrigation
1265 application mechanism used (60 %). In addition, the respondents total irrigated area,
1266 based on their three main irrigated crops, was 12,659 ha, with main crop potatoes and
1267 vegetables accounting for 46 % and 35 % respectively. Lastly, 49 % of the respondents
1268 had a combined storage capacity of 10,134 Ml.

Table 4.2: Respondents' farm attributes ($n = 90$)

Measure	Respondents (%)
Irrigated crop	
- Early potatoes	5
- Main crop potatoes	41
- Sugar beet	11
- Orchard fruit	1
- Small fruit	1
- Vegetables	33
- Grass	1
- Cereals	7
- Other	2
Predominant soil type	
- Sand	36
- Loam	35
- Peat	16
- Mixed	14
Irrigation application used	
- Raingun	60
- Booms	18
- Sprinkle	2
- Trickle	1
- Mixed	19
Summing errors due to rounding	

1269 Figure 4.2 illustrates that the distribution of irrigated areas (top) and storage ca-
1270 pacities (bottom) of the respondents were positively skewed. Irrigated areas ranged
1271 from 3 ha to 1,558 ha, with a median of 47 ha and an interquartile range (IQR) of
1272 116 ha. Storage capacities ranged from 5 ML to 2,273 ML, with a median of 107 ML
1273 and an IQR of 217 ML. Furthermore, 25 % of respondents irrigated less than 25 ha
1274 and 75 % of respondents irrigated less than 141 ha. With regards to those respondents
1275 with storage, 25 % had storage capacities of less than 49 ML whilst 75 % had storage
1276 capacities of less than 266 ML. In addition, for the 49 % of respondents who had stor-
1277 age capacity, there was a strong positive correlation between ranked irrigated area and
1278 ranked storage capacity ($r_s = .76$), which was also statistically significant ($p = <.001$),
1279 thus the null hypothesis could be rejected (see Appendix B).

1280

1281 These results correspond with findings presented by Knox et al. (2010) for the An-
1282 glian region who reported an increase in high value crops, between 1990 and 2010, such

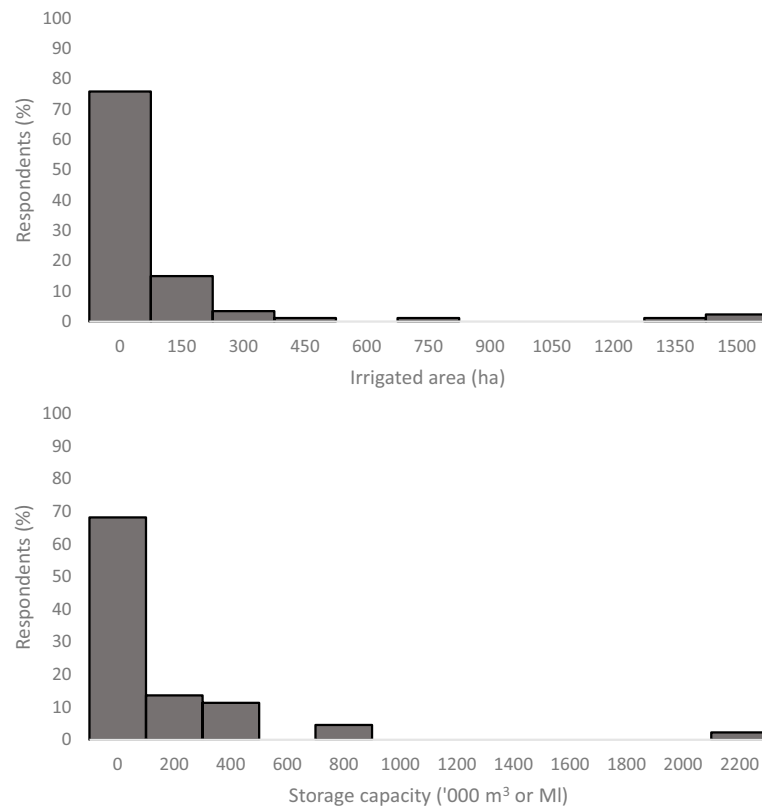


Figure 4.2: Relative frequency histograms of respondents' irrigated areas and storage capacities. (Top) A relative frequency histogram of respondents' irrigated areas. (Bottom) A relative frequency histogram of respondents' storage capacities for those who had storage

1283 as main crop potatoes and vegetables. Furthermore, sand and loam were found to be
 1284 common soil types present in this region (Hodge et al., 1984). Moreover, approximately
 1285 67 % of farmers in England use ranguns as their preferred irrigation application mech-
 1286 anism and 42 % have some level of storage capacity (Weatherhead and Rivas Casado,
 1287 2007).

1288

1289 Social demographics

1290 Table 4.3 provides a summary of the respondents' social demographics including: gen-
 1291 der; age; level of education acquired (measured by National Qualification Framework
 1292 level for England, Wales and Northern Ireland); and total gross annual farm business
 1293 income. The majority of respondents were male (98 %) and aged between 40 and 59 (50

Table 4.3: Respondents' social demographics ($n = 90$)

Measure	Respondents (%)
Gender	
- Male	98
- Female	2
Age	
- 18-39	15
- 40-59	50
- ≥ 60	35
^a Education	
- ≤ 2	37
- 3-5	25
- ≥ 6	38
^b Income	
- £0-49,999	5
- £50-99,999	18
- \geq £100,000	77

^aNational Qualification Framework (NQF) levels for England, Wales and Northern Ireland

^bTotal gross annual farm business income

1294 %). In regards to the level of education acquired the respondents were largely divided
 1295 between those who achieved level two or less and those who achieved level six or more
 1296 (37 % and 38 % respectively). Furthermore, 77 % of respondents reported a total gross
 1297 annual farm business income of more than £100,000.

1298

1299 These results correspond with findings from the farm structure survey in 2013 which
 1300 found that 83 % of farm managers in England were male and 58 % were aged between
 1301 45 and 64 (DEFRA, 2013). Furthermore, although reported farm incomes were greater
 1302 than the average farm income in England (£78,190 in 2010), these are typical within
 1303 the Anglian region where farms are inherently more profitable (DEFRA, 2014a).

1304

1305 **Licence characteristics**

1306 Table 4.4 provides a summary of the respondents' licence characteristics including:
 1307 the percentage of farmers who actively used their licence within the last 10 years;

Table 4.4: Respondents' licence characteristics ($n = 90$)

Measure	Respondents (%)
Licence use (last 10 years)	
- Yes	98
- No	2
Licence type	
- Direct	51
- Storage	9
- Direct and storage	40
Licence source	
- Surface water	78
- Groundwater	22
Time limited	
- Yes	76
- No	24
Summing errors due to rounding	

1308 licence type; licence source; and whether licences were time limited or not. The ma-
1309 jority of farmers actively used their licence within the last 10 years (98 %). However,
1310 with regards to the percentage of respondents who had storage capacity, a discrepancy
1311 occurred between the additional data provided by the EA (31 %) and those who re-
1312 sponded to the survey (49 %). It was assumed that the questionnaire data provided
1313 a more reliable source of information. Therefore, the additional respondents who had
1314 storage were assumed to have updated their licences from direct only to direct and
1315 storage, as both data sources reported the same respondents with storage only licences.
1316 Whereas the majority of respondents (51 %) owned direct licences, which are gener-
1317 ally used during the growing season, only 9 % owned storage only licences, which are
1318 generally used during the winter to refill farm storage reservoirs, whilst the remainder
1319 (40 %) were able to abstract during the whole year as they owned direct and storage
1320 licences. Furthermore, 78 % of respondents owned surface water licences, with the
1321 remainder owning groundwater licences, and 76 % owned time limited licences with
1322 the remainder owning licences which were not time limited. Jointly the respondents
1323 total annual licence limit was 6,645 Ml, whilst their total daily licence limit was 171 Ml.
1324

1325 Figure 4.3 illustrates that the distribution of annual (top) and daily (bottom) li-
1326 cence limits of the respondents were positively skewed, similar to irrigated areas and

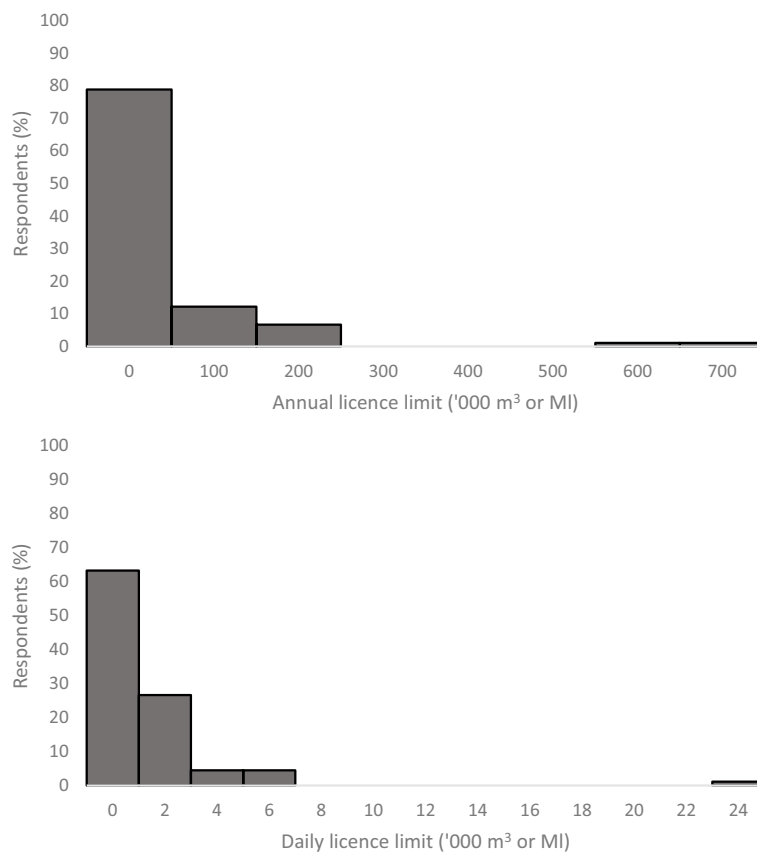


Figure 4.3: Relative frequency histograms of respondents' annual and daily licence limits. (Top) A relative frequency histogram of respondents' annual licence limits. (Bottom) A relative frequency histogram of respondents' daily licence limits

1327 storage capacities. Annual licence limits ranged from 1 MI to 727 MI, with a median of
 1328 45 MI and an IQR of 67 MI. Daily licence limits ranged from <1 MI to 25 MI, with a
 1329 median of 1 MI and an IQR of 2 MI. Furthermore, 25 % of respondents annual licence
 1330 limits were less than 17 MI and 75 % of respondents annual licence limits were less than
 1331 84 MI. Similarly, 25 % of respondents daily licence limits were less than 1 MI and 75 %
 1332 of respondents daily licence limits were less than 2 MI. In addition, there was a strong
 1333 positive correlation between ranked annual licence limit and ranked daily licence limit
 1334 ($r_s = .70$), which was also statistically significant ($p = <.001$), thus the null hypothesis
 1335 could be rejected (see Appendix B).

1337 Past abstractions (2008 to 2012)

1338 On average, the respondents abstracted 36 % of their annual licensed volume between
1339 2008 and 2012 (i.e. 2,395 Ml) (see Appendix B). However, Figure 4.4 (top and middle)
1340 illustrate that annual licensed abstractions varied during this period with 51 % being
1341 abstracted in 2011 and 24 % in 2012, corresponding with the driest and wettest years
1342 since 1973 respectively (see Chapter 3). Furthermore, Figure 4.4 (bottom) illustrates
1343 that the respondents' average monthly abstractions, during the same period, varied
1344 with the largest volume abstracted in June (435 Ml) and the smallest in October (39
1345 Ml). However, Figure 4.4 (bottom) also indicates two general periods of abstractions
1346 in the year. The first corresponds with the growing season (i.e. April to October)
1347 when the majority of direct licence users can abstract, accounting for 70 % of average
1348 monthly abstractions. The second corresponds with the period outside of the growing
1349 season (i.e. November to March) when the majority of storage only licence users can
1350 abstract.

1351

1352 In particular, there was a strong, negative, linear correlation between annual ab-
1353 stractions and precipitation ($r = -.91$), which was also statistically significant ($p = < .05$),
1354 thus this null hypothesis could be rejected (see Appendix B). However, there was a weak
1355 correlation, with no direction or form, between average monthly abstractions and pre-
1356 cipitation ($r = .36$), which was not statistically significant ($p = > .05$), thus this null
1357 hypothesis could not be rejected (see Appendix B). Furthermore, as there were two
1358 general periods of abstractions in the year, during the growing season and outside, two
1359 further tests of association were conducted respectively. During the growing season,
1360 there was a weak correlation, with no direction or form, between average monthly ab-
1361 stractions and precipitation ($r = .36$), which was not statistically significant ($p = > .05$),
1362 thus this null hypothesis could not be rejected (see Appendix B). However, outside
1363 of the growing season, there was a strong, positive, linear correlation between aver-
1364 age monthly abstractions and precipitation ($r = .87$), although it was not statistically
1365 significant ($p = > .05$), thus this null hypothesis could not be rejected (see Appendix B).

1366

1367 These results indicate that although annual abstractions increased when precipi-
1368 tation decreased, other factors had a strong influence on monthly abstractions other
1369 than precipitation. During the growing season, a major contributing factor would likely

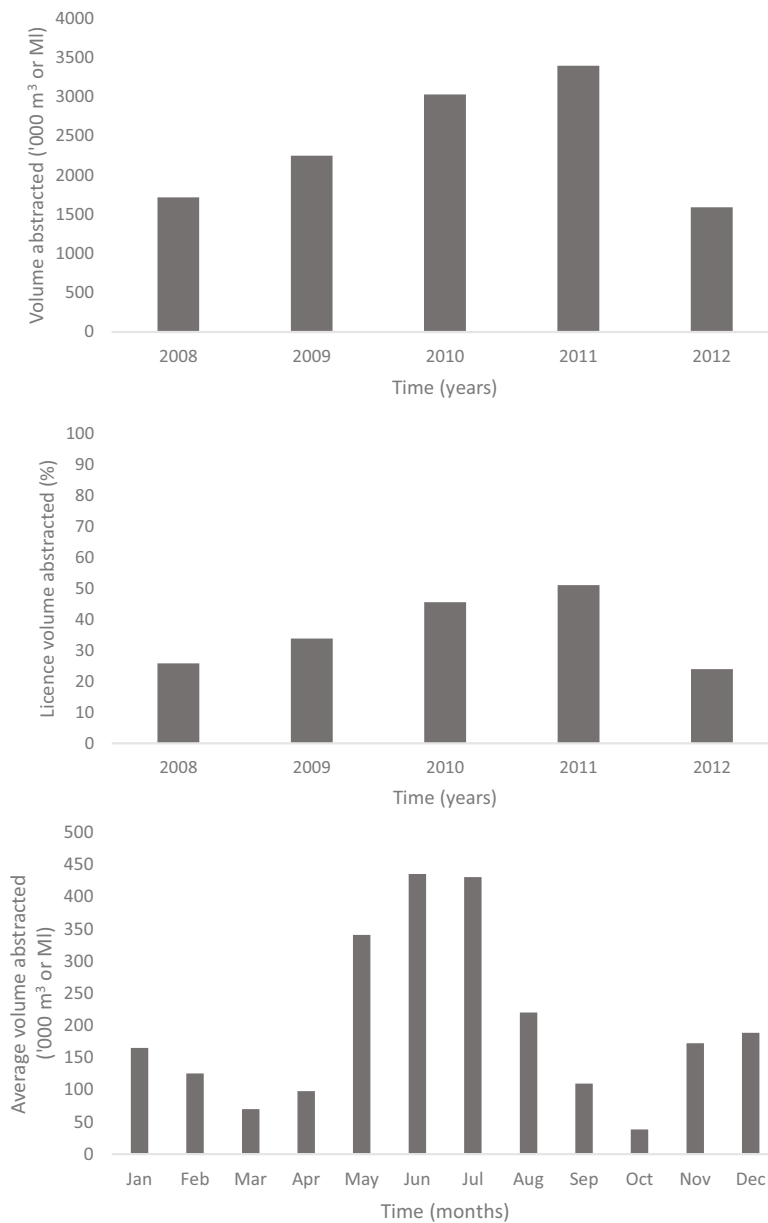


Figure 4.4: Bar charts illustrating respondents' annual, licensed, and average monthly abstractions (2008 to 2012). (Top) A bar chart illustrating respondents' annual abstractions (2008 to 2012). (Middle) A bar chart illustrating the percentage of licence volume abstracted by respondents (2008 to 2012). (Bottom) A bar chart illustrating respondents' average monthly abstractions (2008 to 2012)

1370 include water availability, as less precipitation also means farmers are unable to always
 1371 abstract water due to HoF restrictions. Outside of the growing season, water avail-
 1372 ability is less of an issue due to increased precipitation. Thus, a stronger association

1373 existed between average monthly abstractions and precipitation, although not signif-
1374 icant. Overall, these results indicate that the farm characteristics of the respondents
1375 are largely representative of farmers in the Anglian region, and also in England more
1376 generally.

1377

1378 **4.3.2 Behavioural intentions**

1379 The respondents' ranked preferred behavioural intentions for each of the four scenarios
1380 are presented in Figure 4.5. LWUO were preferred under all four scenarios: increasing
1381 application efficiency under a strategic (long-term)/water shortage scenario; selling sur-
1382 plus water for the duration of the growing season under a strategic (long-term)/water
1383 surplus scenario; only using their maximum abstraction licence to irrigate their most
1384 valuable crops under an in-season (short-term)/water shortage scenario; and just using
1385 their abstraction licence to meet crop water requirements and leaving the remainder of
1386 their licence unused under an in-season (short-term)/water surplus scenario. In addi-
1387 tion, groups interested in crop type and quality were most influential in respondents
1388 decision-making under all scenarios (see Appendix C).

1389

1390 **Preferred strategic (long-term)/water shortage behavioural intention**

1391 In regards to this scenario, 88 % to 97 % of the respondents answered each Likert item,
1392 except for past behaviour, where only 6 % to 8 % answered, due to the fact that only
1393 9 % actually experienced this scenario. Figure 4.5 (top left) indicates that under this
1394 scenario the respondents' ranked preferred behavioural intentions, from most preferred
1395 to least preferred, were as follows: increase application efficiency (LWUO); increase
1396 storage capacity (HWUO); apply for a larger abstraction licence (HWUO); buy more
1397 water for the duration of the growing season (HWUO); grow less water intensive crops
1398 (LWUO); grow the same crops but over a smaller area (LWUO); and lastly change
1399 nothing (LWUO).

1400

1401 Furthermore, a significant median decrease existed between intention to increase
1402 application efficiency and intention to increase storage capacity; intention to increase
1403 storage capacity and intention to apply for a larger abstraction licence; intention to

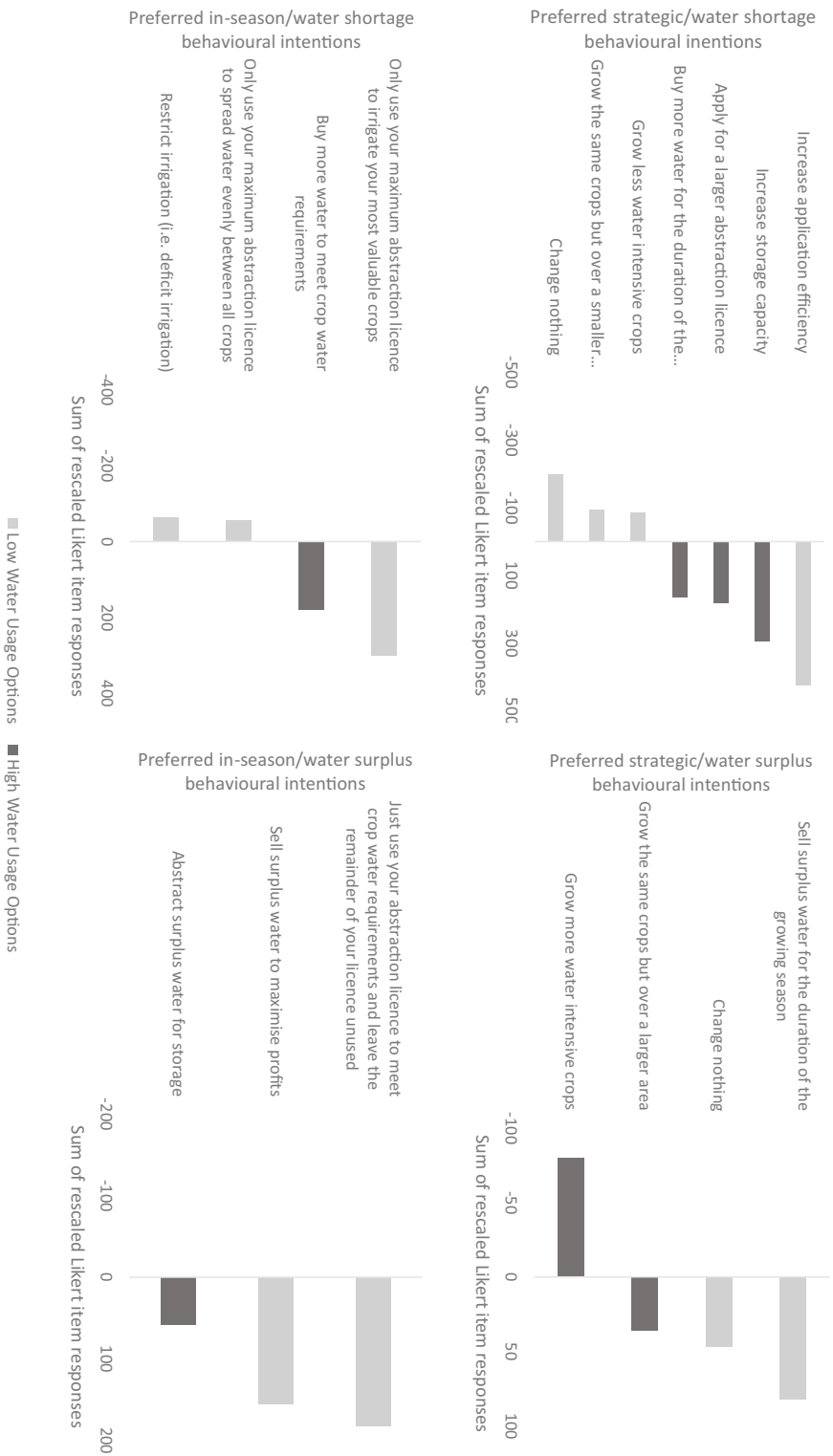


Figure 4.5: Bar charts illustrating respondents' preferred behavioural intentions during strategic (long-term) and in-season (short-term) water shortage and surplus scenarios. (Top left) A bar chart illustrating respondents' preferred strategic (long-term)/water shortage behavioural intentions. (Top right) A bar chart illustrating respondents' preferred strategic (long-term)/water surplus behavioural intentions. (Bottom left) A bar chart illustrating respondents' preferred in-season (short-term)/water shortage behavioural intentions. (Bottom right) A bar chart illustrating respondents' preferred in-season (short-term)/water surplus behavioural intentions

Table 4.5: A binary logistic regression model with regards to respondents' preferred strategic (long-term)/water shortage behavioural intention ($n = 79$)

^a Variables	β	SE	Wald (X^2)	df	Sig.	Exp(β)	95 % C.I. for Exp(β)	
							Lower	Upper
Constant	-1.765	.476	13.748	1	.000	.171		
- Attitude	1.146	.545	4.419	1	.036	3.146	1.081	9.159
- SN	1.641	.547	9.009	1	.003	5.160	1.767	15.068
- PBC	.431	.567	.577	1	.448	1.538	.506	4.676

Coefficient (β); Standard error of estimate (SE); Wald chi-square value (Wald (X^2)); Degree of freedom (df); Significance (Sig.); Exponentiated coefficient (Exp(β)); Confidence interval (C.I.)

^a Subjective norm (SN); Perceived Behavioural Control (PBC)

Model chi-square (X^2) = 22.866 (3 df), $p = .000$, (-2 log likelihood = 85.623).

Hosmer & Lemeshow (X^2) = 1.970 (5 df), $p = .853$. Cox and Snell $R^2 = .251$.

Nagelkerke $R^2 = .337$

1404 buy more water for the duration of the growing season and intention to grow less wa-
 1405 ter intensive crops; and intention to grow the same crops but over a smaller area and
 1406 intention to change nothing. As $p = <.05$ in each case, the null hypotheses could be
 1407 rejected for each pair of these behaviours. In addition, a significant median increase
 1408 existed between intention to grow less water intensive crops and intention to grow the
 1409 same crops but over a smaller area, $p = <.05$, thus the null hypothesis between this
 1410 pair of behaviours could also be rejected (see Appendix C).

1411

1412 A binary logistic regression was performed to examine the effects of attitude, sub-
 1413 jective norm, and perceived behavioural control on the likelihood of increasing respon-
 1414 dents' intention to increase application efficiency. The logistic regression model was
 1415 statistically significant, $X^2(3) = 22.866$, $p = <.001$, thus the null hypothesis could be
 1416 rejected (see Table 4.5). Attitude and subjective norm were both significant predictors
 1417 of intention, $p = <.05$ and $p = <.01$ respectively, thus the null hypotheses for these
 1418 predictor variables could be rejected. However, perceived behavioural control was not
 1419 a significant predictor of intention, $p = >.05$, thus the null hypothesis for this predictor
 1420 could not be rejected. Moreover, a one unit increase in attitude increased the odds of a
 1421 farmers' intention to increase application efficiency by 3.146 times. A one unit increase
 1422 in subjective norm increased the odds of a farmers' intention to increase application
 1423 efficiency by 5.160 times.

1424

Table 4.6: Classification table with regards to respondents' preferred strategic (long-term)/water shortage behavioural intention ($n = 79$)

Observed Intention	Predicted Intention		% correct
	0	1	
0	37	7	84.1
1	12	23	65.7
Overall %			75.9

The cutoff value is .5

1425 Furthermore, no multi-collinearity existed between predictor variables, the number
 1426 of cases per predictor variables were greater than the recommended value, and no out-
 1427 liers were present, based on the thresholds used in this study (see Appendix C). The
 1428 Hosmer-Lemeshow goodness-of-fit statistic was insignificant indicating that the model
 1429 fit the observed data well, $X^2(5) = 1.970$, $p = .853$, thus the null hypothesis could not
 1430 be rejected. Furthermore, the model explained 34 % (Nagelkerke R^2) of the variance
 1431 in intention to increase application efficiency and correctly classified 76 % of cases,
 1432 an improvement over the null model which only classified 56 % of cases correctly (see
 1433 Table 4.6).

1434

1435 Preferred strategic (long-term)/water surplus behavioural intention

1436 In regards to this scenario, 86 % to 96 % of the respondents answered each Likert item,
 1437 except for past behaviour, where only 42 % to 50 % answered, due to the fact that only
 1438 51 % actually experienced this scenario (noticeably more than the previous strategic
 1439 (long-term)/water shortage scenario). Figure 4.5 (top right) indicates that under this
 1440 scenario the respondents' ranked preferred behavioural intentions, from most preferred
 1441 to least preferred, were as follows: sell surplus water for the duration of the growing
 1442 season (LWUO); change nothing (LWUO); grow the same crops but over a larger area
 1443 (HWUO); and lastly grow more water intensive crops (HWUO).

1444

1445 Furthermore, a significant median decrease existed between intention to change
 1446 nothing and intention to grow the same crops but over a larger area; and intention to
 1447 grow the same crops but over a larger area and intention to grow more water intensive
 1448 crops. As $p = <.05$ for both of these cases, the null hypotheses could be rejected for

Table 4.7: A binary logistic regression model with regards to respondents' preferred strategic (long-term)/water surplus behavioural intention ($n = 75$)

^a Variables	β	SE	Wald (X^2)	df	Sig.	Exp(β)	95 % C.I. for Exp(β)	
							Lower	Upper
Constant	-1.085	.387	7.850	1	.005	.338		
- Attitude	2.588	1.107	5.463	1	.019	13.309	1.519	116.635
- SN	1.275	.528	5.821	1	.016	3.579	1.270	10.083
- PBC	.535	.643	.693	1	.405	1.708	.484	6.022

Coefficient (β); Standard error of estimate (SE); Wald chi-square value (Wald (X^2)); Degree of freedom (df); Significance (Sig.); Exponentiated coefficient (Exp(β)); Confidence interval (C.I.)

^a Subjective norm (SN); Perceived Behavioural Control (PBC)

Model chi-square (X^2) = 18.340 (3 df), $p = .000$, (-2 log likelihood = 85.512).

Hosmer & Lemeshow (X^2) = 2.621 (3 df), $p = .454$. Cox and Snell $R^2 = .217$.

Nagelkerke $R^2 = .289$

1449 each pair of these behaviours (see Appendix C). However, although no significant me-
 1450 dian difference existed between intention to sell surplus water for the duration of the
 1451 growing season and intention to change nothing, the prior behaviour was ranked higher.

1452

1453 A binary logistic regression was performed to examine the effects of attitude, sub-
 1454 jective norm, and perceived behavioural control on the likelihood of increasing respon-
 1455 dents' intention to sell surplus water for the duration of the growing season. The logistic
 1456 regression model was statistically significant, $X^2(3) = 18.340$, $p = <.001$, thus the null
 1457 hypothesis could be rejected (see Table 4.7). Attitude and subjective norm were both
 1458 significant predictors of intention, $p = <.05$ for both predictors, thus the null hypotheses
 1459 for these predictor variables could be rejected. However, perceived behavioural control
 1460 was not a significant predictor of intention, $p = >.05$, thus the null hypothesis for this
 1461 predictor could not be rejected. Moreover, a one unit increase in attitude increased the
 1462 odds of a farmers' intention to sell surplus water for the duration of the growing season
 1463 by 13.309 times. A one unit increase in subjective norm increased the odds of a farm-
 1464 ers' intention to sell surplus water for the duration of the growing season by 3.579 times.

1465

1466 Furthermore, no multi-collinearity existed between predictor variables, the number
 1467 of cases per predictor variables were greater than the recommended value, and no out-
 1468 liers were present, based on the thresholds used in this study (see Appendix C). The
 1469 Hosmer-Lemeshow goodness-of-fit statistic was insignificant indicating that the model

Table 4.8: Classification table with regards to respondents' preferred strategic (long-term)/water surplus behavioural intention ($n = 75$)

Observed Intention	Predicted Intention		% correct
	0	1	
0	26	13	66.7
1	10	26	72.2
Overall %			69.3

The cutoff value is .5

1470 fit the observed data well, $X^2(3) = 2.621$, $p = .454$, thus the null hypothesis could not
 1471 be rejected. Furthermore, the model explained 29 % (Nagelkerke R^2) of the variance
 1472 in intention to sell surplus water for the duration of the growing season and correctly
 1473 classified 69 % of cases, an improvement over the null model which only classified 52
 1474 % of cases correctly (see Table 4.8).

1475

1476 Preferred in-season (short-term)/water shortage behavioural intention

1477 In regards to this scenario, 87 % to 94 % of the respondents answered each Likert item,
 1478 except for past behaviour, where only 12 % answered, due to the fact that only 12 %
 1479 actually experienced this scenario, similar to the previous strategic (long-term)/water
 1480 shortage scenario. Figure 4.5 (bottom left) indicates that under this scenario the re-
 1481 spondents' ranked preferred behavioural intentions, from most preferred to least pre-
 1482 ferred, were as follows: only use your maximum abstraction licence to irrigate your most
 1483 valuable crops (LWUO); buy more water to meet crop water requirements (HWUO);
 1484 only use your maximum abstraction licence to spread water evenly between all crops
 1485 (LWUO); and lastly restrict irrigation (i.e. deficit irrigation) (LWUO).

1486

1487 Furthermore, a significant median decrease existed between intention to only use
 1488 your maximum abstraction licence to irrigate your most valuable crops and intention
 1489 to buy more water to meet crop water requirements; and intention to buy more water
 1490 to meet crop water requirements and intention to only use your maximum abstraction
 1491 licence to spread water evenly between all crops. As $p = <.05$ for both of these cases,
 1492 the null hypotheses could be rejected for each pair of these behaviours (see Appendix C).

1493

Table 4.9: A binary logistic regression model with regards to respondents' preferred in-season (short-term)/water shortage behavioural intention ($n = 78$)

^a Variables	β	SE	Wald (X^2)	df	Sig.	Exp(β)	95 % C.I. for Exp(β)	
							Lower	Upper
Constant	-3.108	.680	20.862	1	.000	.045		
- Attitude	1.588	.750	4.479	1	.034	4.896	1.125	21.312
- SN	2.031	.697	8.479	1	.004	7.619	1.942	29.885
- PBC	1.356	.653	4.314	1	.038	3.882	1.079	13.958

Coefficient (β); Standard error of estimate (SE); Wald chi-square value (Wald (X^2)); Degree of freedom (df); Significance (Sig.); Exponentiated coefficient (Exp(β)); Confidence interval (C.I.)

^a Subjective norm (SN); Perceived Behavioural Control (PBC)

Model chi-square (X^2) = 28.167 (3 df), $p = .000$, (-2 log likelihood = 60.639).

Hosmer & Lemeshow (X^2) = 5.055 (4 df), $p = .282$. Cox and Snell $R^2 = .303$.

Nagelkerke $R^2 = .446$

1494 A binary logistic regression was performed to examine the effects of attitude, sub-
1495 jective norm, and perceived behavioural control on the likelihood of increasing respon-
1496 dents' intention to only use their maximum abstraction licence to irrigate their most
1497 valuable crops. The logistic regression model was statistically significant, $X^2(3) =$
1498 28.167, $p = <.001$, thus the null hypothesis could be rejected (see Table 4.9). Attitude,
1499 subjective norm, and perceived behavioural control were all significant predictors of
1500 intention, $p = <.05$ for all predictors, thus the null hypotheses for all of the predictor
1501 variables could be rejected. Moreover, a one unit increase in attitude increased the
1502 odds of a farmers' intention to only use their maximum abstraction licence to irrigate
1503 their most valuable crops by 4.896 times. A one unit increase in subjective norm in-
1504 creased the odds of a farmers' intention to only use their maximum abstraction licence
1505 to irrigate their most valuable crops by 7.619 times. Lastly, a one unit increase in per-
1506 ceived behavioural control increased the odds of a farmers' intention to only use their
1507 maximum abstraction licence to irrigate their most valuable crops by 3.882 times.

1508

1509 However, although no multi-collinearity existed between predictor variables, and
1510 the number of cases per predictor variables were greater than the recommended value,
1511 there was one noticeable outlier present, based on the thresholds used in this study (see
1512 Appendix C). This outlier represented the only respondent who was categorised as one
1513 for intention but zero for each of the predictors. Therefore, a second binary logistic
1514 regression was performed, after removing the outlier, to compare changes in the model.

Table 4.10: A binary logistic regression model with regards to respondents' preferred in-season (short-term)/water shortage behavioural intention after removing an outlier ($n = 77$)

^a Variables	β	SE	Wald (X^2)	df	Sig.	Exp(β)	95 % C.I. for Exp(β) Lower	Upper
Constant	-3.643	.817	19.876	1	.000	.026		
- Attitude	1.708	.774	4.876	1	.027	5.519	1.212	25.139
- SN	2.403	.777	9.556	1	.002	11.056	2.410	50.730
- PBC	1.638	.707	5.366	1	.021	5.146	1.287	20.577

Coefficient (β); Standard error of estimate (SE); Wald chi-square value (Wald (X^2)); Degree of freedom (df); Significance (Sig.); Exponentiated coefficient (Exp(β)); Confidence interval (C.I.)

^a Subjective norm (SN); Perceived Behavioural Control (PBC)

Model chi-square (X^2) = 32.129 (3 df), $p = .000$, (-2 log likelihood = 53.853).

Hosmer & Lemeshow (X^2) = 7.215 (4 df), $p = .125$. Cox and Snell $R^2 = .342$.

Nagelkerke $R^2 = .508$

1515 The second binary logistic regression model was statistically significant, $X^2(3) =$
1516 32.129, $p = <.001$, thus the null hypothesis could be rejected (see Table 4.10). Atti-
1517 tude, subjective norm, and perceived behavioural control all continued to be significant
1518 predictors of intention, $p = <.05$ for all predictors, thus the null hypotheses for all of
1519 the predictor variables could be rejected. Moreover, a one unit increase in attitude in-
1520 creased the odds of a farmers' intention to only use their maximum abstraction licence
1521 to irrigate their most valuable crops by 5.519 times. A one unit increase in subjective
1522 norm increased the odds of a farmers' intention to only use their maximum abstraction
1523 licence to irrigate their most valuable crops by 11.056 times. Lastly, a one unit increase
1524 in perceived behavioural control increased the odds of a farmers' intention to only use
1525 their maximum abstraction licence to irrigate their most valuable crops by 5.146 times.
1526 Therefore, removing the one noticeable outlier increased the odds ratios of each of the
1527 predictor variables.

1528

1529 Furthermore, no multi-collinearity existed between predictor variables, the number
1530 of cases per predictor variables were greater than the recommended value, and no out-
1531 liers were present, based on the thresholds used in this study (see Appendix C). The
1532 Hosmer-Lemeshow goodness-of-fit statistic was insignificant indicating that the model
1533 fit the observed data well, $X^2(4) = 7.215$, $p = .125$, thus the null hypothesis could not
1534 be rejected. Furthermore, the model explained 51 % (Nagelkerke R^2) of the variance
1535 in intention to sell surplus water for the duration of the growing season and correctly

Table 4.11: Classification table with regards to respondents' preferred in-season (short-term)/water shortage behavioural intention ($n = 77$)

Observed Intention	Predicted Intention		% correct
	0	1	
0	52	6	89.7
1	5	14	73.7
Overall %			85.7

The cutoff value is .5

1536 classified 86 % of cases, an improvement over the null model which only classified 75
 1537 % of cases correctly (see Table 4.11).

1538

1539 Preferred in-season (short-term)/water surplus behavioural intention

1540 In regards to this scenario, 79 % to 91 % of the respondents answered each Likert
 1541 item, except for past behaviour, where only 42 % to 50 % answered, due to the fact
 1542 that only 53 % actually experienced this scenario, which is noticeably more than the
 1543 previous in-season (short-term)/water shortage scenario and similar to the strategic
 1544 (long-term)/water surplus scenario. Figure 4.5 (bottom right) indicates that under
 1545 this scenario the respondents' ranked preferred behavioural intentions, from most preferred
 1546 to least preferred, were as follows: just use your abstraction licence to meet crop
 1547 water requirements and leave the remainder of your licence unused (LWUO); sell surplus
 1548 water to maximise profits (LWUO); and lastly abstract surplus water for storage
 1549 (HWUO).

1550

1551 The only significant median decrease existed between intention to sell surplus wa-
 1552 ter to maximise profits and intention to abstract surplus water for storage, as $p =$
 1553 $< .05$. Therefore the null hypotheses could be rejected for this pair of behaviours (see
 1554 Appendix C). However, although no significant median difference existed between in-
 1555 tention to just use your abstraction licence to meet crop water requirements and leave
 1556 the remainder of your licence unused and intention to sell surplus water to maximise
 1557 profits, the prior behaviour was ranked higher, similar to the previous strategic (long-
 1558 term)/water surplus scenario.

1559

Table 4.12: A binary logistic regression model with regards to respondents' preferred in-season (short-term)/water surplus behavioural intention ($n = 72$)

^a Variables	β	SE	Wald (X^2)	df	Sig.	$\text{Exp}(\beta)$	95 % C.I. for $\text{Exp}(\beta)$	
							Lower	Upper
Constant	-21.819	6027.378	.000	1	.997	.000		
- Attitude	$+\infty$	6027.378	.000	1	.997	$+\infty$	-	-
- SN	-.431	.826	.272	1	.602	.650	.129	3.281
- PBC	$+\infty$	6027.378	.000	1	.997	$+\infty$	-	-

Coefficient (β); Standard error of estimate (SE); Wald chi-square value (Wald (X^2)); Degree of freedom (df); Significance (Sig.); Exponentiated coefficient ($\text{Exp}(\beta)$); Confidence interval (C.I.)

^a Subjective norm (SN); Perceived Behavioural Control (PBC)

Model chi-square (X^2) = 40.960 (3 df), $p = .000$, (-2 log likelihood = 37.744).

Hosmer & Lemeshow (X^2) = 2.003 (5 df), $p = .849$. Cox and Snell $R^2 = .434$.

Nagelkerke $R^2 = .653$

1560 A binary logistic regression was performed to examine the effects of attitude, sub-
 1561 jective norm, and perceived behavioural control on the likelihood of increasing respon-
 1562 dents' intention to just use their abstraction licence to meet crop water requirements
 1563 and leave the remainder of their licence unused. However, the model failed to converge
 1564 after 20 iterations suggesting that one or more of the predictor variables were causing
 1565 complete or quasi-complete separation, due to the sample size being too small. On
 1566 inspection of the iteration history it was clear that attitude and perceived behavioural
 1567 control were the responsible variables (see Appendix C). Although the most common
 1568 approach is to remove the variable or variables causing the issue, others suggest a more
 1569 desirable approach is to simply do nothing and report the results of the full model,
 1570 largely due to interaction effects between predictor variables. Either approach should
 1571 be made in consideration to the original research question attempting to be addressed.
 1572 Therefore, as this study was interested in understanding the relative importance of each
 1573 of the underlying constructs, the latter approach was adopted (Allison, 2008).

1574

1575 The binary logistic regression model was statistically significant, $X^2(3) = 40.960$, p
 1576 = $<.001$, thus the null hypothesis could be rejected (see Table 4.12). However, none of
 1577 the predictor variables were individually significant, $p = >.05$ for all predictors, thus the
 1578 null hypothesis could not be rejected for any of the predictor variables. The standard
 1579 errors and Wald chi-square statistics for attitude and perceived behavioural control are
 1580 certainly incorrect but those for subjective norm remain valid (Allison, 2008).

Table 4.13: Classification table with regards to respondents' preferred in-season (short-term)/water surplus behavioural intention ($n = 72$)

Observed Intention	Predicted Intention		% correct
	0	1	
0	55	0	100.0
1	9	8	47.1
Overall %			87.5

The cutoff value is .5

1581 Furthermore, no multi-collinearity existed between predictor variables, the number
 1582 of cases per predictor variables were greater than the recommended value, and no out-
 1583 liers were present, based on the thresholds used in this study (see Appendix C). The
 1584 Hosmer-Lemeshow goodness-of-fit statistic was insignificant indicating that the model
 1585 fit the observed data well, $X^2(5) = 2.003$, $p = .849$, thus the null hypothesis could not
 1586 be rejected. Furthermore, the model explained 65 % (Nagelkerke R^2) of the variance
 1587 in intention to just use their abstraction licence to meet crop water requirements and
 1588 leave the remainder of their licence unused and correctly classified 88 % of cases, an
 1589 improvement over the null model which only classified 76 % of cases correctly (see Ta-
 1590 ble 4.13).

1591

1592 4.4 Summary

1593 Overall, this chapter has analysed empirical data which identified farmers' preferred
 1594 behavioural intentions under different scenarios of water shortage and surplus with re-
 1595 gards to the proposed water allocation systems in England. The results of this study
 1596 are intended to inform policy decision-makers involved in designing the proposed water
 1597 allocation systems by: 1) understanding how the proposed water allocation systems
 1598 might be received by farmers' on the ground; and 2) identifying which underlying pre-
 1599 dictors of intention most influence farmers' decision-making. Therefore, a questionnaire
 1600 was sent to 826 farmers in the study area of which 11 % responded. Analyses were
 1601 conducted in regards to respondents' farm characteristics and preferred behavioural
 1602 intentions based on the theoretical framework of the TPB.

1603

1604 Generally, respondents were representative of farmers in the Anglian region, and
1605 in England, as farm characteristics were similar to those presented in other studies
1606 (Hodge et al., 1984; Weatherhead and Rivas Casado, 2007; Knox et al., 2010; DE-
1607 FRA, 2013, 2014a). Under a strategic (long-term)/water shortage scenario farmers'
1608 preferred behavioural intention was to increase application efficiency. Under a strate-
1609 gic (long-term)/water surplus scenario farmers' preferred behavioural intentions was to
1610 sell surplus water for the duration of the growing season. Under an in-season (short-
1611 term)/water shortage scenario farmers' preferred behavioural intention was to only use
1612 their maximum abstraction licence to irrigated their most valuable crops. Lastly, under
1613 an in-season (short-term)/water surplus scenario farmers' preferred behavioural inten-
1614 tion was to just use their abstraction licence to meet crop water requirements and leave
1615 the remainder of their licence unused. Interestingly, all of the preferred behaviours were
1616 LWUO as defined in this study.

1617

1618 In addition, the results of the binary logistic regression analyses indicated that atti-
1619 tude and subjective norm were both significant predictors of farmers' intentions under
1620 three of the four scenarios (not under an in-season (short-term)/water surplus scenario).
1621 Perceived behavioural control was only a significant predictor under an in-season (short-
1622 term)/water shortage scenario. Interestingly, subjective norm had a larger influence on
1623 farmers' intentions under both strategic (long-term) and in-season (short-term)/water
1624 shortage scenarios, whilst attitude had a larger influence on farmers' intention under a
1625 strategic (long-term)/water surplus scenario. Past behaviour was not included in the
1626 regression analyses due to too few responses for this construct.

1627

1628 Therefore, the results of this study indicate that the proposed water allocation
1629 systems, which strongly encourage water licence trading, are not likely to be adopted
1630 quickly by farmers in the Anglian region based on their current preferred behavioural in-
1631 tentions, except with regards to selling surplus water during strategic (long-term)/water
1632 surplus scenarios. However, if attitude and subjective norm are assumed to be the most
1633 influential predictors of farmers' intentions, based on the results of this study, then it
1634 would not be surprising to see a gradual increase in the number of farmers trading as
1635 more farmer's increasingly understood the operations and the advantages, similar to
1636 what occurred in Australia when they first introduced water trading (Bjornlund, 2003).

1637 Finally, this study is intended to inform policy decision-makers involved with the
1638 design and implementation of the proposed water allocation systems in England. The
1639 results of this study have indicated how the proposals might be received by farmers
1640 on the ground under different strategic (long-term) and in-season (short-term) water
1641 shortage and surplus scenarios. It also identified the main underlying constructs of the
1642 TPB which have the greatest effect in influencing their decision-making.

1643

1644 Chapter 5

1645 A farm typology based on 1646 preferred behavioural intentions

1647 This chapter presents a farm typology based on the preferred behavioural intentions of
1648 farmers identified in the previous chapter. In particular this chapter offers one of the
1649 first typologies designed for policy decision-makers to understand how the proposed
1650 water allocation systems in England might be received by particular groups of farmers
1651 on the ground. As previously discussed, traditional farm typologies tend to concentrate
1652 on particular farm attributes such as farm income, size, or output. However, in respect
1653 to policy decision-making a growing number of researchers, and policy decision-makers,
1654 are beginning to realise the necessity of identifying farmers' behavioural intentions in
1655 aiding the design, implementation, and overall success of policy changes. Therefore,
1656 this chapter addresses the second of the three research questions presented in the in-
1657 troductory chapter:

1658

1659 **Research question 2:** *If farmers share similar behavioural intentions under dif-*
1660 *ferent scenarios how can traditional farm typologies incorporate these preferred be-*
1661 *haviours?*

1662

1663 5.1 Introduction

1664 The importance of understanding farmers' preferred behavioural intentions under dif-
1665 ferent scenarios, and the underlying predictors of intention, for policy decision-making
1666 have been discussed (see Chapter 1). Furthermore, the use and development of farm

1667 typologies to aid policy decision-makers in effectively designing and implementing pol-
1668 icy changes have also been discussed (see Chapter 2). However, an important research
1669 gap identified in Chapter 2 highlighted that traditional farm typologies, which focus
1670 on physical or economic descriptors, fail to incorporate true farmer behaviour and are
1671 therefore of limited value in supporting policy formulations (Guillem et al., 2012).
1672 Therefore, this chapter extends the analyses conducted in Chapter 4 by exploring
1673 whether farmers who share similar behavioural intention traits also share particular
1674 farm characteristics such as: farm attributes; social demographics; licence character-
1675 istics; and past abstraction behaviour. This method could provide an alternative tool
1676 for policy decision-makers when developing and implementing proposed policy changes
1677 by targeting specific farm types based on behavioural intentions which may naturally
1678 incorporate elements of more traditional typologies.

1679

1680 5.2 Materials and methods

1681 A conceptual approach was used to categorise farmers into three farm types based on
1682 their responses to the behavioural intention scenarios of the questionnaire (sections B
1683 and C). The categories used to define the three farm types related to the Low Water
1684 Usage Options (LWUO) and High Water Usage Options (HWUO) previously discussed
1685 (see section 4.2.1). Individual preferences were determined by rescaling Likert psycho-
1686 metric responses to the Theory of Planned Behaviour (TPB) constructs (Ajzen, 1985),
1687 including past behaviour (see Figure 4.1), from -3 to 3 and calculating a median re-
1688 sponse for each of the 18 behaviours. These median values were then used to calculate
1689 an overall median response for all LWUO and all HWUO for each respondent. Data
1690 were treated as ordinal and therefore median values were used rather than mean val-
1691 ues. Respondents were categorised depending on which median value was greatest. If
1692 both median values were tied then this would suggest that the irrigation farmer had
1693 no preference overall. Therefore, three farm types existed in this conceptual approach:
1694 farmers who preferred HWUO; farmers who preferred LWUO; and farmers who had no
1695 preference.

1696

1697 Farm characteristics including: farm attributes; social demographics; licence char-
1698 acteristics; and past abstractions were analysed to compare similarities and differences

1699 between the three farm types. Pearson's Chi-square tests (X^2) were conducted to test
1700 for statistically significant differences between farm types when examining categorical
1701 variables. This test required two main assumptions: independence of observations; and
1702 that no cells had an expected frequency of less than one and no more than 20 % had
1703 an expected frequency of less than five (Cochran, 1952). Although the first assumption
1704 was satisfied by the fact that only one farmer could belong to a single farm type, the
1705 second assumption occasionally required amalgamating particular categories, if suit-
1706 able, to increase expected frequency. Where this occurred, the amalgamated categories
1707 are stated along with the test results. The null hypothesis for the Pearson's Chi-square
1708 test was that there was no statistically significant difference between farm type and
1709 the categorical variable being tested. Therefore, if the *p-value* was less than the chosen
1710 alpha level, which for this study was .05, then the null hypothesis was rejected suggest-
1711 ing that there was a significant difference between farm types.

1712

1713 In regards to continuous variables, such as size of irrigated areas, Shapiro-Wilk
1714 tests, in addition to normal probability plots, were used to determine whether vari-
1715 ables were normally distributed (see Appendix D). Depending on the results of these
1716 tests, the appropriate parametric or non-parametric test was then used to explore the
1717 null hypothesis that a statistically significant difference existed between the three farm
1718 types if the *p-value* was less than the chosen alpha level, which for these tests was
1719 .05, (one-way analysis of variance (ANOVA) or Kruskal-Wallis respectively). If such
1720 a statistically significant difference existed, then further relevant parametric or non-
1721 parametric tests were conducted to test the same null hypothesis but between two
1722 farm types (independent *t*-test or Mann-Whitney *U* test respectively). However, in
1723 regards to Mann-Whitney *U* tests the difference in median values was reported if the
1724 two distributions were similar in shape; otherwise the difference in mean ranks was
1725 reported if distributions were not similar.

1726

1727 5.3 Results and discussion

1728 Figure 5.1 illustrates the prevalence of the three farm types: those who preferred LWUO
1729 accounting for 23 % of the respondents; those who had no preference accounting for
1730 27 % of the respondents; and those who preferred HWUO accounting for 50 % of the

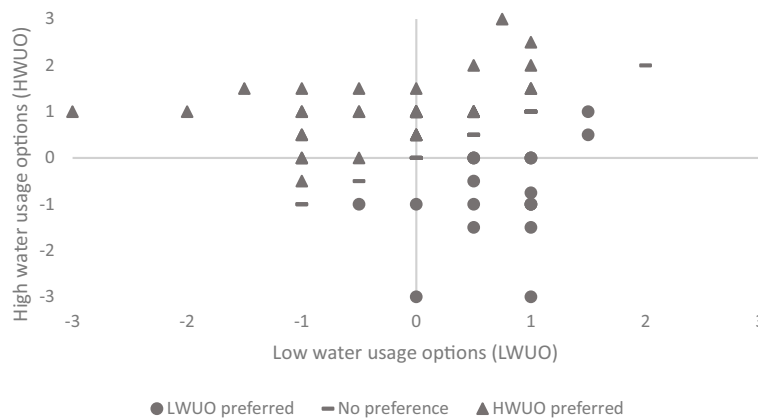


Figure 5.1: A scatter plot indicating three farm types derived from median responses to low water usage options (LWUO) and high water usage options (HWUO), including: respondents who preferred LWUO; respondents who had no preference; and respondents who preferred HWUO ($n = 90$)

1731 respondents. The results indicate that the largest percentage of respondents preferred
 1732 HWUO whilst the smallest percentage of respondents preferred LWUO. Although it
 1733 appears farmers from different farm types are clustered close together this is simply a
 1734 result of the conceptual approach used.

1735

1736 5.3.1 Farm attributes

1737 Table 5.1 provides a comparison of the three farm types' farm attributes. A statistically
 1738 significant difference existed with regards to irrigation application mechanisms used be-
 1739 tween those who preferred LWUO and those who had no preference ($X^2 = 5.033$) as
 1740 $p = < .05$, and between those who preferred LWUO and those who preferred HWUO
 1741 ($X^2 = 9.260$) as $p = < .05$. Thus, the null hypothesis for this farm attribute could be
 1742 rejected as those who preferred LWUO used significantly more rainguns for irrigating
 1743 (80 %) compared to those who had no preference or preferred HWUO (58 % and 53
 1744 % respectively). In addition, a statistically significant difference existed with regards
 1745 to storage availability between those who preferred LWUO and those who preferred
 1746 HWUO ($X^2 = 7.508$) as $p = < .05$. Thus, the null hypothesis for this farm attribute
 1747 could also be rejected as those who preferred LWUO consisted of significantly fewer
 1748 farmers with storage available (24 %) compared with those who preferred HWUO (60
 1749 %). However, in regards to irrigated crops and predominant soil types, no statistically

Table 5.1: Farm types' farm attributes ($n = 90$)

Measure	^a LWUO preferred (%)	No preference (%)	^b HWUO preferred (%)
Irrigated crop			
- Main crop potatoes	41	39	42
- Vegetables	24	41	32
- Other (amalgamated)	34	20	26
Pearson chi-square (X^2)(2 <i>df</i>)			
- LWUO preferred		3.427	1.290
- No preference			1.137
Predominant soil type			
- Sand	32	45	33
- Loam	32	23	42
- Peat	22	17	12
- Mixed	15	15	13
Pearson chi-square (X^2)(3 <i>df</i>)			
- LWUO preferred		1.784	2.611
- No preference			4.705
Irrigation application used			
- Raingun	80	58	53
- Other (amalgamated)	19	42	46
Pearson chi-square (X^2)(1 <i>df</i>)			
- LWUO preferred		5.033*	9.260*
- No preference			.407
Storage available			
- Yes	24	50	60
- No	76	50	40
Pearson chi-square (X^2)(1 <i>df</i>)			
- LWUO preferred		3.268	7.508*
- No preference			.637
Summing errors due to rounding. (<i>df</i>) = Degree of freedom. (*) = $p < .05$			
^a Low water usage options (LWUO)			
^b High water usage options (HWUO)			

1750 significant difference existed between the three farm types, as $p = >.05$, thus the null
1751 hypotheses for these two farm attributes could not be rejected.

1752

1753 Kruskal-Wallis tests showed that statistically significant differences existed between
1754 farm types with regards to irrigated areas ($X^2 = 6.525$) as $p = <.05$, and storage capac-
1755 ities ($X^2 = 9.280$) as $p = <.05$ (see Figure 5.2). In particular, a statistically significant
1756 difference existed with regards to irrigated area between those who preferred LWUO
1757 and those who preferred HWUO ($U = 264$) as $p = <.05$. Thus, the null hypothesis for

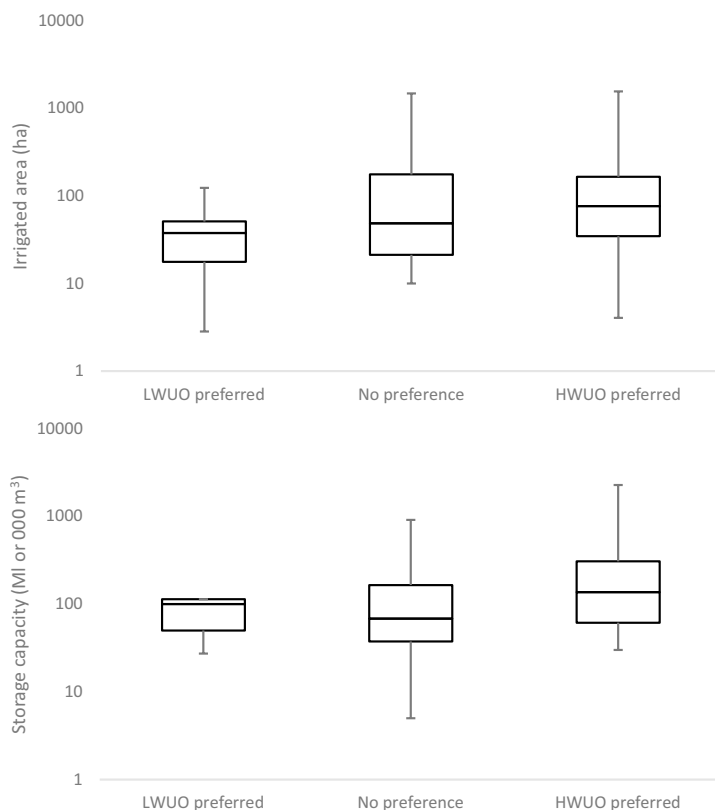


Figure 5.2: Box plots illustrating farm types' irrigated areas and storage capacities. (Top) A box plot illustrating farm types' irrigated areas. (Bottom) A box plot illustrating farm types' storage capacities

1758 this farm attribute could be rejected as those who preferred LWUO irrigated signifi-
 1759 cantly smaller areas (median 38 ha and IQR 33 ha) compared to those who preferred
 1760 HWUO (median 76 ha and IQR 130 ha). Similarly, a statistically significant difference
 1761 existed with regards to storage capacity between those who preferred LWUO and those
 1762 who preferred HWUO ($U = 273$) as $p = < .05$. Thus, the null hypothesis for this farm
 1763 attribute could also be rejected as those who preferred LWUO had significantly smaller
 1764 storage capacities (median 100 Ml and IQR 64 Ml) compared to those who preferred
 1765 HWUO (median 136 Ml and IQR 245 Ml).

1766

1767 5.3.2 Social demographics

1768 Table 5.2 provides a comparison of the three farm types' social demographics. A sta-
 1769 tistically significant difference existed with regards to age between those who preferred

Table 5.2: Farm types' social demographics

Measure	^a LWUO preferred (%)	No preference (%)	^b HWUO preferred (%)
Gender ($n = 86$)			
- Male	100	96	98
- Female	0	4	2
Age ($n = 88$)			
- 18-59	47	63	73
- ≥ 60	53	38	27
Pearson chi-square (X^2)(1 <i>df</i>)			
- LWUO preferred		.985	3.993*
- No preference			.868
^c Education ($n = 79$)			
- ≤ 5	87	65	51
- ≥ 6	13	35	49
Pearson chi-square (X^2)(1 <i>df</i>)			
- LWUO preferred		2.154	5.785*
- No preference			1.173
^d Income ($n = 77$)			
- £0-99,999	29	23	22
- \geq £100,000	71	77	78
Pearson chi-square (X^2)(1 <i>df</i>)			
- LWUO preferred		.156	.253
- No preference			.005
Summing errors due to rounding. (<i>df</i>) = Degree of freedom. (*) = $p < .05$			
^a Low water usage options (LWUO)			
^b High water usage options (HWUO)			
^c National Qualification Framework (NQF) levels for England, Wales and Northern Ireland			
^d Total gross annual farm business income			

1770 LWUO and those who preferred HWUO ($X^2 = 3.993$) as $p = < .05$. Thus, the null hy-
1771 pothesis for this social demographic could be rejected as those who preferred LWUO
1772 consisted of significantly older farmers (53 % were aged 60 or over) compared with those
1773 who preferred HWUO (only 27 % were aged 60 or over). In addition, a statistically
1774 significant difference existed with regards to education between those who preferred
1775 LWUO and those who preferred HWUO ($X^2 = 5.785$) as $p = < .05$. Thus, the null
1776 hypothesis for this social demographic could also be rejected as those who preferred
1777 LWUO consisted of significantly fewer farmers who had achieved a National Qualifi-
1778 cation Level (NQF) of 6 or higher (only 13 % had a university degree or equivalent)
1779 compared with those who preferred HWUO (49 %). However, in regards to gender, the

1780 assumptions for conducting Pearson's Chi-square test could not be satisfied as very few
1781 female farmers existed. Furthermore, in regards to income, no statistically significant
1782 difference existed between the three farm types, as $p = >.05$, thus the null hypothesis
1783 for this social demographic could not be rejected.

1784

1785 5.3.3 Licence characteristics

1786 Table 5.3 provides a comparison of the three farm types' licence characteristics. A sta-
1787 tistically significant difference existed with regards to licence type between those who
1788 preferred LWUO and those who preferred HWUO ($X^2 = 7.508$) as $p = <.05$. Thus, the
1789 null hypothesis for this licence characteristic could be rejected as those who preferred
1790 LWUO consisted of significantly fewer storage, or direct and storage, licence users (76
1791 % owned direct licences only) compared with those who preferred HWUO (only 40 %).
1792 In addition, a statistically significant difference existed with regards to time limited
1793 licences between those who preferred LWUO and those who preferred HWUO ($X^2 =$
1794 7.698) as $p = <.05$. Thus, the null hypothesis for this licence characteristic could also
1795 be rejected as those who preferred LWUO consisted of significantly fewer farmers who
1796 owned time limited licences (52 %) compared with those who preferred HWUO (84
1797 %). However, in regards to licence use during the last 10 years, the assumptions for
1798 conducting Pearson's Chi-square test could not be satisfied as very few non-active users
1799 existed. Furthermore, in regards to licence source, no statistically significant difference
1800 existed between the three farm types, as $p = >.05$, thus the null hypothesis for this
1801 licence characteristic could not be rejected.

1802

1803 Kruskal-Wallis tests showed that statistically significant differences existed between
1804 farm types with regards to annual licence limits ($X^2 = 11.078$) as $p = <.05$, and daily
1805 licence limits ($X^2 = 15.172$) as $p = <.05$ (see Figure 5.3). In particular, a statistically
1806 significant difference existed with regards to annual licence limit between those who
1807 preferred LWUO and those who had no preference ($U = 139$) as $p = <.05$, and between
1808 those who preferred LWUO and those who preferred HWUO ($U = 240$) as $p = <.05$.
1809 Thus, the null hypothesis for this licence characteristic could be rejected as those who
1810 preferred LWUO had significantly smaller annual licence limits (median 18 Ml and IQR
1811 26 Ml) compared to those who had no preference (median 49 Ml and IQR 68 Ml), and

Table 5.3: Farm types' licence characteristics ($n = 90$)

Measure	^a LWUO preferred (%)	No preference (%)	^b HWUO preferred (%)
Licence use (last 10 years)			
- Yes	100	92	100
- No	0	8	0
Licence type			
- Direct	76	50	40
- Other (amalgamated)	24	50	60
Pearson chi-square (X^2)(1 <i>df</i>)			
- LWUO preferred		3.268	7.508*
- No preference			.637
Licence source			
- Surface water	81	75	78
- Groundwater	19	25	22
Pearson chi-square (X^2)(1 <i>df</i>)			
- LWUO preferred		.230	.086
- No preference			.068
Time limited			
- Yes	52	79	84
- No	48	21	16
Pearson chi-square (X^2)(1 <i>df</i>)			
- LWUO preferred		3.616	7.698*
- No preference			.303
Summing errors due to rounding. (<i>df</i>) = Degree of freedom. (*) = $p < .05$			
^a Low water usage options (LWUO)			
^b High water usage options (HWUO)			

1812 compared to those who preferred HWUO (median 55 Ml and IQR 74 ml). Similarly,
1813 a statistically significant difference existed with regards to daily licence limit between
1814 those who preferred LWUO and those who preferred HWUO ($U = 193$) as $p = < .05$, and
1815 between those who had no preference and those who preferred HWUO ($U = 382$) as $p =$
1816 $< .05$. Thus, the null hypothesis for this licence characteristic could also be rejected as
1817 those who preferred LWUO, and those who had no preference, had significantly smaller
1818 daily licence limits (median 1 Ml and IQR < 1 ml, and median 1 Ml and IQR 2 Ml
1819 respectively) compared to those who preferred HWUO (median 2 Ml and IQR 2 ml).
1820 Therefore, although those who preferred LWUO had a significantly smaller annual and
1821 daily licence limit than those who preferred HWUO, those who had no preference had
1822 a significantly larger annual licence limit than those who preferred LWUO but had a
1823 significantly smaller daily licence limit than those who preferred HWUO.

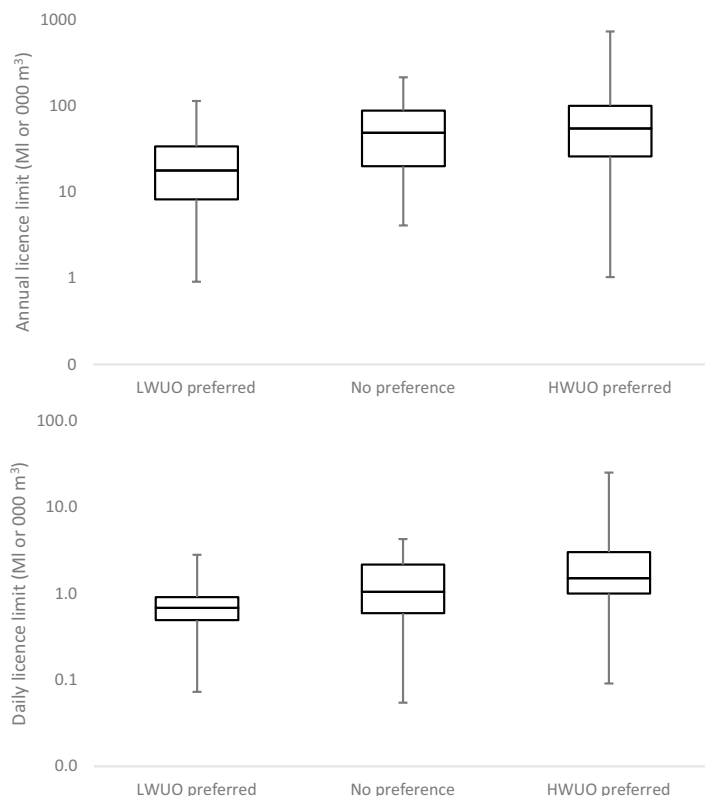


Figure 5.3: Box plots illustrating farm types' annual and daily licence limits. (Top) A box plot illustrating farm types' annual licence limit. (Bottom) A box plot illustrating farm types' daily licence limit

1824 5.3.4 Past abstractions (2008 to 2012)

1825 Kruskal-Wallis tests showed that statistically significant differences existed between
 1826 farm types with regards to average annual volume abstracted ($X^2= 10.294$) as $p=$
 1827 $<.05$, and average in-season (i.e. April to October) volume abstracted ($X^2= 7.072$) as
 1828 $p= <.05$ (see Figure 5.4). In particular, a statistically significant difference existed with
 1829 regards to average annual volume abstracted between those who preferred LWUO and
 1830 those who had no preference ($U= 127$) as $p= <.05$, and between those who preferred
 1831 LWUO and those who preferred HWUO ($U= 269$) as $p= <.05$. Thus, the null hypoth-
 1832 esis for this past abstraction characteristic could be rejected as those who preferred
 1833 LWUO abstracted significantly less on average annually (median 3 MI and IQR 7 MI)
 1834 compared to those who had no preference (median 15 MI and IQR 31 MI), and to those
 1835 who preferred HWUO (median 12 MI and IQR 20 MI). Similarly, a statistically signif-
 1836 icant difference existed with regards to average in-season volume abstracted between

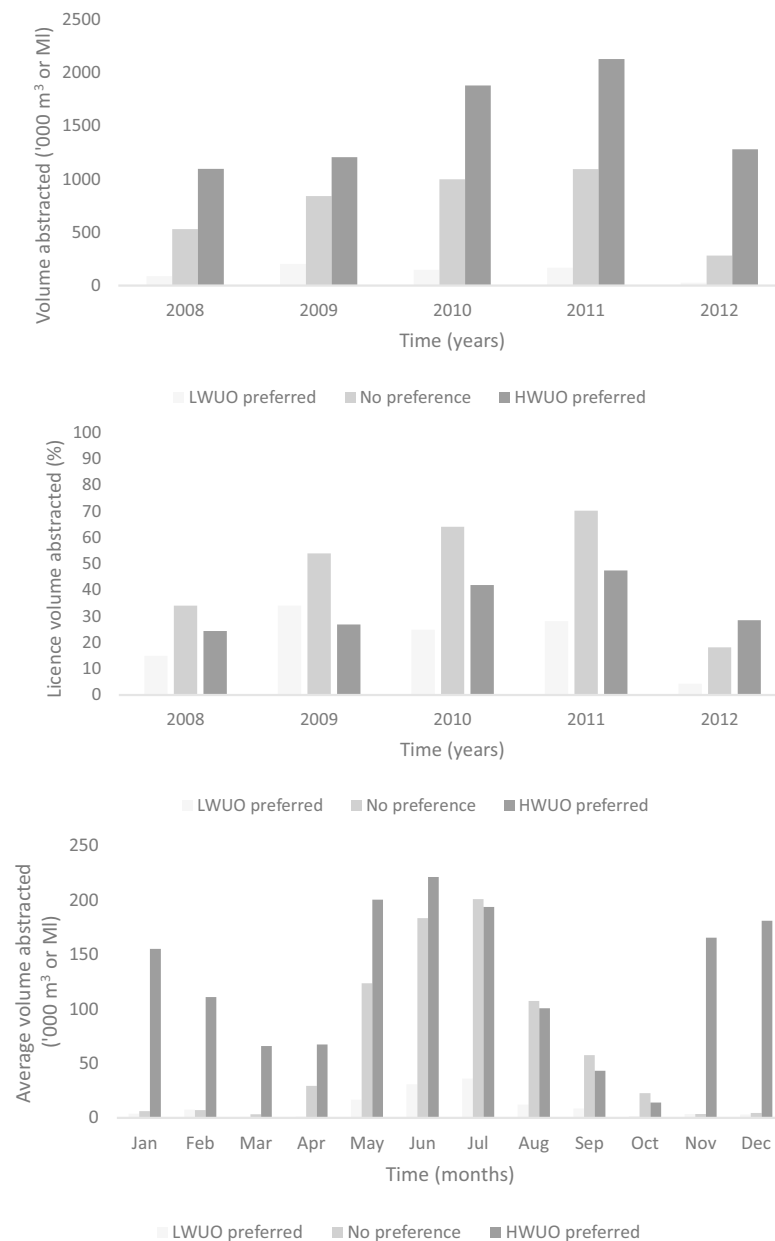


Figure 5.4: Bar charts illustrating farm types and total respondents annual, licensed, and average monthly abstractions (2008 to 2012). (Top) A bar chart illustrating annual abstractions by farm type. (Middle) A bar chart illustrating the percentage of licence volume abstracted by farm type and total respondents. (Bottom) A bar chart illustrating average monthly abstractions by farm type

1837 those who preferred LWUO and those who had no preference ($U = 141$) as $p = < .05$.
 1838 Thus, the null hypothesis for this past abstraction characteristic could also be rejected
 1839 as those who preferred LWUO abstracted significantly less on average in-season (me-

1840 dian <1 MI and IQR 1 MI) compared to those who had no preference (median 2 MI
1841 and IQR 5 MI).

1842

1843 Interestingly, despite those who preferred LWUO and those who had no preference
1844 abstracting only 20 % and 5 % out-of-season compared to 53 % for those who preferred
1845 HWUO, no statistically significant differences existed between farm types with regards
1846 to average out-of-season volume abstracted ($X^2= 5.513$) as $p= >.05$. Nonetheless, this
1847 higher percentage of water abstracted out-of-season to refill farm storage reservoirs does
1848 support earlier findings as those who preferred HWUO consisted of significantly more
1849 farmers with storage available, and significantly more farmers who owned storage, and
1850 direct and storage, licences compared to those who preferred LWUO.

1851

1852 5.4 Summary

1853 Overall, this study was driven by a growing number of researchers and policy decision-
1854 makers realising the necessity of incorporating farmers behavioural intentions into more
1855 traditional farm typologies in order to aid the design, implementation, and overall suc-
1856 cess of policy changes. In particular, this chapter aimed to inform policy decision-
1857 makers on how the proposed water policy reform options in England might be received
1858 by particular groups, or farm types, on the ground. In doing so, this chapter addressed
1859 the second of three research questions set out in the introduction to this thesis, and
1860 highlighted again at the start of this chapter. That was, if farmers shared similar
1861 behavioural intentions under different scenarios how could traditional farm typologies
1862 incorporate these preferred behaviours. This chapter has identified three farm types
1863 based on farmers' overall preferred behaviours analysed in the previous chapter: those
1864 who preferred LWUO; those who had no preference; and those who preferred HWUO.
1865 Furthermore, a series of statistical tests were conducted in order to examine similari-
1866 ties and differences between the three farm types in regards to: farm attributes; social
1867 demographics; licence characteristics, and past abstractions.

1868

1869 Table 5.4 provides a summary of the significant differences in farm characteristics
1870 identified between the three farm types, in addition to each farm types' preferred be-
1871 havioural intention under each scenario. Despite farm types being determined based on

Table 5.4: A conceptual farm typology based on respondents' overall preferred behavioural intentions across all four scenarios ($n = 90$)

Measure	^a LWUO preferred	No preference	^b HWUO preferred
Farm characteristics (Significant differences)			
- Irrigation application	Mainly rainguns	Mixed	Mixed
- Storage available	Minority		Majority
- Irrigated area	Small		Large
- Storage capacity	Small		Large
- Age (18-59)	Minority		Majority
- ^c Education (≥ 6)	Minority		Majority
- Licence type	Mainly direct		^d Mainly storage
- Time limited licences	Mixed		Mainly time limited
- Annual licence limit	Small	Large	Large
- Daily licence limit	Small	Small	Large
- Annual abstractions	Small	Large	Large
- In-season abstractions	Small	Large	
Preferred behavioural intentions			
- Strategic (long-term)/ water shortage		Increase application efficiency (LWUO)	
- Strategic (long-term)/ water surplus	Change nothing (LWUO)	Sell surplus water for the duration of the growing season (LWUO)	Grow the same crops but over a larger area (HWUO)
- In-season (short-term)/ water shortage		Only use their maximum abstraction licence to irrigate their most valuable crops (LWUO)	
- In-season (short-term)/ water surplus	Just use their abstraction licence to meet crop water requirements and leave the remainder of their licence unused (LWUO)	Sell surplus water to maximise profits (LWUO)	Abstract surplus water for storage (HWUO)

^a Low water usage options (LWUO)

^b High water usage options (HWUO)

^cNational Qualification Framework (NQF) levels for England, Wales and Northern Ireland

^d Includes both storage, and direct and storage licences

See Tables 5.1, 5.2, and 5.3, and Figures 5.2, 5.3, and 5.4 for actual values

See Appendix D for farm types' ranked preferred behavioural intentions

1872 respondents' overall behavioural intentions across all four scenarios, similarities and dif-
 1873 ferences in preferred behavioural intentions existed within each scenario (see Appendix
 1874 D).

1875 Those who preferred LWUO were generally representative of farmers who irrigated
1876 smaller areas, had fewer resources available, and were older and had a lower level of
1877 education compared to farmers who preferred HWUO. In particular, fewer resources
1878 refer to less available storage and capacity (and therefore fewer storage related licences),
1879 smaller annual and daily licence limits, and therefore smaller average annual abstrac-
1880 tions compared to those who preferred HWUO. In addition, those who preferred LWUO
1881 also relied predominantly on rainguns for irrigation rather than a more mixed approach
1882 used by those who had no preference, and those who preferred HWUO.

1883

1884 Conversely, those who preferred HWUO were generally representative of farmers
1885 who irrigated larger areas, had greater resources available, and were younger and had
1886 a higher level of education compared to farmers who preferred LWUO. In particular,
1887 greater resources refer to more available storage and capacity (and therefore more stor-
1888 age related licences), larger annual and daily licence limits, and therefore larger average
1889 annual abstractions compared to those who preferred LWUO. In addition, those who
1890 preferred HWUO also owned mainly time limited licences compared to a more mixed
1891 ratio by those who preferred LWUO. The most likely explanation for this relates to
1892 the EA slowly phasing out none-time limited licences, in order for greater control over
1893 protecting the environment, as new licences are issued. Therefore, licences issued to
1894 younger farmers are perhaps more likely to be time limited compared to older farmers
1895 who have retained their none-time limited licences.

1896

1897 Lastly, those who had no preference were unsurprisingly more difficult to define
1898 as they shared significant differences with both other farm types, although more in
1899 common with those who preferred HWUO. In particular, they relied on a more mixed
1900 approach to irrigation, similar to those who preferred HWUO. In addition, they had a
1901 greater resource with regards to annual licence limit and therefore had similarly large
1902 annual abstractions in comparison to those who preferred HWUO. However, they had a
1903 smaller daily licence limit similar to those who preferred LWUO. In addition, those who
1904 had no preference abstracted more in-season compared to those who preferred LWUO.

1905

1906 In regards to similarities, all farm types irrigated similar proportions of main crop
1907 potatoes and vegetables, on sand and loam which were the predominant soil types.

1908 The majority of farmers were all male, and total gross annual farm business income
1909 was over £100,000. Furthermore, the majority had all used their abstraction licences
1910 within the last 10 years, and were predominantly all surface water licences.

1911

1912 Interestingly, all farm types shared similar LWUO preferred behavioural intentions
1913 under strategic (long-term) and in-season (short-term) water shortage scenarios. That
1914 is, under the strategic (long-term)/water shortage scenario they all preferred to increase
1915 application efficiency, and under the in-season (short-term)/water shortage scenario
1916 they all preferred to only use their maximum abstraction licence to irrigate their most
1917 valuable crops. This is not particularly surprising for those who preferred LWUO, or
1918 perhaps for those who had no preference. However, with regards to those who preferred
1919 HWUO, one explanation could simply be the HWUO available were not regarded as
1920 feasible or practical in comparison to the preferred LWUO. For example, increasing
1921 storage capacity for many of those who preferred HWUO may not be possible if many
1922 already have storage available. Furthermore, many of the farmers would not have ex-
1923 perience with trading so buying more water for the duration of the growing season,
1924 or buying more water to meet crop water requirements, may have seemed very diffi-
1925 cult compared to increasing application efficiency. The same is true with applying for
1926 a larger abstraction licence. However, these HWUO were ranked second, third and
1927 fourth during a strategic (long-term)/ water shortage, and second during an in-season
1928 (short-term)/water shortage by those who preferred HWUO, and by those who had no
1929 preference (see Appendix D).

1930

1931 During strategic (long-term) and in-season (short-term) water surplus scenarios
1932 each farm type had different preferred behavioural intentions. For example, under a
1933 strategic (long-term)/water surplus scenario: those who preferred LWUO preferred to
1934 change nothing; those who had no preference preferred to sell surplus water for the
1935 duration of the growing season; and those who preferred HWUO preferred to grow the
1936 same crops but over a larger area. Similarly, under an in-season (short-term)/water
1937 surplus scenario: those who preferred LWUO preferred to just use their abstraction li-
1938 cence to meet crop water requirements and leave the remainder of their licence unused;
1939 those who had no preference preferred to sell surplus water to maximise profits; and
1940 those who preferred HWUO preferred to abstract surplus water for storage. As perhaps

1941 would be expected, those who preferred LWUO favoured LWUO, whilst those who pre-
1942 ferred HWUO favoured HWUO. More interestingly, those who had no preference also
1943 favoured LWUO, although rather than doing nothing with the surplus water they pre-
1944 ferred to sell. This would suggest that those who had no preference had slightly more
1945 in common with those who preferred HWUO in regards to farm characteristics but
1946 more in common with those who preferred LWUO in regards to preferred behavioural
1947 intentions within each scenario.

1948

1949 This study has shown how farmers' preferred behavioural intentions can be incor-
1950 porated into more traditional farm typologies, which are typically based on physical or
1951 economic descriptors only, and offer valuable support with regards to policy formula-
1952 tion. Therefore, this study has addressed the second research gap identified in Chapter
1953 2, and presented again at the start of this chapter. Furthermore, the typology indi-
1954 cate that the most likely farm types to engage in water licence trading are those who
1955 have no preference and those who prefer HWUO. Therefore, the typology could also
1956 be used to determine which catchments would likely benefit most from the proposed
1957 enhanced water allocation system based on the proportion of farmers who have simi-
1958 lar farm characteristics to those who had no preference and those who preferred HWUO.

1959

1960 Chapter 6

1961 Modelling system level patterns 1962 of abstraction behaviour

1963 This chapter offers one of the first insights into: 1) how the proposed water allocation
1964 systems, both in basic and enhanced catchments, are likely to change water availability
1965 for farmers; and 2) the importance of understanding abstraction behaviour in determin-
1966 ing the overall success of these systems. The proposed water allocation systems, which
1967 are to be implemented by the early 2020s, aim to improve efficiency whilst balancing
1968 the needs of licence users and the environment. However, climate change is expected
1969 to exacerbate this challenge. Therefore, this chapter presents an empirically based
1970 Agent-Based Model (ABM) to understand what system level patterns of abstraction
1971 behaviour emerge from individual farmer decisions under different policy (the current
1972 and proposed water allocation systems) and climate scenarios (during a selected dry
1973 and wet year). Therefore, this chapter addresses the final of the three research ques-
1974 tions presented in the introductory chapter:

1975

1976 **Research question 3:** *What potential system level patterns of abstraction be-*
1977 *haviour emerge from individual farm scale decisions under different policy and climate*
1978 *scenarios?*

1979

1980 6.1 Introduction

1981 Major water allocation system reforms have been proposed in England to address the
1982 climate and demand change pressures which the current system is failing to address

(see Chapter 1). A critical review of the literature highlighted the importance of policy decision-makers, involved in the design and implementation of the proposed water allocation systems, to understand farmers' behaviour at the farm and system level yet very little is understood in either of these areas (see Chapter 2). The Great Ouse catchment was selected as the study area due to the importance of available freshwater resources for agricultural spray irrigation despite it being one of the driest and most water-stressed regions in the UK (see Chapter 3). Chapters 4 and 5 have presented empirical data which identified farmers' preferred behavioural intentions under different scenarios of water shortage and surplus with regards to the proposed water allocation systems, and used this data to develop a farm typology based on preferred behavioural intentions rather than traditional farm characteristics. This chapter presents the development and scenario simulation results of an ABM, which utilises the empirical data and typology discussed in the previous chapters, to understand what system level patterns of abstraction behaviour emerge from individual farm decisions under different policy and climate scenarios.

1998

1999 **6.2 Materials and methods**

2000 The methodological procedure used in this study can be categorised into three stages:
2001 1) model description; 2) policy and climate scenario selection; and 3) statistical analysis.

2002

2003 **6.2.1 Model description**

2004 The model, developed using NetLogo 5.1.0 (Wilensky, 1999), is described following
2005 the standardised Overview, Design concepts, and Details (ODD) protocol designed by
2006 Grimm et al. (2006), and updated by (Grimm et al., 2010) (see section 2.3.2).

2007

2008 **Purpose**

2009 The purpose of this model was to understand the system level patterns of abstraction
2010 behaviour which emerge based on farmers' preferred behavioural intentions under dif-
2011 ferent policy and climate scenarios (see section 6.2.2). The aims of the proposed water
2012 allocation systems are to improve efficiency whilst balancing the needs of licence users

2013 and the environment. Therefore, as this study was only simulating farmers, and the
2014 environment was considered to be better protected under the proposed water allocation
2015 systems due to the introduction of low flow conditions, four model indicators were used.
2016 The four indicators were: monthly licensed volume abstracted; surplus water available;
2017 shortage water required; and the percentage of farm Irrigation Water Need (IWN) sat-
2018 isfied (i.e. of main crop potatoes and vegetables) (see Appendix E). Monthly licensed
2019 volume abstracted was used to understand the overall system level pattern of farmers
2020 abstractions. Surplus water available, and shortage water required, were considered im-
2021 portant to highlight the potential for trading, thus improve efficiency, between farmers
2022 and other sectors at different times of the year. Finally, the percentage of farm IWN
2023 satisfied was used to understand the effectiveness of the different policy scenarios.

2024

2025 **Entities, state variables, and scales**

2026 There are three types of entities in this model: agents (i.e. farmers); grid cells (i.e.
2027 the landscape); and the observer (i.e. the global environment) all of which have their
2028 own unique set of state variables (see Table 6.1). Farmers' state variables included two
2029 types: those which remain the same during each model setup (i.e. identification and
2030 location within the projected landscape of the study area); and those which depend on
2031 the farmers' farm type, which is randomly assigned at the start of each model setup.
2032 However, the proportion of farmers within each farm type remain the same during each
2033 model run and is presented in Section 5.3: 23 % of farmers are those who preferred Low
2034 Water Usage Options (LWUO); 27 % are farmers who had no preference; and 50 % are
2035 farmers who preferred High Water Usage Options (HWUO). Irrigated areas, storage
2036 capacities, and annual licence limits are calculated from distributions, using median,
2037 25th and 75th percentile values, which are unique to each farm type. Therefore, farmers
2038 within the same farm type can have different individual values for each of these state
2039 variables at the start of each model setup. However, farmers' other state variables
2040 are single values depending on their farm type including: the percentage of irrigated
2041 area covered by main crop potatoes and vegetables; the percentage of farmers with
2042 storage available; the percentage of farmers with storage only licences; and preferred
2043 behavioural strategies. In addition, the percentage of crop cover excludes any other
2044 crops besides main crop potatoes and vegetables and therefore values differ between

Table 6.1: Entities, state variables, and scales

Variable	Unit	Value (value derived from)
Farmers' state variables		
- Identification	n	1-836
- Location	^a Coordinates	Figure 3.1
- Farm type	^b n	Section 5.3
- Irrigated area	m^2	Figure 5.2
- Main crop potato cover	%	Table 5.1
- Vegetable cover	%	Table 5.1
- Storage availability	%	Table 5.1
- Storage only licence	%	Table 5.3
- Storage capacity	m^3	Figure 5.2
- Annual licence limit	m^3	Figure 5.3
- Preferred behavioural strategies		Table 5.4
Landscape state variables		
- ^c AP location	^a Coordinates	Figure 3.1
- ^c AP name		Figure 3.1
- ^c AP low flow condition (Q95)	m^3	Table 3.1
- ^c AP high flow condition (Q30)	m^3	Table 3.1
Global environment state variables		
- ^d IWN of main crop potatoes	mm	Appendix E
- ^d IWN of vegetables (i.e. carrots)	mm	Appendix E
- ^c AP river discharges (dry year)	m^3	Appendix E
- ^c AP river discharges (wet year)	m^3	Appendix E

^a British National Grid

^b Farm types are randomly assigned but proportional to the empirical data in the farm typology

^c Assessment Point

^d Irrigation water need

2045 the model and the farm typology. Furthermore, carrots were selected to represent veg-
 2046 etables due to the availability of data with regards to calculating IWN (see Appendix E).

2047

2048 In regards to farmers' preferred behavioural strategies, only in-season (short-term)
 2049 water shortage and surplus strategies were simulated. The added complexity and large
 2050 assumptions involved with incorporating the strategic (long-term) water shortage and
 2051 surplus strategies, particularly with regards to how farmers would increase application
 2052 efficiency or grow the same crops but over a larger area, were considered to over com-
 2053 plicate the model and not provide the most efficient and effective way of addressing
 2054 the research question. The strategic (long-term) water shortage and surplus strategies
 2055 were therefore omitted in favour of a simpler but nonetheless more reliable model.

2056 During an in-season (short-term)/water shortage, all three farm types preferred to
2057 only use their maximum abstraction licence to irrigate their most valuable crops. In the
2058 model, this was defined by which of the two crops represented the largest percentage
2059 of irrigated area, as this was considered likely to be the crop which generated the most
2060 income. Main crop potatoes represented 63 % and 57 % of irrigated areas for those who
2061 preferred LWUO and HWUO respectively, whilst vegetables (i.e. carrots) represented
2062 51 % of irrigated area for those who had no preference. During an in-season (short-
2063 term)/water surplus, those who preferred LWUO preferred to just use their abstraction
2064 licence to meet crop water requirements and leave the remainder of their licence unused.
2065 Those who had no preference preferred to sell surplus water to maximise profits, whilst
2066 those who preferred HWUO preferred to abstract surplus water for storage. However,
2067 this final behavioural strategy was designed to automatically occur for all farm types
2068 within the model who had storage available. Therefore, those who preferred HWUO
2069 performed their second highest ranked behavioural strategy which was to sell surplus
2070 water to maximise profits. In addition, although two of the farm types behavioural
2071 strategies involved selling surplus water to maximise profits, none of the farm types
2072 behavioural strategies involved buying more water to meet crop water requirements,
2073 and therefore no trading occurred. Nonetheless, the quantity of surplus water available
2074 and shortfall in water required were used as indicators to evaluate the potential for
2075 trading, and thus a measure of potential increase in efficiency.

2076

2077 Landscape state variables included: Assessment Point (AP) location, which was a
2078 projected shapefile of the study area; AP name; AP low flow conditions (Q95); and AP
2079 high flow conditions (Q30). However, low flow conditions were dependent on policy
2080 scenario, and were therefore not included when simulating the current water allocation
2081 system. Furthermore, HoF data was unavailable for individual farmers and was there-
2082 fore not included in the model. State variables associated with the global environment,
2083 which drive the behaviour and dynamics of the other entities, included IWN of main
2084 crop potatoes and vegetables (i.e. carrots), and river discharge of each AP area during
2085 both the dry and wet year climate scenarios.

2086

2087 In regards to scales, spatially each grid cell represents 1 km² with the extent of the
2088 model reflecting the projected landscape of the study area (8,596 km²) (see Figure 6.1).

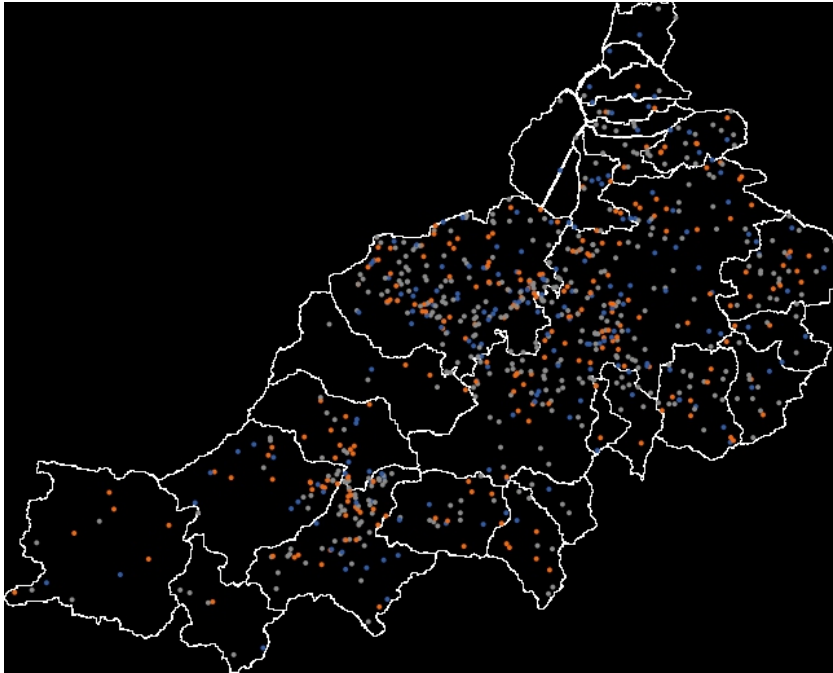


Figure 6.1: Model setup in Netlogo 5.1.0 highlighting each AP location within the Great Ouse catchment and the location of farmers by farm type: those who preferred low water usage options (LWUO) (blue); those who had no preference (orange); and those who preferred high water usage options (HWUO) (grey)

2089 Temporally, each time step represents one month, as this corresponded with available
2090 precipitation and temperature data used to derive IWN (see Appendix E). Simulations
2091 ran for one year, for both dry and wet years, to indicate in-season (short-term) water
2092 shortage and surplus abstraction behaviour during extreme climate scenarios.

2093

2094 **Process overview and scheduling**

2095 The main order of procedures simulated asynchronously each month in the model are
2096 presented in pseudo-code in Figure 6.2. First, the observer runs the monthly param-
2097 eters including IWN for main crop potatoes and vegetables (i.e. carrots) and river
2098 discharge of each AP area. Farmers then calculate how much water they require to
2099 satisfy IWN for both crops based on the size of their irrigated area, the percentage
2100 of irrigated area each crop represents, and the IWN of both crops. Grid cells, within
2101 the projected landscape, then calculate how much water they have available, based on
2102 their AP name and the river discharge of that AP area, in addition to low and high

```
to go
  run-monthly-parameters (observer procedure)
  calculate-farm-water-requirements (farmer procedure)
  calculate-ap-water-availability (grid and farmer procedure)
  calculate-farm-water-availability (farmer procedure)
  calculate-farm-water-abstractions (farmer procedure)
  perform-preferred-behavioural-strategies (farmer procedure)
end
```

Figure 6.2: Pseudo code highlighting the main procedures simulated each month in the model

2103 flow conditions derived from river flow data from 1973 to 2012 (see Chapter 3). Within
2104 this same procedure, farmers determine how much potential water there is within their
2105 AP area as it can vary depending on the water allocation system being simulated, in
2106 addition to the climate scenario.

2107

2108 After the previous procedures have been performed, farmers calculate how much
2109 water they have available which is determined based on licensed water availability and
2110 storage water availability, both of which are dependent on the water allocation system
2111 being simulated, in addition to variability of particular variables during initialisation.
2112 However, as a farmer may have one or both of these sources of water available during a
2113 period when there is no IWN or when there is an IWN the remaining farmer procedures
2114 highlighted in the pseudo-code are more simply illustrated in Figure 6.3.

2115

2116 If a farmer has no licensed water available and there is no IWN then they assess
2117 whether they need to replenish their storage reservoirs. However, regardless of the
2118 result, they have no water available and they are therefore unable to refill their stor-
2119 age reservoirs and as they have not used any of their annual licence limit it remains
2120 the same for the procedures next month. Similarly, if a farmer has no licensed water
2121 available but there is an IWN then they assess whether they have enough storage water
2122 available to satisfy IWN. If they do not have enough to cover IWN then they perform
2123 their in-season (short-term)/water shortage behavioural strategy (dependent on farm
2124 type) and update their storage water availability for next month accordingly. If, on the
2125 other hand, they do have enough to cover IWN they simply update their storage water

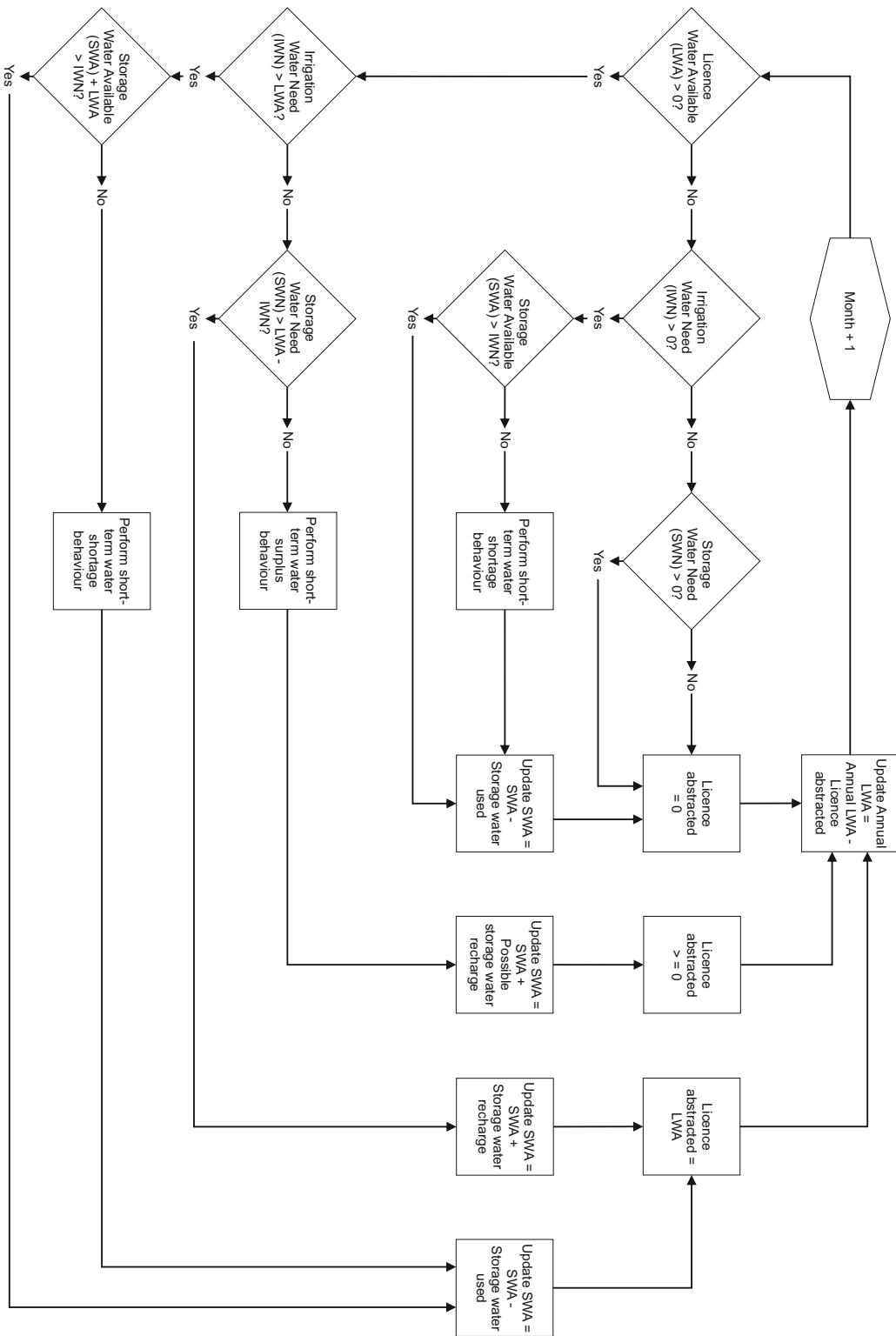


Figure 6.3: Flow diagram illustrating monthly farmer decisions

2126 availability for next month. Either way, no licensed water has been used and therefore
2127 their annual licence limit remains the same for next month.

2128

2129 Alternatively, if a farmer has licensed water available and it is greater than total
2130 IWN then they assess whether they need to replenish their storage reservoirs with
2131 the surplus available. If the surplus available is more than they require to replenish
2132 their storage reservoirs then they perform their in-season (short-term)/water surplus
2133 behavioural strategy (dependent on farm type) and update their storage water avail-
2134 ability and annual licence limit (if they used their licence limit) for next month accord-
2135 ingly. If, on the other hand, the surplus available is less than required to replenish their
2136 storage reservoirs then they simply update their storage water availability and annual
2137 licence limit (as they used all their licence for that month) for next month accord-
2138 ingly. However, if a farmer has licensed water available and it is less than total IWN then
2139 they assess whether they have enough licensed water available plus storage water avail-
2140 able to satisfy IWN. If they do not have enough to cover IWN then they perform their
2141 in-season (short-term)/water shortage behavioural strategy (dependent on farm type)
2142 and update their storage water availability and annual licence limit (as they used all
2143 their storage and all their licence for that month) for next month accordingly. If, on the
2144 other hand, they do have enough to cover IWN then they simply update their storage
2145 water availability and annual licence limit for next month accordingly.

2146

2147 **Design concepts**

2148 Design concept elements of the ODD protocol have been grouped together to avoid
2149 repetition. The general concept underlying this model is the farm typology presented
2150 in Chapter 5. In particular the main concept is that individual behavioural strategies
2151 and attributes at the farm scale, based on empirical data, will lead to different over-
2152 all patterns of behaviour at the system level under different water allocation systems
2153 (policy scenarios) during dry and wet years (climate scenarios). Furthermore, the indi-
2154 cators which are used to evaluate the different water allocation systems at the system
2155 level are model results which emerge from individual farmers performing behavioural
2156 strategies at the farm scale. In addition, the main objective of farmers is to satisfy
2157 their monthly IWN based on licensed and storage water availability.

2158 In addition, depending on the water allocation system being simulated and the type
2159 of licence a farmer possesses, they implicitly assume that they will require some of that
2160 licence later in the year and therefore divide their annual licence by the number of
2161 months they are permitted to abstract from the environment. This highlights a design
2162 in the model which could be improved by incorporating a more sophisticated method
2163 for farmers to assume how much water they will require each month. Nonetheless,
2164 in the model presented, under the current system those with direct licences can only
2165 abstract between April and October (i.e. in-season), those with storage only licences
2166 can only abstract between November and March (i.e. out-of-season), whilst those with
2167 direct and storage licences can abstract all year. However, this self-imposed limit does
2168 not accumulate from one month to the next if unused due to water availability.

2169

2170 Furthermore, all of the global environment state variables, and the landscape state
2171 variables on which the farmer is located, are sensed by the individual farmers as these
2172 drive their behaviours and dynamics. In regards to the landscape state variables this
2173 was considered a reasonable assumption as these simply include location and low and
2174 high flow conditions which they would know as part of their licence. However, in re-
2175 gards to global environment state variables, it is more likely that farmers would know
2176 the approximate IWN of their crops and river discharge in their AP area, rather than
2177 the exact values. Therefore, this highlights another design in the model which could
2178 be improved by incorporating a more realistic method for farmers to estimate IWN.
2179 Although available water resources within each AP area fluctuate from month to month
2180 this only affects farmers under the enhanced water allocation system, where licences
2181 increase and decrease proportionally with water availability in the AP area and there-
2182 fore no direct or indirect interaction between farmers occurs. Finally, stochasticity of
2183 particular variables associated with farmers occur during model initialisation including:
2184 farm type; irrigated areas; storage capacities; and annual licence limits. However, all
2185 of these are based on values presented in the farm typology (see Chapter 5).

2186

2187 **Initialisation**

2188 At the start of each simulation, each of the 836 farmers are located on their actual ge-
2189 ographical location within the projected landscape (i.e. the study area) with a unique

Table 6.2: Summary of scenarios simulated

Scenario	Climate	
	Dry year (1983)	Wet year (1988)
Policy	Current	Current
	Basic	Basic
	Enhanced	Enhanced

2190 identification number. Furthermore, although the proportion of farmers within each
 2191 farm type are kept the same as the empirical data, their farm types are randomly as-
 2192 signed. This was due to reasons of confidentiality and the EA only providing farmers'
 2193 locations, except for the 90 respondents which received additional data, and therefore
 2194 no assumptions could be made regarding which farm type the majority of farmers were
 2195 more likely to belong to. Farmers' farm type was used to determine their individual
 2196 state variables as previously discussed.

2197

2198 **Input data and submodels**

2199 All input data was derived from the previous chapters and the actual model input data,
 2200 along with the submodels highlighted in Figure 6.2, are presented in Appendix E.

2201

2202 **6.2.2 Policy and climate scenarios**

2203 Three policy scenarios under two climate scenarios were simulated in this study (see
 2204 Table 6.2). The three policy scenarios related to: the current water allocation system;
 2205 the proposed basic water allocation system; and the proposed enhanced water alloca-
 2206 tion system (see Section 1.1.4). The two climate scenarios related to a dry year and a
 2207 wet year based on total annual IWN of main crop potatoes and vegetables (i.e. carrots)
 2208 combined. Appendix E provides a description of how monthly and annual IWN were
 2209 calculated from 1973 to 2012. Furthermore, 1983 and 1988 were selected as the dry and
 2210 wet years respectively, as IWN during these years equalled or exceed 20 % and 80 % of
 2211 annual IWN from 1973 to 2012. This method is commonly adopted for these types of
 2212 studies (Knox et al., 1997).

2213

2214 **6.2.3 Statistical analysis**

2215 Statistical analysis in this study was conducted and is presented in two parts: model
2216 validation; and policy and climate scenario simulations.

2217

2218 **Validation procedure**

2219 In order to validate the model, and to explore the sensitivity of the model to variables
2220 which vary during initial model setup, 100 simulations were conducted under the cur-
2221 rent water allocation system scenario, for the period 2008 to 2012, in order to examine
2222 whether simulated abstractions for all 836 farmers were similar to measured abstrac-
2223 tions of the 90 respondents (see section 5.3.4). Two-tailed tests of association were
2224 conducted between the simulated abstractions of the 836 farmers and the measured
2225 abstractions of the 90 respondents, for both annual and average monthly abstractions.
2226 The null hypotheses for these tests were that there was no significant association be-
2227 tween simulated and measured abstractions. As in previous analysis, the Shapiro-Wilk
2228 tests, in addition to normal probability plots, were used to determine whether variables
2229 were normally distributed (see Appendix E). Depending on the results of these tests,
2230 the appropriate parametric (Pearson's correlation) or non-parametric (Spearman's rank
2231 correlation) test of association was used.

2232

2233 Therefore, with regards to the tests of association, if the *p-value* was less than the
2234 chosen alpha level, which for these tests were .05, then the null hypothesis was rejected
2235 suggesting that the variables were associated. Furthermore, a Pearson's correlation
2236 coefficient (r), or a Spearman's rank correlation coefficient (r_s), of plus or minus 0.7 or
2237 greater was considered a strong correlation; plus or minus 0.5 a moderate correlation;
2238 and plus or minus 0.3 a weak correlation.

2239

2240 **Policy and climate change scenario simulation procedure**

2241 As with the model validation, 100 simulations were conducted and the average values for
2242 each model indicator were reported (see Appendix E). Shapiro-Wilk tests, in addition
2243 to normal probability plots, were used to determine whether variables were normally
2244 distributed (see Appendix E). Depending on the results of these tests, the appropriate

2245 parametric or non-parametric test was then used to explore the null hypothesis that
2246 a statistically significant difference existed between the three policy scenarios, during
2247 both the dry and wet year climate scenarios, if the *p-value* was less than the chosen
2248 alpha level, which for these tests was .05, (one-way analysis of variance (ANOVA) or
2249 Kruskal-Wallis respectively). If such a statistically significant difference existed, then
2250 further relevant parametric or non-parametric tests were conducted to test the same null
2251 hypothesis but between each pair of farm types (independent *t*-test or Mann-Whitney
2252 *U* test respectively). However, in regards to Mann-Whitney *U* tests the difference in
2253 median values was reported if the two distributions were similar in shape; otherwise
2254 the difference in mean ranks was reported if distributions were not similar.
2255

2256 6.3 Results and discussion

2257 6.3.1 Model validation

2258 Figure 6.4 illustrates there were strong, positive, linear correlations between annual ab-
2259 stractions of the 90 respondents and annual abstractions of the 836 simulated farmers
2260 with regards to those who preferred LWUO ($r = .92$), which was statistically significant
2261 ($p = <.05$), thus the null hypothesis could be rejected; and with regards to those who
2262 had no preference ($r = .95$), which was also statistically significant ($p = <.05$), thus
2263 the null hypothesis could also be rejected. However, only a moderate, positive, linear
2264 correlation existed with regards to those who preferred HWUO ($r = .61$), which was
2265 not statistically significant ($p = >.05$), thus the null hypothesis could not be rejected.
2266

2267 Similarly, Figure 6.4 also illustrates there were strong, positive, linear correlations
2268 between average monthly abstractions of the 90 respondents and average monthly ab-
2269 stractions of the 836 simulated farmers with regards to those who preferred LWUO
2270 ($r = .77$), which was statistically significant ($p = <.05$), thus the null hypothesis could
2271 be rejected; and with regards to those who had no preference ($r = .91$), which was
2272 also statistically significant ($p = <.05$), thus the null hypothesis could also be rejected.
2273 Furthermore, only a moderate, positive, linear correlation existed with regards to those
2274 who preferred HWUO ($r = .58$), however this was statistically significant ($p = <.05$),
2275 thus the null hypothesis could be rejected.

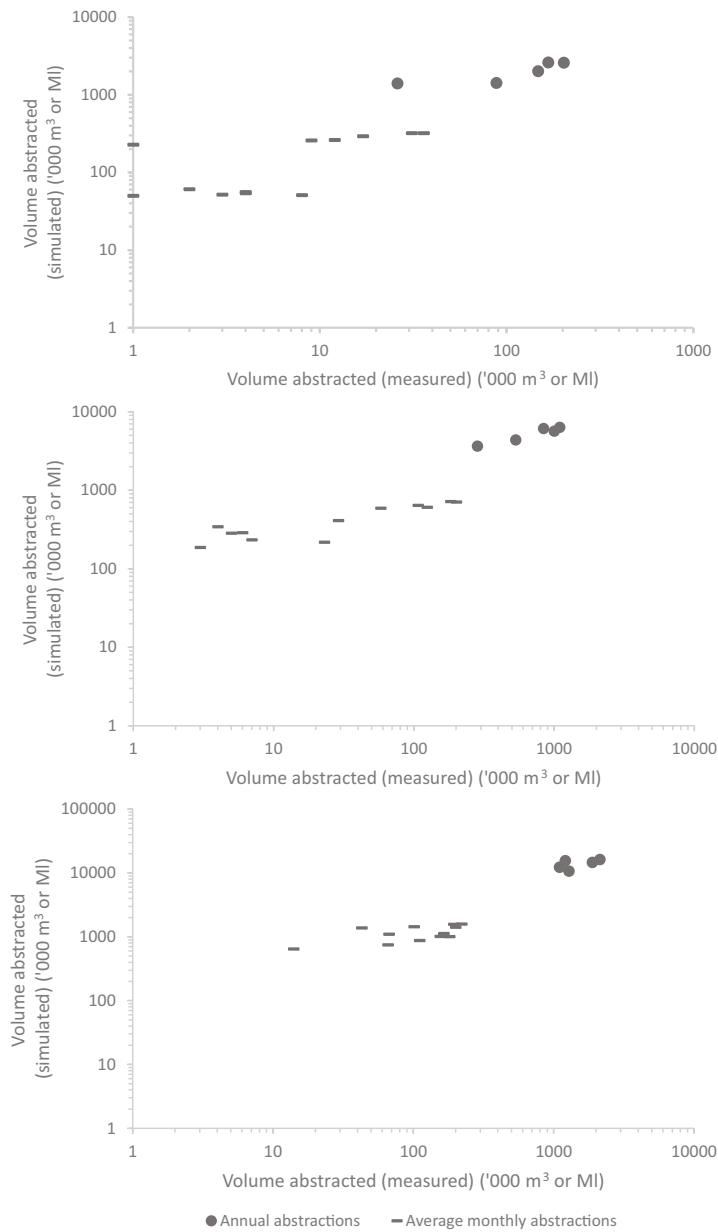


Figure 6.4: Scatter plots indicating the strength of association between simulated annual, and average monthly, abstractions for all 836 farmers and measured annual, and average monthly, abstractions for the 90 respondents, based on the average of 100 simulations. (Top) Pearson's correlation coefficients were $r = .92$, $p = <.05$, and $r = .77$, $p = <.05$, for annual and average monthly abstractions respectively for those who preferred low water usage options (LWUO). (Middle) Pearson's correlation coefficients were $r = .95$, $p = <.05$, and $r = .91$, $p = <.001$, for annual and average monthly abstractions respectively for those who had no preference. (Bottom) Pearson's correlation coefficients were $r = .61$, $p = >.05$, and $r = .58$, $p = <.05$ for annual and average monthly abstractions respectively for those who preferred high water usage options (HWUO)

2276 Overall, although differences between the measured and simulated abstractions oc-
2277 curred most likely due to over simplification of the model, particularity as HoF for
2278 individual licences were not included, the results indicate that the model, based on
2279 farmers' preferred behavioural intentions, was relatively reliable at simulating abstrac-
2280 tions for each farm type under the current water allocation system between 2008 and
2281 2012. The model could therefore be used to simulate system level patterns of abstrac-
2282 tion behaviour under the proposed basic and enhanced water allocation systems during
2283 dry and wet years with relative confidence.

2284

2285 6.3.2 Policy and climate change scenario simulations

2286 Kruskal-Wallis tests showed that statistically significant differences existed between
2287 policy scenarios with regards to monthly licence volume abstracted during the dry year
2288 climate scenario ($X^2= 7.002$) as $p= <.05$, and during the wet year climate scenario
2289 ($X^2= 18.029$) as $p= <.05$. Furthermore, Kruskal-Wallis tests showed that a statis-
2290 tically significant difference existed between policy scenarios with regards to surplus
2291 water available during the wet year climate scenario ($X^2= 8.979$) as $p= <.05$, but not
2292 during the dry year climate scenario ($X^2= .295$) as $p= >.05$. In addition, Kruskal-
2293 Wallis tests showed that no statistically significant differences existed between policy
2294 scenarios with regards to shortage water required during the dry year climate scenario
2295 ($X^2= .065$) as $p= >.05$, or during the wet year climate scenario ($X^2= .278$) as $p=$
2296 $>.05$. Similarly, Kruskal-Wallis tests showed that no statistically significant differences
2297 existed between policy scenarios with regards to the percentage of farm IWN satisfied
2298 (i.e. main crop potatoes and vegetables combined) during the dry year climate scenario
2299 ($X^2= .023$) as $p= >.05$, or during the wet year climate scenario ($X^2= .357$) as $p= >.05$
2300 (see Figure 6.5).

2301

2302 In particular, statistically significant differences existed with regards to monthly
2303 licence volume abstracted between the current and proposed enhanced water allocation
2304 systems during the dry year climate scenario ($U= 32$) as $p= <.05$, and during the
2305 wet year climate scenario ($U= 8$) as $p= <.05$. Furthermore, statistically significant
2306 differences also existed with regards to monthly licence volume abstracted between the
2307 proposed basic and proposed enhanced water allocation systems during the dry year

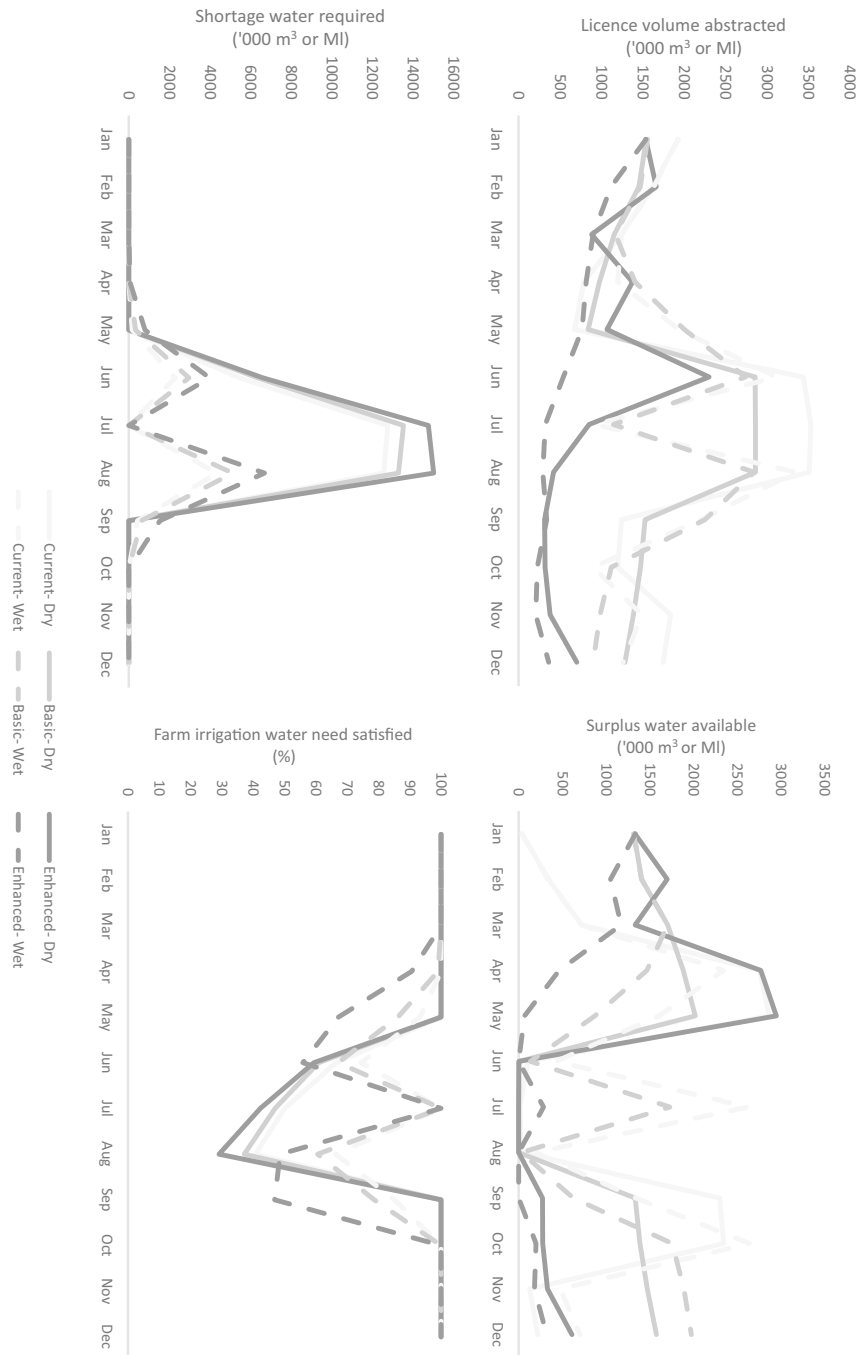


Figure 6.5: Line graphs indicating the potential system level patterns of abstraction behaviour which emerge from individual farm scale decisions under different policy and climate scenarios. (Top left) Line graph indicating the difference in monthly licensed abstractions. (Top right) Line graph indicating the difference in monthly surplus water available. (Bottom left) Line graph indicating the difference in monthly shortage water required. (Bottom right) Line graph indicating the difference in monthly farm irrigation water need (IWN) satisfied (i.e. main crop potatoes and vegetables combined)

2308 climate scenario ($U = 34$) as $p = <.05$, and during the wet year climate scenario ($U =$
2309 10) as $p = <.05$. Thus, the null hypothesis for this model indicator could be rejected
2310 as farmers abstracted significantly less under the proposed enhanced water allocation
2311 system, during both the dry (median 868 MI and IQR 1,106 MI) and wet (median 452
2312 MI and IQR 579 MI) year climate scenarios, compared to the current water allocation
2313 system, during both the dry (median 1,684 MI and IQR 1,854 MI) and wet (median
2314 1,561 MI and IQR 928 MI) year climate scenarios, and also compared to the proposed
2315 basic water allocation system, during both the dry (median 1,464 MI and IQR 1,340
2316 MI) and wet (median 1,440 MI and IQR 1,064 MI) year climate scenarios.

2317

2318 Similarly, statistically significant differences existed during the wet year climate
2319 scenario, with regards to surplus water available between the current and proposed
2320 enhanced water allocation systems ($U = 32$) as $p = <.05$, and between the proposed
2321 basic and proposed enhanced water allocation systems ($U = 24$) as $p = <.05$. Thus, the
2322 null hypothesis for this model indicator could be rejected as farmers during the wet
2323 year climate scenario had significantly less surplus water available under the proposed
2324 enhanced water allocation system (median 242 MI and IQR 887 MI) compared to the
2325 current (median 712 MI and IQR 1,805 MI) and proposed basic (median 1,437 MI and
2326 IQR 1,064 MI) water allocation systems.

2327

2328 Overall, the system level patterns of abstraction behaviour which emerged from
2329 farmers' preferred behavioural intentions indicate that farmers under the proposed en-
2330 hanced water allocation system abstracted less on average each month and therefore
2331 had less surplus water available compared to the current or proposed basic water allo-
2332 cation systems. This was due to farmers' monthly licensed water availability varying
2333 depending on water availability within AP areas (see Appendix E). Nonetheless, de-
2334 spite farmers abstracting less under this system, the percentage of farm IWN satisfied
2335 did not significantly vary under any of the policy scenarios. Therefore, the results of
2336 the simulations based on farmers' preferred behavioural intentions, would indicate that
2337 the proposed enhanced water allocation system, at least at the system level, offered the
2338 greatest protection for the environment as low flow conditions were enforced, and fewer
2339 abstractions occurred overall without significantly effecting the percentage of farm IWN
2340 satisfied.

2341 However, perhaps unsurprisingly, the results also indicate that surplus water was
2342 generally available outside of the growing season whilst a shortage of water was exhib-
2343 ited during the growing season. This would suggest that regardless of policy scenario,
2344 few trades are likely to occur between farmers. However, this is partly a model lim-
2345 itation as the growing season started at the same time for all farmers, thus increases
2346 and decreases in IWN occurred at the same time, whilst in reality it may be staggered.
2347 Nonetheless, even if the growing season was staggered, few trades are likely to occur
2348 but the results do highlight the potential for water licence trading with other sectors.

2349

2350 6.4 Summary

2351 Overall, this study has provided a first insight into how the proposed water allocation
2352 reforms, both in basic and enhanced catchments, are likely to change water availability
2353 for farmers. Furthermore, it has indicated the importance of understanding abstraction
2354 behaviour in determining the overall success of these reforms. In doing so, this chapter
2355 addressed the third of three research questions set out in the introduction to this thesis,
2356 and highlighted again at the start of this chapter. That was, what potential system
2357 level patterns of abstraction behaviour emerge from individual farm scale decisions un-
2358 der different policy and climate scenarios.

2359

2360 The development of an ABM was presented and validated by comparing simulated
2361 abstractions for the period 2008 to 2012 with measured abstractions for the same period
2362 based on farmers' preferred behavioural intentions. The policy and climate scenario
2363 simulations indicated that the proposed enhanced water allocation system offered the
2364 greatest protection to the environment at the system level, as a result of less water
2365 abstracted and low flow conditions being introduced, whilst not significantly reducing
2366 the percentage of farm IWN satisfied. Furthermore, despite model limitations, the sim-
2367 ulation results indicate few trades are likely to occur between farmers, even if preferred
2368 behavioural intentions were to change, as surplus water was generally only available
2369 outside of the growing season, whilst a shortage of water was indicated during the
2370 growing season. Nonetheless, the proposed enhanced water allocation system appears
2371 to offer the greatest potential of increasing efficiency with regards to balancing the
2372 needs of licence users, at least farmers, with the environment.

2373 Chapter 7

2374 Discussion and conclusions

2375 As discussed in the introduction to this thesis, water resource management in England
2376 is currently in a long-term transition from large-scale, high-cost, centralised infrastruc-
2377 ture projects to low-cost, community-scale, decentralised systems. A major aspect of
2378 this transition is the introduction of water markets where water licence trading can
2379 occur to improve efficiency. Water licence trading has been in operation in several
2380 countries for several decades and has been available in England since 2003 (DEFRA,
2381 2011). However, due to several reasons, mainly regarding institutional barriers which
2382 deter farmers from trading, very few trades have actually occurred in England. Similar
2383 institutional barriers were found to exist in other countries (Thobani, 1998; Zhang,
2384 2007). However, the current water allocation system in England is considered by the
2385 regulatory authorities as inadequate to provide users with the ability to adapt to climate
2386 change whilst providing adequate protection for the environment (EA, 2008; DEFRA,
2387 2011).

2388

2389 In response, DEFRA proposed the basic and enhanced water allocation systems to
2390 address the inadequacies of the current system and encourage water licence trading (see
2391 Chapter 1). However, it was argued that the success of the proposed water allocation
2392 systems, particularly with regards to water licence trading, very much depended on the
2393 preferred behavioural intentions of the users on the ground (Feola et al., 2015). This
2394 factor formed the rationale for this research as very little was understood with regards
2395 to how the proposed water allocation systems in England would be received by licence
2396 users, in particular by farmers, who were the main focus of this study (see Chapter 3).

2397

2398 Therefore, this thesis has presented three studies which have attempted to un-
2399 derstand farmers' behavioural intentions under different climate and proposed water
2400 allocation system scenarios, in order to assess whether the proposals will achieve their
2401 intended aim. These studies included: an empirical investigation within the Great Ouse
2402 catchment in eastern England to understand farmers' preferred behavioural intentions
2403 under different scenarios of water shortage and surplus with regards to the proposed
2404 water allocation systems; the development of a conceptual typology based on farmers
2405 preferred behavioural intentions rather than more traditional attributes such as income
2406 or farm size; and the development, and policy and climate scenario simulation results,
2407 of an ABM used to understand the system level patterns of abstraction behaviour which
2408 emerge from individual farm level decisions. This chapter discusses some of the main
2409 limitations of the methods used and attempts to place the results within the context
2410 of the wider literature.

2411

2412 **7.1 Understanding farmer behaviour**

2413 There is an increasing body of literature which supports the rationale that any individ-
2414 uals or group of individuals, who are directly affected by changes in policy, particularly
2415 with regards to water and land management, should be a key consideration in the
2416 design and implementation of that policy change (Austin et al., 1998; Burton, 2004;
2417 Emtage et al., 2006; Guillem et al., 2012). Moreover, several studies have shown that
2418 farmers' behaviour very much influences how and to what success policy proposals are
2419 realised on the ground (Bjornlund, 2003; Moon and Cocklin, 2011; Home et al., 2014;
2420 Feola et al., 2015). However, several challenges remain in understanding individuals'
2421 intentions and associated behaviours, and methods to incorporate these into practical
2422 policy applications.

2423

2424 The main challenges in understanding individual behaviours' are associated with the
2425 complexity and limitations of the theoretical behavioural models used. In particular, to
2426 what extent can current behavioural models accurately measure an individual's inten-
2427 tion, and to what extent is that intention a true measure of an individual's behaviour.
2428 The more reliable the behavioural model, the more likely policy decision-makers are
2429 to incorporate them into the design and implementation stages of policy changes. Al-

2430 though each model offers particular advantages and disadvantages, this research was
2431 based on one of the most widely used and influential models for predicting human so-
2432 cial behaviour, the TPB (Ajzen, 1985). However, as previously discussed one of the
2433 main criticisms of this model, in addition to normal survey limitation biases such as
2434 respondents answering how they think they should rather than honestly or whether
2435 only those responded who were interested in the subject, concerns efficacy (i.e. the
2436 predictive ability of the model) (Sniehotta et al., 2014).

2437

2438 Meta-analytical reviews found that the TPB explained between 40-49 % of the
2439 variance in intention (Sheeran et al., 1999; Armitage and Conner, 2001; Hagger et al.,
2440 2002; Trafimow et al., 2002; Schulze and Wittmann, 2003; McEachan et al., 2011). The
2441 adapted TPB model used in this study explained between 29-65 % of the variance in
2442 intention based on Nagelkerke's R^2 . However, as binary logistic regression was used
2443 rather than multiple linear regression, there are no direct analogues to the coefficient of
2444 determination (i.e. R^2) used in ordinary least squares, and therefore the aforementioned
2445 pseudo R^2 value was used (Bewick et al., 2005). As a result, these different method-
2446 ological approaches should be considered when comparing the results of this study with
2447 those reported in meta-analytical reviews. Although, the results of this study suggest
2448 a comparatively high percentage of variance explained, incorporating associated salient
2449 beliefs for each construct would likely improve the results. Therefore, a more robust
2450 methodology may consider using farmer focus groups to determine salient beliefs. This
2451 would have likely provided a more robust set of behaviours and potentially explained
2452 a higher percentage of variance in intention (Ajzen, 2011).

2453

2454 Attitude and subjective norm were found to be the main predictors of farmers' in-
2455 tention for three of the four scenarios: strategic (long-term) water shortage and water
2456 surplus scenarios; and the in-season (short-term) water shortage scenario. For the latter
2457 scenario, perceived behavioural control was also found to be a significant predictor of
2458 farmers' intention. Interestingly, previous studies also found attitude to be a significant
2459 predictor of farmers' behaviour, and not perceived behavioural control, suggesting that
2460 farmers tend to have complete volitional control with regards to long-term decision-
2461 making at the farm scale (Wauters et al., 2010; Poppenborg and Koellner, 2013).

2462

2463 Although this study has provided a first insight into how farmers' may respond to
2464 the proposed water allocation systems in England, there is clearly a considerable per-
2465 centage of variance in intention which was left unexplained and therefore other factors,
2466 not measured, could influence farmers' behavioural intentions. As previously men-
2467 tioned, a more robust methodology which measures salient beliefs may highlight some
2468 of these other factors. Furthermore, the intention-behaviour gap remains which is likely
2469 to exacerbate the longer the time interval between measuring intentions and farmers
2470 actually performing the behaviours (Orbell and Sheeran, 1998; Sniehotta et al., 2005).
2471 As the proposed water allocation systems are only being introduced in the early 2020s,
2472 farmers' intentions may change again, and therefore, this study should only be used
2473 as a preliminary understanding at this current time. Nonetheless, when water trading
2474 was first introduced in southeastern Australia in 1989 very few farmers participated in
2475 water licence trading for the first ten years. However, an increasing number of farm-
2476 ers participated in temporary licence trades as more farmers increasingly understood
2477 the operations and the advantages of a secure, reliable, fast, and cheap water transfer
2478 (Bjornlund, 2003).

2479

2480 **7.2 Validity and utility of a behavioural farm typology**

2481 Although farm typologies clearly offer policy decision-makers a useful tool for cate-
2482 gorising farms, and making more informed policy decisions, the main criticism of farm
2483 typologies, which this study attempted to address, is that they fail to fully capture true
2484 farmer behaviour and are therefore of limited value in supporting policy formulation
2485 (Guillem et al., 2012). Although the conceptual approach used in this study iden-
2486 tified three farm types, which featured significant differences in farm characteristics,
2487 alternative data-driven approaches such as cluster analysis (Barnes et al., 2011), or
2488 stratification of farmers based on particular farm characteristics, might have identified
2489 different farm types, and thus different preferred behavioural intentions. However, the
2490 conceptual approach used was considered more appropriate when using ordinal Likert-
2491 item measures of behaviour.

2492

2493 Nonetheless, although the validity of the typology is difficult to assess unless used in
2494 practice, it certainly offers utility for policy decision-makers who can use the informa-

2495 tion with regards to implementing the proposed water allocation systems in England.
2496 In particular, identifying which farmers are more or less likely to engage with particular
2497 aspects of the proposals, such as water licence trading, may be used in deciding which
2498 catchments should be categorised as basic or enhanced.

2499

2500 Within the wider sphere of policy decision-making, behavioural typologies clearly
2501 offer greater value in supporting policy formulation compared to more traditional ty-
2502 pologies. However, limitations also exist with regards to the added complexity involved
2503 in developing a behavioural typology (i.e. measuring individual behavioural intentions),
2504 and the effects of the intention-behaviour gap. In this sense, behavioural farm typolo-
2505 gies may be more valuable for policy formulation (Guillem et al., 2012) rather than
2506 continual monitoring of policy interventions, and policy decision-makers should not
2507 rely solely on behavioural intentions but also on more traditional farm characteristics
2508 (Schmitzberger et al., 2005; Emtage et al., 2006, 2007; Gorton et al., 2008; Pike, 2008;
2509 Barnes et al., 2011; Poppenborg and Koellner, 2013).

2510

2511 **7.3 Utility of the Agent-Based Model (ABM)**

2512 This research has highlighted the importance of understanding farmers' preferred be-
2513 havioural intentions with regards to the proposed water allocation systems in England.
2514 Furthermore, understanding the system level patterns of abstraction behaviour which
2515 could emerge from individual decision-making at the farm scale provides valuable infor-
2516 mation for policy decision-makers. However, although the ABM developed highlighted
2517 the potential effects of the proposed water allocation systems, several limitations of the
2518 model exist.

2519

2520 One of the main limitations of this model was its ability to only model in-season
2521 (short-term) water shortage and surplus preferred behavioural strategies, during a se-
2522 lected dry and wet year climate scenario, rather than model behaviours over a longer
2523 time period. With the latter approach, strategic (long-term) water shortage and sur-
2524 plus preferred behaviours could have been implemented, despite the added complexity
2525 of the model, to highlight the long-term effects of proposed policy changes. If the model
2526 had incorporated strategic behavioural strategies, then all farm types, during strategic

2527 (long-term)/water shortage scenarios, would have simply increased their application
2528 efficiency which would ultimately reduce abstractions at the system level during the
2529 growing season, regardless of policy or climate scenario. However, during strategic
2530 (long-term)/water surplus scenarios, more water would be abstracted during the grow-
2531 ing season as a result of those who preferred high water usage options (HWUO) growing
2532 the same crops but over a larger area. Furthermore, as the strategies for those who
2533 preferred low water usage options (LWUO), and those who had no preference, were to
2534 change nothing and sell surplus water for the duration of the growing season respec-
2535 tively, very little in regards to overall system level patterns of abstraction would have
2536 changed, as no trading occurred as buying was not a preferred behavioural strategy of
2537 any of the farmers. However, this highlights another limitation of the model.

2538

2539 The purpose of the model was to understand the potential system level patterns
2540 of abstraction behaviour which were likely to emerge based on farmers' preferred be-
2541 havioural intentions. However, as none of the farmers' preferred behavioural intentions
2542 involved buying more water, either strategically or in-season, no trading occurred in
2543 the model. Therefore, only the potential for trading at the system level was simulated
2544 based on surplus water available and shortage water indicated. Furthermore, although
2545 previous studies, such as that in southeastern Australia, have found that an increas-
2546 ing number of farmers participated in temporary water licence trading (i.e. weekly,
2547 monthly, or seasonal), very few participated in permanent water licence trading due to:
2548 differential tax treatment; policy uncertainty; institutional barriers such as administra-
2549 tive complexity and costs; and farmers' perception that water rights are an inherent
2550 part of their property (Bjornlund, 2003). Nonetheless, the results of the simulations
2551 highlighted that even if trading was available, very few trades would likely occur, even
2552 if farmers growing seasons would be staggered, as very little overlap would exist with
2553 regards to when surplus water was available and shortage water was required. There-
2554 fore, despite these limitations, the results of the model would suggest that farmers, who
2555 do not already have storage available, are best investing in storage to make use of the
2556 surplus water available out-of-season.

2557

2558 Another limitation of the model concerns the lack of planning and learning to change
2559 adaptive strategies. A more sophisticated model would perhaps incorporate the abil-

2560 ity of farmers to better plan their irrigation water needs (IWN) during the year. In
2561 the current model, this is simply divided by the number of months their licence was
2562 available. However, in reality farmers know, approximately, from previous years when
2563 demand is likely to be greatest and when it is likely to be less. Furthermore, if farmers
2564 continually experienced particular water shortage or surplus scenarios they may very
2565 well change their behavioural strategies, which would certainly alter the system level
2566 patterns of abstraction behaviour. Therefore, a model which simulates abstraction be-
2567 haviour over consecutive years could incorporate planning and learning from previous
2568 seasons. Data regarding how farmers actually plan their abstractions each year could
2569 be obtained during focus groups used to measure salient beliefs.

2570

2571 Overall, the model presented, as with any, was a simplified representation of reality.
2572 Nonetheless, despite the limitations discussed, the scenario simulation results indicated
2573 that the proposed enhanced water allocation system was likely to achieve its intended
2574 aim of increasing efficiency whilst balancing the needs of licence users and the environ-
2575 ment, more than the current or proposed basic water allocation system. Furthermore,
2576 incorporating other licence users would likely open up greater opportunities for water
2577 licence trading. However, although the proposed enhanced water allocation system was
2578 designed to facilitate trading more easily than the proposed basic system, the method
2579 used to derive water shares under the enhanced system may result in fewer trades as
2580 licences adjust to reflect users needs, and therefore less surplus water would be available
2581 compared to the proposed basic system.

2582

2583 Previous studies have also used ABM to simulate different policy or climate sce-
2584 narios related to individual behavioural strategies and changes in land use or resources
2585 (Schlüter and Pahl-Wostl, 2007; Galán et al., 2009; Acosta et al., 2014). However, one of
2586 the main limitations that each of these studies concluded with regards to ABM concerns
2587 the fact that as the models are developed based on behavioural strategies of particular
2588 groups of individuals, often within a particular geographic location, then generalising
2589 the model for a wider population or different geographic location may be difficult.
2590 Therefore, although the simulation results presented in this study are designed based
2591 on farmers currently operating within the Great Ouse catchment, the framework of the
2592 model, after adapting it to another area, could be generalised to other catchments.

7.4 Overview of conclusions

Farmer water abstraction behaviour in response to different policy and climate scenarios was investigated, in an area of eastern England, to assess whether the proposed water allocation systems in England would achieve their intended aim (i.e. to increase efficiency whilst balancing the needs of licences users and the environment). Three studies were conducted to understand farmers' preferred behavioural intentions, explore the utility of a farm typology developed based on farmers' behavioural intentions, and to examine the potential system level patterns of abstraction behaviour which emerge, under different policy and climate scenarios, based on farmers' preferred behavioural intentions. The main conclusions of this research are:

1. The TPB provides a relatively reliable method of successfully measuring farmers' intentions and underlying predictors. The TPB explained between 29-65 % of the variance in intention which was similar to the range found by meta-analytical reviews (i.e. 40-49 %). Furthermore, attitude and subjective norm were found to be the main significant predictors of farmers' intentions, at least in three of the four water shortage and surplus scenarios including: the strategic (long-term) water shortage and surplus scenarios; and the in-season (short-term)/water shortage scenario. In addition, with regards to the latter scenario, perceived behavioural control was also found to be a significant predictor of intention. Therefore, these results indicate that farmers, at least in regards to these behaviours, believe they have greater volitional control with regards to decision-making in the long-term compared to the short-term.

2. Behavioural farm typologies offer greater value in supporting policy formulations compared to traditional farm typologies. Developing a typology based on individual behavioural intentions provides greater utility to policy decision-makers, compared to more traditional typologies, as physical or economic attributes alone may fail to fully capture an individuals' intention. Depending on the policy formulation, this is likely to be of great interest, and very valuable to those involved in the policy formulation. However, the additional insights come at a cost with regards to added complexity in developing the typology (i.e. having to measure individual intentions), and inherent limitations with any behavioural approach such as the intention-

2625 behaviour gap (i.e. the utility of a behavioural typology becomes less the longer the
2626 period between policy formulation and intervention).

2627

2628 **3. Significant differences in farm characteristics existed between the**
2629 **farm types derived in this study.** Those who preferred Low Water Usage Op-
2630 tions (LWUO) were generally representative of farmers who irrigated smaller areas,
2631 used predominantly rainguns as their irrigation application mechanism, had less stor-
2632 age available, and were older and had a lower level of education compared to farmers
2633 who preferred High Water Usage Option (HWUO). In addition, those who preferred
2634 LWUO predominantly used rainguns as their main irrigation application mechanism
2635 whilst those who preferred HWUO, and those who had no preference, used a mix of
2636 different application mechanisms. Furthermore, those who preferred LWUO generally
2637 had a mix of both time limited and none-time limited licences whilst those who pre-
2638 ferred HWUO predominantly had time limited licences. Interestingly, those who had no
2639 preference were generally representative of farmers who had larger annual licence lim-
2640 its and past annual abstractions (similar to those who preferred HWUO) compared to
2641 farmers who preferred LWUO. However, those who had no preference also had smaller
2642 daily licence limits (similar to those who preferred LWUO) compared to those who
2643 preferred HWUO. In addition, those who had no preference abstracted more in-season
2644 compared to those who preferred LWUO.

2645

2646 **4. All farm types' preferred Low Water Usage Option (LWUO) under**
2647 **strategic (long-term) and in-season (short-term) water shortage scenarios.**
2648 Despite farmers' overall preferred behaviour in regards to all 18 behaviours assessed,
2649 each farm type preferred LWUO during water shortage scenarios. Under a strategic
2650 (long-term) scenario this was to increase application efficiency. Under an in-season
2651 (short-term) scenario this was to only use their maximum abstraction licence to irri-
2652 gate their most valuable crops. Although this is not surprising for those who preferred
2653 LWUO, or perhaps even for those who had no preference. The feasibility or perhaps
2654 lack of experience under the current water allocation system could have contributed
2655 to the preferred behaviour of those who preferred HWUO. However, during water sur-
2656 plus scenarios, those who preferred LWUO and those who had no preference preferred
2657 LWUO, and those who preferred HWUO preferred HWUO.

2658 **5. ABM offer a valuable tool for policy decision-makers involved with**
2659 **policy formulation.** In the context of this study, the ABM developed and used to
2660 understand what system level patterns of abstraction behaviour emerge based on in-
2661 dividual farm level decision-making, indicated the proposed enhanced water allocation
2662 system provided the greatest ability to balance the needs of licence users whilst protect-
2663 ing the environment. However, this ABM did not incorporate licence users from other
2664 sectors in the catchment and therefore generalising ABM results to wider populations
2665 or other geographic locations may be difficult. Therefore, although ABM can provide
2666 an insight into system level patterns of policy interventions, they are perhaps limited
2667 in their utility to the area or population of interest.

2668

2669 **6. The success of the proposed water allocation systems, with regards**
2670 **to encouraging water licence trading is dependent on both institutional**
2671 **changes and changes in users' intentions.** The importance of institutions being
2672 designed to encourage and facilitate water licence trading was discussed in the introduc-
2673 tion to this thesis (see Chapter 1). However, as evident in several countries, including
2674 the UK, one of the main reasons farmers do not trade was due to high transaction costs
2675 and where transaction costs were low reasons included: management, legal, adminis-
2676 trative, and fiscal barriers (Zhang, 2007). Therefore, although policy decision-makers
2677 can utilise the results of this study to identify which farmers are more or less likely to
2678 engage with water licence trading, without changing current institutional barriers, the
2679 proposed water allocation systems are not likely to succeed.

2680

2681 **7. Although water licence trading is likely to increase efficiency at the**
2682 **system level this should be coupled with a long-term strategy to increase**
2683 **storage capacity and application efficiency at the farm level.** The results of the
2684 policy and climate scenario simulations suggest that very few trades are likely to occur
2685 between farmers. However, this is also due to several limitations within the model, but
2686 nonetheless very little overlap appears to occur between surplus water being available
2687 and shortage water being required. Where overlap does occur, trading would likely
2688 provide a viable method of improving efficiency. However, for the majority of farmers,
2689 policies focussing on increasing storage capacity and improving application efficiency
2690 are more likely to increase efficiency at the system level.

2691 7.5 Further research

2692 This research provides a first investigation into understanding farmers' preferred be-
2693 havioural intentions with regards to the proposed water allocation systems in England.
2694 In particular, the three studies presented attempted to address the three research gaps
2695 identified in Chapter 2. These included: understanding farmers' preferred behavioural
2696 intentions under different scenarios of water shortage and surplus with regards to the
2697 proposed water allocation systems; the potential utility of developing a farm typology
2698 based on farmers' preferred behavioural intentions with regards to the proposed wa-
2699 ter allocation systems; and lastly understanding the potential system level patterns of
2700 abstraction behaviour which emerge from individual farm level decision-making, under
2701 different policy and climate scenario, with regards to the proposed water allocation
2702 systems. However, although the methods and results of this research have direct policy
2703 application they have also highlighted several limitations and areas for further research.

2704

2705 In regards to measuring individual behaviour using the TPB, this research has
2706 identified several limitations with the approach used in this study. Further research
2707 is required to improve the variance in intention explained by predictor variables by
2708 either measuring farmers' salient beliefs, within the theoretical framework of the TPB
2709 (Ajzen, 2011), or adopting an alternative behavioural approach such as the temporal
2710 self-regulation theory (Hall and Fong, 2007), which emphasises temporal dynamics and
2711 lends itself more readily to experimental tests (Sniehotta et al., 2014). In addition,
2712 focus groups may provide a more valuable approach in understanding individual be-
2713 haviours (Wauters et al., 2010). However, measuring salient beliefs was considered in
2714 the design of the study but a more direct approach was adopted in order to assess a
2715 greater number of behaviours. Therefore, further research may consider focussing on
2716 fewer behaviours but in more detail.

2717

2718 In addition, further research is required to understand the full validity and utility of
2719 behavioural farm typologies used for policy formulation, intervention, and monitoring
2720 (Barnes et al., 2011). This research only highlighted the potential utility of the typology
2721 for policy formulation based on farmers' preferred behavioural intentions. However, if
2722 behavioural typologies are to become more widely used, then further studies are re-

2723 quired to assess their practical applications, and the extent of the usefulness in policy
2724 decision-making (i.e. do individuals perform the behaviours they intended).

2725

2726 Further research is also required to simulate water trading at the system level un-
2727 der the different policy and climate scenarios. Although this study highlighted the
2728 potential for water trading by measuring surplus water available and shortage water
2729 required, a more sophisticated model could incorporate an economic element to indi-
2730 vidual farmers' strategies. Zhang et al. (2014) concluded a similar observation, arguing
2731 that further research with regards to water trading between farmers is required, and
2732 should preferably use a dataset with sufficient observations on water transactions in
2733 order to test the assumptions we draw regarding factors which drive developing water
2734 markets. In addition, more work is required to understand how the proposed water al-
2735 location systems in England would operate when other licence users are included from
2736 other sectors, such as public water supply companies and electricity supply industries.

2737

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Appendix A

Questionnaire

Section A: This section is about your general irrigation practice			
What are the three main crops you usually irrigate (e.g. early potatoes, main crop potatoes, sugar beet, orchard fruit, small fruit, vegetables, grass, cereals)?			
How large an irrigated area do these main crops usually cover (ha or acres)?	ha acres	ha acres	ha acres
What is the predominant soil type these crops usually grow on?	Sand <input type="checkbox"/> Loam <input type="checkbox"/> Peat <input type="checkbox"/>	Sand <input type="checkbox"/> Loam <input type="checkbox"/> Peat <input type="checkbox"/>	Sand <input type="checkbox"/> Loam <input type="checkbox"/> Peat <input type="checkbox"/>
What type of irrigation application infrastructure do you usually use?	Rainguns <input type="checkbox"/> Booms <input type="checkbox"/> Sprinklers <input type="checkbox"/> Trickle <input type="checkbox"/>	Rainguns <input type="checkbox"/> Booms <input type="checkbox"/> Sprinklers <input type="checkbox"/> Trickle <input type="checkbox"/>	Rainguns <input type="checkbox"/> Booms <input type="checkbox"/> Sprinklers <input type="checkbox"/> Trickle <input type="checkbox"/>
If applicable, what is the approximate size of your water storage capacity? (Please indicate unit of measurement such as mega litres, cubic metres, millions of gallons, acre-inches)			

Section D: Demographic Questions

What is your gender?

Male

Female

What is your age?

18-29

30-39

40-49

50-59

60-69

70 +

What is your highest level of education completed?

NQF Level *Example Qualification*

Entry Level: Entry level certificate/ Foundation diploma/ BTEC Level 1

Level 1-2: O-Levels (A*-G)/ Higher Dip

Level 3-5: A-Levels (A*-E)/ Advance Dip/ Foundation Deg/ HND/ Dip FE

Level 6: Bachelors (Honours) Degree/ Graduate Dip/ Professional CE

Level 7: Masters Degree/ PGDip/ PGCert/ Postgraduate CE

Level 8: Doctoral Degree (Doctorates and Higher Doctorates)

On average, what is your total gross annual farm business income?

£0-4,999 £5-9,999 £10-£14,999 £15-£19,999 £20-£29,999

£30-£39,999 £40-£49,999 £50-74,999 £75-£99,999 £100,000+

Have you used your licence at all in the last 10 years?

Yes

No

Appendix B

Summary statistics of farm characteristics

Table B.1: Farm attributes (section A of survey)

Crop category	Irrigated (%)	Area (ha)	Soil type (%)				Irrigation method (%)				
			Sa	Lo	Pe	M	R	B	S	T	M
Main crop 1 ($n=90$)											
- Early pot.	10	503	56	33	11	0	67	11	0	0	22
- Main crop pot.	72	4,819	28	40	20	12	65	11	2	0	23
- Sugar beet	2	110	50	0	50	0	100	0	0	0	0
- Orchard fruit	1	3	0	100	0	0	0	0	0	100	0
- Small fruit	0	0	0	0	0	0	0	0	0	0	0
- Vegetables	13	1,243	50	25	8	17	50	17	8	0	25
- Grass	0	0	0	0	0	0	0	0	0	0	0
- Cereals	0	0	0	0	0	0	0	0	0	0	0
- Other	1	20	0	0	0	0	0	0	0	0	100
- Missing	0	-	-	-	-	-	-	-	-	-	-
Total	100	6,699	34	37	18	11	62	11	2	1	23
Main crop 2 ($n=54$)											
- Early pot.	0	0	0	0	0	0	0	0	0	0	0
- Main crop pot.	10	930	11	33	33	22	44	11	0	0	44
- Sugar beet	12	191	20	50	10	20	70	30	0	0	0
- Orchard fruit	0	0	0	0	0	0	0	0	0	0	0
- Small fruit	1	1	0	100	0	0	0	0	0	100	0
- Vegetables	32	1,634	57	18	14	11	62	24	0	0	14
- Grass	0	0	0	0	0	0	0	0	0	0	0
- Cereals	3	200	0	67	0	33	67	0	0	0	33
- Other	1	4	100	0	0	0	100	0	0	0	0
- Missing	40	-	-	-	-	-	-	-	-	-	-
Total	100	2,959	38	31	15	15	60	21	0	2	17
Main crop 3 ($n=39$)											
- Early pot.	0	0	0	0	0	0	0	0	0	0	0
- Main crop pot.	1	16	100	0	0	0	100	0	0	0	0
- Sugar beet	9	504	43	14	0	43	63	13	0	0	25
- Orchard fruit	0	0	0	0	0	0	0	0	0	0	0
- Small fruit	0	0	0	0	0	0	0	0	0	0	0
- Vegetables	21	1,540	37	37	16	11	42	37	5	0	16
- Grass	1	12	0	100	0	0	100	0	0	0	0
- Cereals	10	925	22	44	11	22	67	33	0	0	0
- Other	1	4	100	0	0	0	100	0	0	0	0
- Missing	57	-	-	-	-	-	-	-	-	-	-
Total	100	3,001	37	34	11	18	56	28	3	0	13
Total (amalgamation of main crops 1, 2 and 3 excluding missing data)											
- Early pot.	5	503	56	33	11	0	67	11	0	0	22
- Main crop pot.	41	5,765	27	39	21	13	63	11	1	0	25
- Sugar beet	11	805	32	32	11	26	70	20	0	0	10
- Orchard fruit	1	3	0	100	0	0	0	0	0	100	0
- Small fruit	1	1	0	100	0	0	0	0	0	100	0
- Vegetables	33	4,417	49	25	14	12	53	27	3	0	17
- Grass	1	12	0	100	0	0	100	0	0	0	0
- Cereals	7	1,125	17	50	8	25	67	25	0	0	8
- Other	2	28	100	0	0	0	67	0	0	0	33
Total	100	12,659	36	35	16	14	60	18	2	1	19

Summing errors due to rounding. Potatoes (pot.); Sand (Sa); Loam (Lo); Peat (Pe); Mixed (M); Raingun (R); Boom (B); Sprinkler (S); Trickle (T)

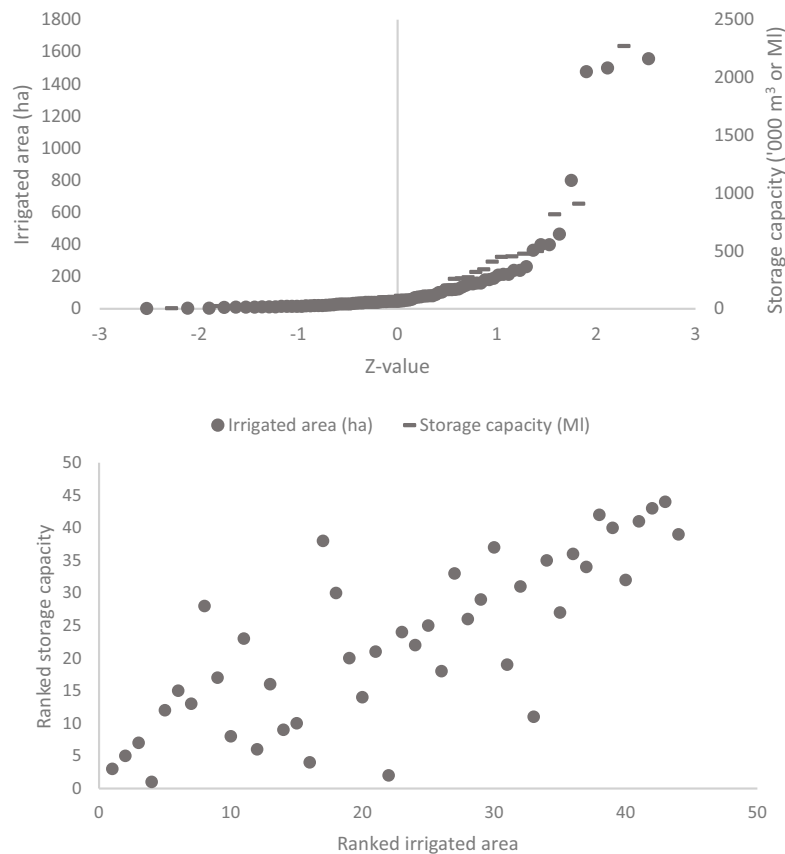


Figure B.1: A normal probability plot illustrating distribution, and a scatter plot indicating strength of association, with regards to irrigated area and storage capacity. (Top) A normal probability plot illustrating that neither variables were normally distributed which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for both variables. (Bottom) A scatter plot indicating the strength of association between ranked irrigated area and ranked storage capacity (Spearman's rank correlation coefficient $r_s = .76$, $p = <.001$)

Table B.2: Social demographics (section D of survey) ($n = 90$)

Measure	Respondents (%)
Gender	
- Male	93
- Female	2
- Missing	4
Age	
- 18-29	7
- 30-39	8
- 40-49	21
- 50-59	28
- 60-69	23
- ≥ 70	11
- Missing	2
^a Education (NQF level)	
- <1	3
- 1-2	29
- 3-5	22
- 6	28
- 7	6
- 8	0
- Missing	12
^b Income (£)	
- 0-4,999	0
- 5-9,999	0
- 10-14,999	0
- 15-19,999	1
- 20-29,999	3
- 30-39,999	0
- 40-49,999	0
- 50-74,999	8
- 75-99,999	8
- $\geq 100,000$	66
- Missing	14
Licence use (last 10 years)	
- Yes	96
- No	2
- Missing	2

^aNational Qualification Framework (NQF) levels for England, Wales and Northern Ireland

^bTotal gross annual farm business income

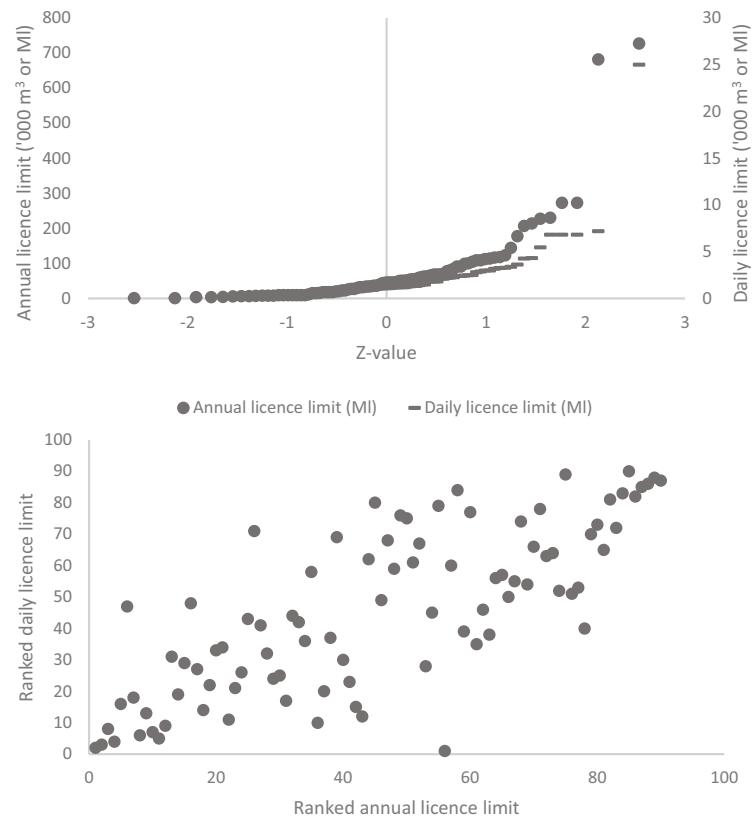


Figure B.2: A normal probability plot illustrating distribution, and a scatter plot indicating strength of association, with regards to annual and daily licence limits. (Top) A normal probability plot illustrating that neither variables were normally distributed which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for both variables. (Bottom) A scatter plot indicating the strength of association between ranked annual licence limit and ranked daily licence limit (Spearman's rank correlation coefficient $r_s = .70$, $p = <.001$)

Table B.3: Past abstractions between 2008 and 2012 ('000 m³ or Ml)

Month	2008	2009	2010	2011	2012	Average	S.D.
Jan	107	96	232	324	67	165	108.89
Feb	46	33	208	285	57	126	113.81
Mar	35	4	13	149	151	70	73.36
Apr	26	35	50	216	162	98	85.87
May	152	274	348	661	269	341	192.10
Jun	289	646	561	509	172	435	197.82
Jul	428	334	883	388	120	431	279.61
Aug	176	271	172	289	193	220	55.21
Sep	61	237	35	104	112	110	77.71
Oct	4	72	6	78	32	39	35.27
Nov	201	79	237	148	199	173	61.17
Dec	191	169	284	242	58	189	85.64
Annual	1,715	2,249	3,028	3,393	1,591	2,395	794.71

Summing errors due to rounding

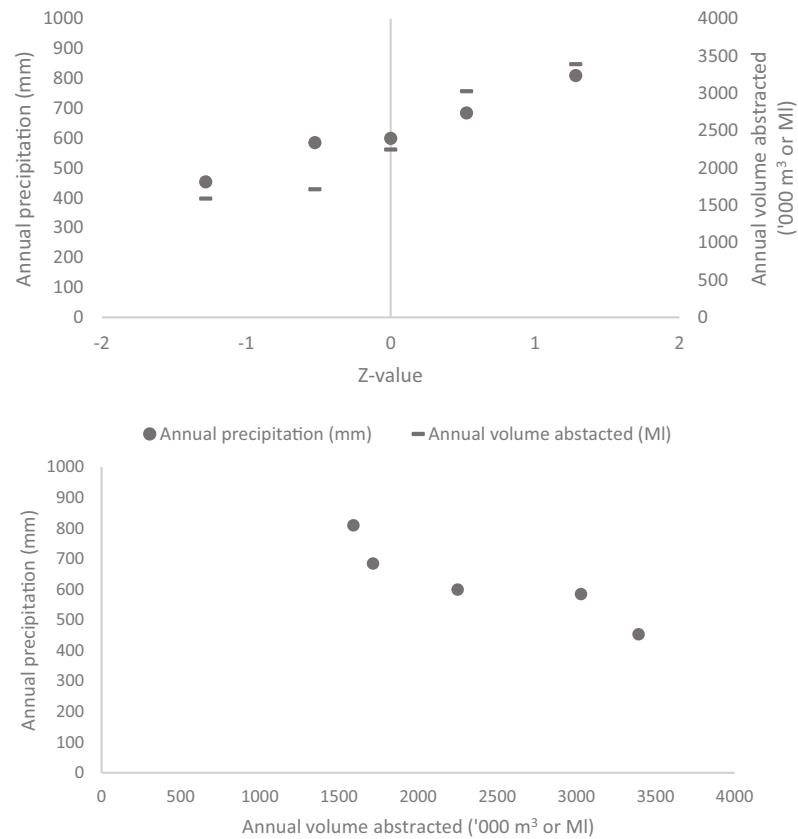


Figure B.3: A normal probability plot illustrating distribution, and a scatter plot indicating strength of association, with regards to annual abstractions and precipitation. (Top) A normal probability plot illustrating that both variables were normally distributed which supports the results of the Shapiro-Wilk tests of normality as $p = >.05$ for both variables. (Bottom) A scatter plot indicating the strength of association between annual abstractions and annual precipitation (Pearson's correlation coefficient $r = -.91$, $p = <.05$)

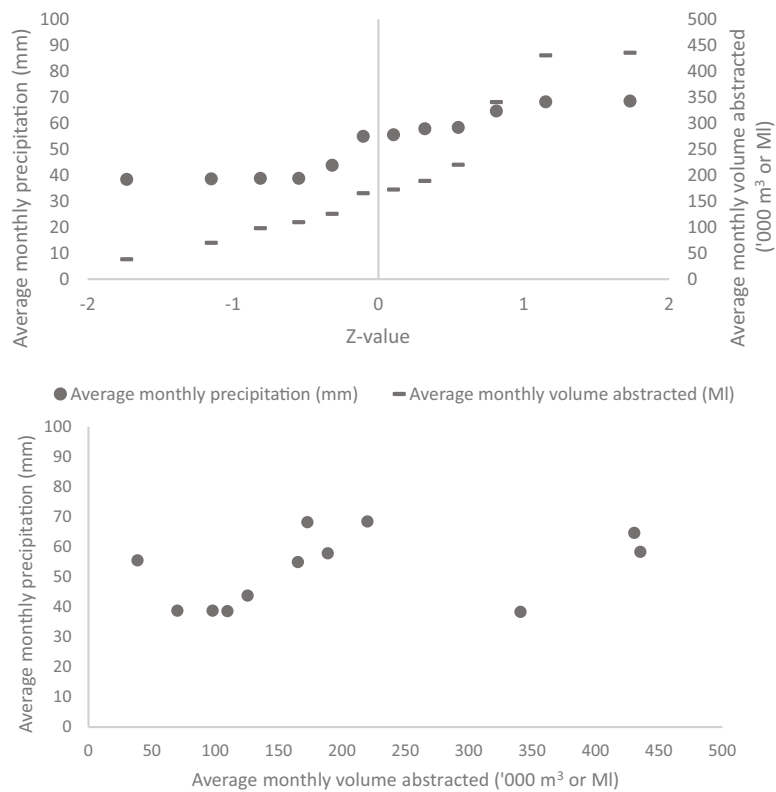


Figure B.4: A normal probability plot illustrating distribution, and a scatter plot indicating strength of association, with regards to average monthly abstractions and precipitation. (Top) A normal probability plot illustrating that both variables were approximately normally distributed which supports the results of the Shapiro-Wilk tests of normality as $p = >.05$ for both variables. (Bottom) A scatter plot indicating the strength of association between average monthly abstractions and average monthly precipitation (Pearson's correlation coefficient $r = .36$, $p = >.05$)

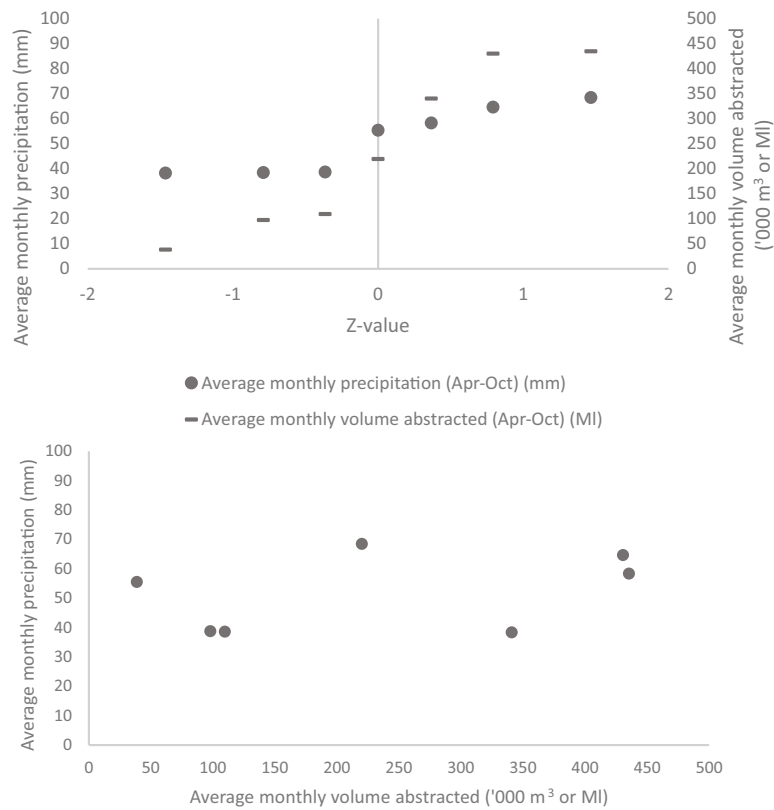


Figure B.5: A normal probability plot illustrating distribution, and a scatter plot indicating strength of association, with regards to average monthly abstractions and precipitation in-season (Apr-Oct). (Top) A normal probability plot illustrating that both variables were approximately normally distributed which supports the results of the Shapiro-Wilk tests of normality as $p = >.05$ for both variables. (Bottom) A scatter plot indicating the strength of association between average monthly abstractions and average monthly precipitation in-season (Apr-Oct) (Pearson's correlation coefficient $r = .36$, $p = >.05$)

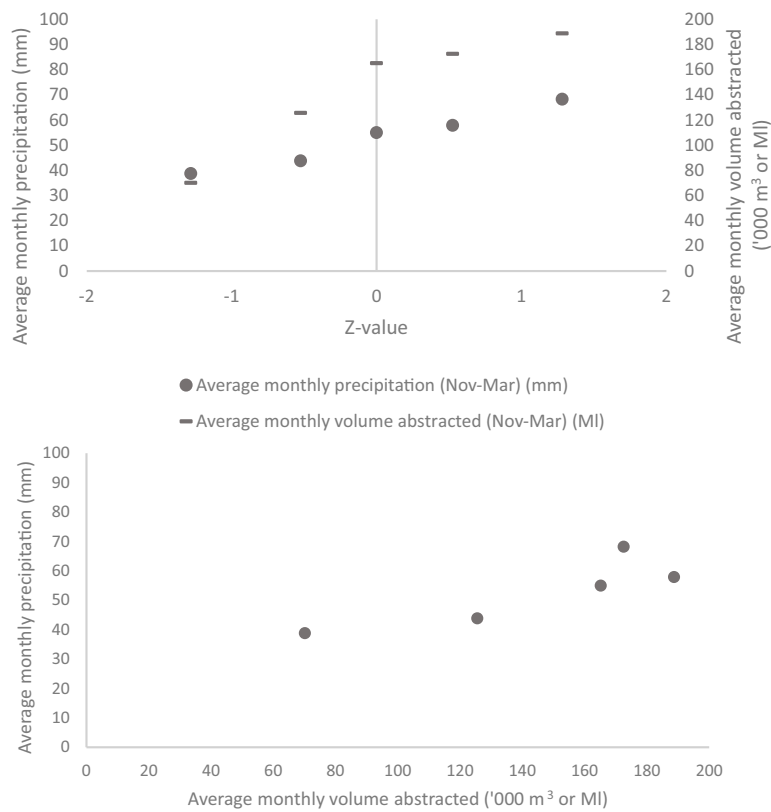


Figure B.6: A normal probability plot illustrating distribution, and a scatter plot indicating strength of association, with regards to average monthly abstractions and precipitation out-of-season (Nov-Mar). (Top) A normal probability plot illustrating that both variables were approximately normally distributed which supports the results of the Shapiro-Wilk tests of normality as $p = >.05$ for both variables. (Bottom) A scatter plot indicating the strength of association between average monthly abstractions and average monthly precipitation out-of-season (Nov-Mar) (Pearson's correlation coefficient $r = .87$, $p = >.05$)

Appendix C

Summary statistics of behavioural intentions

Table C.1: Strategic (long-term)/water shortage behavioural intentions (section B of survey)

Behaviour	^b Likert item responses (%) (<i>n</i> =90)								Median	IQR
^a Construct	1	2	3	4	5	6	7	Missing		
Grow the same crops but over a smaller area										
- I	16	7	16	14	32	6	7	3	4	2
- A	17	28	41	4	6	0	0	4	3	1
- SN	7	10	18	23	28	7	0	8	4	2
- PBC	7	9	12	3	32	10	22	4	5	3
- PB	0	1	3	1	1	0	0	93	3	2
Grow less water intensive crops										
- I	17	9	22	12	26	9	0	6	3	3
- A	11	17	38	16	13	0	0	6	3	2
- SN	9	10	21	16	30	4	0	10	4	2
- PBC	8	2	20	2	24	16	22	6	5	3
- PB	0	1	2	1	2	0	0	93	4	2
Increase storage capacity										
- I	8	4	9	10	23	13	23	9	5	3
- A	4	2	6	6	32	19	19	12	5	1
- SN	2	1	2	13	37	12	20	12	5	1
- PBC	10	4	23	2	22	14	16	8	5	3
- PB	1	0	2	1	1	0	1	93	4	3
Increase application efficiency										
- I	1	2	2	10	36	24	18	7	5	1
- A	1	0	0	2	46	29	17	6	5	1
- SN	1	0	0	8	38	24	20	9	5	1
- PBC	11	1	20	4	26	17	14	7	5	3
- PB	0	0	1	1	2	0	1	94	5	2
Buy more water for the duration of the growing season										
- I	10	4	13	22	31	7	7	6	4	2
- A	2	7	10	22	27	19	7	7	5	2
- SN	2	0	4	32	32	12	4	12	5	1
- PBC	6	6	19	4	32	13	13	7	5	3
- PB	0	0	2	1	2	2	0	92	5	3
Apply for a larger abstraction licence										
- I	7	8	14	13	36	3	11	8	5	2
- A	2	3	7	21	30	14	13	9	5	2
- SN	4	1	11	23	26	14	9	11	5	2
- PBC	9	8	20	4	19	12	20	8	5	3
- PB	0	1	0	1	2	1	0	94	5	3
Change nothing										
- I	33	8	22	13	9	2	7	6	3	3
- A	17	11	36	21	7	1	1	7	3	2
- SN	17	10	30	23	8	1	0	11	3	2
- PBC	12	0	12	3	30	13	22	7	5	3
- PB	1	1	3	1	0	0	1	92	3	2

^aTheory of Planned Behaviour constructs: Intention (I); Attitude (A); Subjective Norm (SN); Perceived Behavioural Control (PBC); and Past Behaviour (PB)

^b(I), (PBC), and (PB) = strongly disagree (1) to strongly agree (7); (A) = extremely disadvantageous (1) to extremely advantageous (7); and (SN) = definitely shouldn't (1) to definitely should (7) (4 = not sure for all constructs)

Table C.2: Preferred strategic (long-term)/water shortage behavioural intentions

Behaviour	^b Rescaled Likert item responses							\sum Totals	^c Sign test	
^a Construct	-3	-2	-1	(n=)	1	2	3	Total	(-,+,=)	
Increase application efficiency									425	
- I	-3	-4	-2	(84)	32	44	48	115		29,15,37*
- A	-3	0	0	(85)	41	52	45	135		
- SN	-3	0	0	(82)	34	44	54	129		
- PBC	-30	-2	-18	(84)	23	30	39	42		
- PB	0	0	-1	(5)	2	0	3	4		
Increase storage capacity									294	
- I	-21	-8	-8	(82)	21	24	63	71		33,15,32*
- A	-12	-4	-5	(79)	29	34	51	93		
- SN	-6	-2	-2	(79)	33	22	54	99		
- PBC	-27	-8	-21	(83)	20	26	42	32		
- PB	-3	0	-2	(6)	1	0	3	-1		
Apply for a larger abstraction licence									182	
- I	-18	-14	-13	(83)	32	6	30	23		26,19,38
- A	-6	-6	-6	(82)	27	26	36	71		
- SN	-12	-2	-10	(80)	23	26	24	49		
- PBC	-24	-14	-18	(83)	17	22	54	37		
- PB	0	-2	0	(5)	2	2	0	2		
Buy more water for the duration of the growing season									164	
- I	-27	-8	-12	(85)	28	12	18	11		42,21,21*
- A	-6	-12	-9	(84)	24	34	18	49		
- SN	-6	0	-4	(79)	29	22	12	53		
- PBC	-15	-10	-17	(84)	29	24	36	47		
- PB	0	0	-2	(7)	2	4	0	4		
Grow less water intensive crops									-85	
- I	-45	-16	-20	(85)	23	16	0	-42		15,32,38*
- A	-30	-30	-34	(85)	12	0	0	-82		
- SN	-24	-18	-19	(81)	27	8	0	-26		
- PBC	-21	-4	-18	(85)	22	28	60	67		
- PB	0	-2	-2	(6)	2	0	0	-2		
Grow the same crops but over a smaller area									-95	
- I	-42	-12	-14	(87)	29	10	18	-11		52,17,14*
- A	-45	-50	-37	(86)	5	0	0	-127		
- SN	-18	-18	-16	(83)	25	12	0	-15		
- PBC	-18	-16	-11	(86)	29	18	60	62		
- PB	0	-2	-3	(6)	1	0	0	-4		
Change nothing									-199	
- I	-90	-14	-20	(85)	8	4	18	-94		
- A	-45	-20	-32	(84)	6	2	3	-86		
- SN	-45	-18	-27	(80)	7	2	0	-81		
- PBC	-33	0	-11	(84)	27	24	60	67		
- PB	-3	-2	-3	(7)	0	0	3	-5		

^aTheory of Planned Behaviour constructs: Intention (I); Attitude (A); Subjective Norm (SN); Perceived Behavioural Control (PBC); and Past Behaviour (PB)

^bLikert item scores rescaled from -3 to 3 and multiplied by number of responses

^cPositive difference (+); negative difference (-); and ties (=). (*) = $p < .05$.

Table C.3: Tests for multi-collinearity, and recommended number of cases, in relation to respondents' preferred strategic (long-term)/water shortage behavioural intention. Tests for multi-collinearity of the predictor variables (i.e. attitude, subjective norm (SN), and perceived behavioural control (PBC)) in relation to the binary response variable (i.e. intention) where 0 = strongly disagree to agree and 1 = very much agree to strongly agree with the intention to increase application efficiency ($n = 79$)

Measure	Attitude	SN	PBC	VIF	Frequency	
					0	1
- Intention	.388	.467	.135			
- Attitude		.416	.044	1.210	42	37
- SN			.134	1.230	41	38
- PBC				1.019	53	26

Variance Inflation Factor (VIF)

Table C.4: Strategic (long-term)/water surplus behavioural intentions (section B of survey)

Behaviour - ^a Construct	^b Likert item responses (%) (<i>n</i> =90)							Missing	Median	IQR
	1	2	3	4	5	6	7			
Grow the same crops but over a larger area										
- I	3	2	41	17	21	3	3	9	3	2
- A	4	7	24	19	32	6	3	4	4	2
- SN	3	4	19	28	24	6	1	14	4	2
- PBC	4	2	17	2	39	6	23	7	5	4
- PB	7	2	21	2	9	1	1	57	3	2
Grow more water intensive crops										
- I	4	4	52	14	12	3	0	9	3	1
- A	4	6	32	26	23	4	0	4	4	2
- SN	7	3	31	27	17	2	0	13	4	1
- PBC	4	2	22	2	34	7	21	7	5	3
- PB	9	1	24	1	6	2	0	57	3	0
Sell surplus water for the duration of the growing season										
- I	7	7	21	14	34	4	6	7	4	2
- A	7	2	6	20	48	8	6	4	5	1
- SN	3	2	9	32	31	7	2	13	4	1
- PBC	11	2	21	8	28	8	16	7	5	3
- PB	8	1	17	4	7	4	1	58	3	2
Change nothing										
- I	8	3	18	19	30	8	8	7	4	2
- A	7	4	26	34	18	1	6	4	4	2
- SN	6	7	23	30	16	4	0	14	4	1
- PBC	9	2	20	9	26	4	24	6	5	4
- PB	4	1	12	4	17	1	10	50	5	2

^aTheory of Planned Behaviour constructs: Intention (I); Attitude (A); Subjective Norm (SN); Perceived Behavioural Control (PBC); and Past Behaviour (PB)

^b(I), (PBC), and (PB) = strongly disagree (1) to strongly agree (7); (A) = extremely disadvantageous (1) to extremely advantageous (7); and (SN) = definitely shouldn't (1) to definitely should (7) (4 = not sure for all constructs)

Table C.5: Preferred strategic (long-term)/water surplus behavioural intentions

Behaviour	^b Rescaled Likert item responses								\sum Totals	^c Sign test
^a Construct	-3	-2	-1	(n=)	1	2	3	Total		(-,+,=)
Sell surplus water for the duration of the growing season									82	
- I	-18	-12	-19	(84)	31	8	15	5		29,29,21
- A	-18	-4	-5	(86)	43	14	15	45		
- SN	-9	-4	-8	(78)	28	12	6	25		
- PBC	-30	-4	-19	(84)	25	14	42	28		
- PB	-21	-2	-15	(38)	6	8	3	-21		
Change nothing									47	
- I	-21	-6	-16	(84)	27	14	21	19		41,21,17*
- A	-18	-8	-23	(86)	16	2	15	-16		
- SN	-15	-12	-21	(77)	14	8	0	-26		
- PBC	-24	-4	-18	(85)	23	8	66	51		
- PB	-12	-2	-11	(45)	15	2	27	19		
Grow the same crops but over a larger area									36	
- I	-9	-4	-37	(82)	19	6	9	-16		25, 6,51*
- A	-12	-12	-22	(86)	29	10	9	2		
- SN	-9	-8	-17	(77)	22	10	3	1		
- PBC	-12	-4	-15	(84)	35	10	63	77		
- PB	-18	-4	-19	(39)	8	2	3	-28		
Grow more water intensive crops									-80	
- I	-12	-8	-47	(82)	11	6	0	-50		
- A	-12	-10	-29	(86)	21	8	0	-22		
- SN	-18	-6	-28	(78)	15	4	0	-33		
- PBC	-12	-4	-20	(84)	31	12	57	64		
- PB	-24	-2	-22	(39)	5	4	0	-39		

^aTheory of Planned Behaviour constructs: Intention (I); Attitude (A); Subjective Norm (SN); Perceived Behavioural Control (PBC); and Past Behaviour (PB)

^bLikert item scores rescaled from -3 to 3 and multiplied by number of responses

^cPositive difference (+); negative difference (-); and ties (=). (*) = $p < .05$.

Table C.6: Tests for multi-collinearity, and recommended number of cases, in relation to respondents' preferred strategic (long-term)/water surplus behavioural intention. Tests for multi-collinearity of the predictor variables (i.e. attitude, subjective norm (SN), and perceived behavioural control (PBC)) in relation to the binary response variable (i.e. intention) where 0 = strongly disagree to not sure and 1 = agree to strongly agree with the intention to sell surplus water for the duration of the growing season ($n = 75$)

Measure	Attitude	SN	PBC	VIF	Frequency	
					0	1
- Intention	.356	.332	.181			
- Attitude		.141	.136	1.033	64	11
- SN			.196	1.054	40	35
- PBC				1.053	58	17

Variance Inflation Factor (VIF)

Table C.7: In-season (short-term)/water shortage behavioural intentions (section C of survey)

Behaviour	^b Likert item responses (%) ($n=90$)							Median	IQR	
^a Construct	1	2	3	4	5	6	7	Missing		
Only use your maximum abstraction licence to spread water evenly between all crops										
- I	16	6	32	7	16	9	10	6	3	2
- A	7	17	42	10	11	4	2	7	3	2
- SN	10	2	20	22	24	6	4	11	4	2
- PBC	10	6	17	3	27	13	17	8	5	3
- PB	2	1	3	1	3	1	0	88	3	3
Only use your maximum abstraction licence to irrigate your most valuable crops										
- I	7	0	6	1	33	23	24	6	6	2
- A	3	4	19	11	40	8	8	7	5	2
- SN	3	1	4	12	31	24	12	11	5	1
- PBC	9	4	14	6	29	13	17	8	5	3
- PB	0	1	1	2	4	1	2	88	5	2
Restrict irrigation (i.e. deficit irrigation)										
- I	7	3	22	18	33	4	2	10	4	2
- A	10	18	40	16	7	0	2	8	3	2
- SN	8	4	20	28	17	8	3	12	4	2
- PBC	9	8	17	4	29	11	13	9	5	3
- PB	1	1	2	2	6	0	0	88	4	2
Buy more water to meet crop water requirements										
- I	9	3	8	14	33	14	7	11	5	1
- A	4	2	11	17	36	12	8	10	5	1
- SN	3	0	4	26	36	9	9	13	5	1
- PBC	9	6	13	10	31	7	13	11	5	2
- PB	2	0	1	1	6	0	2	88	5	2

^aTheory of Planned Behaviour constructs: Intention (I); Attitude (A); Subjective Norm (SN); Perceived Behavioural Control (PBC); and Past Behaviour (PB)

^b(I), (PBC), and (PB) = strongly disagree (1) to strongly agree (7); (A) = extremely disadvantageous (1) to extremely advantageous (7); and (SN) = definitely shouldn't (1) to definitely should (7) (4 = not sure for all constructs)

Table C.8: Preferred in-season (short-term)/water shortage behavioural intentions

Behaviour	^b Rescaled Likert item responses							\sum Totals	^c Sign test
^a Construct	-3	-2	-1	(n=)	1	2	3	Total	(-,+,=)
Only use your maximum abstraction licence to irrigate your most valuable crops									
								301	
- I	-18	0	-5	(85)	30	42	66	115	42,13,25*
- A	-9	-8	-17	(84)	36	14	21	37	
- SN	-9	-2	-4	(80)	28	44	33	90	
- PBC	-24	-8	-13	(83)	26	24	45	50	
- PB	0	-2	-1	(11)	4	2	6	9	
Buy more water to meet crop water requirements									
								179	
- I	-24	-6	-7	(80)	30	26	18	37	43,19,18*
- A	-12	-4	-10	(81)	32	22	21	49	
- SN	-9	0	-4	(78)	32	16	24	59	
- PBC	-24	-10	-12	(80)	28	12	36	30	
- PB	-6	0	-1	(11)	5	0	6	4	
Only use your maximum abstraction licence to spread water evenly between all crops									
								-56	
- I	-42	-10	-29	(85)	14	16	27	-24	21,29,31
- A	-18	-30	-38	(84)	10	8	6	-62	
- SN	-27	-4	-18	(80)	22	10	12	-5	
- PBC	-27	-10	-15	(83)	24	24	45	41	
- PB	-6	-2	-3	(11)	3	2	0	-6	
Restrict irrigation (i.e. deficit irrigation)									
								-65	
- I	-18	-6	-20	(81)	30	8	6	0	
- A	-27	-32	-36	(83)	6	0	6	-83	
- SN	-21	-8	-18	(79)	15	14	9	-9	
- PBC	-24	-14	-15	(82)	26	20	36	29	
- PB	-3	-2	-2	(11)	5	0	0	-2	

^aTheory of Planned Behaviour constructs: Intention (I); Attitude (A); Subjective Norm (SN); Perceived Behavioural Control (PBC); and Past Behaviour (PB)

^bLikert item scores rescaled from -3 to 3 and multiplied by number of responses

^cPositive difference (+); negative difference (-); and ties (=). (*) = $p < .05$.

Table C.9: Tests for multi-collinearity, and recommended number of cases, in relation to respondents' preferred in-season (short-term)/water shortage behavioural intention. Tests for multi-collinearity of the predictor variables (i.e. attitude, subjective norm (SN), and perceived behavioural control (PBC)) in relation to the binary response variable (i.e. intention) where 0 = strongly disagree to very much agree and 1 = strongly agree with the intention to only use your maximum abstraction licence to irrigate your most valuable crops ($n = 78$)

Measure	Attitude	SN	PBC	VIF	Frequency	
					0	1
- Intention	.446	.483	.313			
- Attitude		.410	.181	1.219	65	13
- SN			.180	1.219	47	31
- PBC				1.048	51	27
Variance Inflation Factor (VIF)						

Table C.10: Tests for multi-collinearity, and recommended number of cases, in relation to respondents' preferred in-season (short-term)/water shortage behavioural intention after removing an outlier. Tests for multi-collinearity of the predictor variables (i.e. attitude, subjective norm (SN), and perceived behavioural control (PBC)) in relation to the binary response variable (i.e. intention) where 0 = strongly disagree to very much agree and 1 = strongly agree with the intention to only use your maximum abstraction licence to irrigate your most valuable crops ($n = 77$)

Measure	Attitude	SN	PBC	VIF	Frequency	
					0	1
- Intention	.466	.513	.337			
- Attitude		.408	.177	1.216	64	13
- SN			.174	1.215	46	31
- PBC				1.046	50	27
Variance Inflation Factor (VIF)						

Table C.11: In-season (short-term)/water surplus behavioural intentions (section C of survey)

Behaviour	^b Likert item responses (%) ($n=90$)							Median	IQR	
^a Construct	1	2	3	4	5	6	7	Missing		
Sell surplus water to maximise profits										
- I	8	0	14	27	14	16	9	12	4	3
- A	3	0	3	16	40	18	9	11	5	2
- SN	6	1	10	27	26	7	7	18	4	1
- PBC	9	1	18	4	32	7	18	11	5	3
- PB	9	1	19	2	9	1	1	58	3	2
Just use your abstraction licence to meet crop water requirements and leave the remainder of your licence unused										
- I	3	2	16	12	37	11	10	9	5	1
- A	2	2	23	28	28	1	3	12	4	2
- SN	2	4	13	22	30	1	9	18	4	1
- PBC	10	6	13	8	28	6	19	11	5	3
- PB	0	1	1	3	30	3	11	50	5	1
Abstract surplus water for storage										
- I	6	6	23	18	18	6	6	19	4	2
- A	6	2	10	18	36	8	4	17	5	1
- SN	6	1	10	26	26	3	8	21	4	1
- PBC	10	1	17	11	24	4	18	14	5	3
- PB	7	2	19	4	10	1	1	56	3	2

^aTheory of Planned Behaviour constructs: Intention (I); Attitude (A); Subjective Norm (SN); Perceived Behavioural Control (PBC); and Past Behaviour (PB)

^b(I), (PBC), and (PB) = strongly disagree (1) to strongly agree (7); (A) = extremely disadvantageous (1) to extremely advantageous (7); and (SN) = definitely shouldn't (1) to definitely should (7) (4 = not sure for all constructs)

Table C.12: Preferred in-season (short-term)/water surplus behavioural intentions

Behaviour	^b Rescaled Likert item responses							\sum Totals	^c Sign test
^a Construct	-3	-2	-1	(n=)	1	2	3	Total	(-,+,=)
Just use your abstraction licence to meet crop water requirements and leave the remainder of your licence unused									
								182	
- I	-9	-4	-14	(82)	33	20	27	53	37,29,11
- A	-6	-4	-21	(79)	25	2	9	5	
- SN	-6	-8	-12	(74)	27	2	24	27	
- PBC	-27	-10	-12	(80)	25	10	51	37	
- PB	0	-2	-1	(45)	27	6	30	60	
Sell surplus water to maximise profits									
								155	
- I	-21	0	-13	(79)	13	28	24	31	33,16,23*
- A	-9	0	-3	(80)	36	32	24	80	
- SN	-15	-2	-9	(74)	23	12	18	27	
- PBC	-24	-2	-16	(80)	29	12	48	47	
- PB	-24	-2	-17	(38)	8	2	3	-30	
Abstract surplus water for storage									
								58	
- I	-15	-10	-21	(73)	16	10	15	-5	
- A	-15	-4	-9	(75)	32	14	12	30	
- SN	-15	-2	-9	(71)	23	6	21	24	
- PBC	-27	-2	-15	(77)	22	8	48	34	
- PB	-18	-4	-17	(40)	9	2	3	-25	

^aTheory of Planned Behaviour constructs: Intention (I); Attitude (A); Subjective Norm (SN); Perceived Behavioural Control (PBC); and Past Behaviour (PB)

^bLikert item scores rescaled from -3 to 3 and multiplied by number of responses

^cPositive difference (+); negative difference (-); and ties (=). (*) = $p < .05$.

Table C.13: Tests for multi-collinearity, and recommended number of cases, in relation to respondents' preferred in-season (short-term)/water surplus behavioural intention. Tests for multi-collinearity of the predictor variables (i.e. attitude, subjective norm (SN), and perceived behavioural control (PBC)) in relation to the binary response variable (i.e. intention) where 0 = strongly disagree to agree and 1 = very much agree to strongly agree with the intention to just use your abstraction licence to meet crop water requirements and leave the remainder of your licence unused ($n = 72$)

Measure	Attitude	SN	PBC	VIF	Frequency	
					0	1
- Intention	.467	.129	.483			
- Attitude		.158	.003	1.026	46	26
- SN			.097	1.035	38	34
- PBC				1.010	50	22

Variance Inflation Factor (VIF)

Table C.14: Iteration history with regards to respondents' preferred in-season (short-term)/water surplus behavioural intention ($n = 72$)

Iteration	-2log likelihood	Coefficients			
		Constant	Attitude	SN	PBC
1	50.304	-2.205	1.641	.032	1.773
2	42.001	-3.492	2.771	-.102	2.849
3	39.176	-4.677	3.912	-.297	3.941
4	38.243	-5.779	5.006	-.406	5.013
5	37.925	-6.809	6.035	-.429	6.038
6	37.811	-7.816	7.041	-.430	7.044
7	37.769	-8.818	8.043	-.431	8.046
8	37.753	-9.818	9.044	-.431	9.047
9	37.747	-10.819	10.044	-.431	10.047
10	37.745	-11.819	11.044	-.431	11.047
11	37.745	-12.819	12.044	-.431	12.047
12	37.744	-13.819	13.044	-.431	13.047
13	37.744	-14.819	14.044	-.431	14.047
14	37.744	-15.819	15.044	-.431	15.047
15	37.744	-16.819	16.044	-.431	16.047
16	37.744	-17.819	17.044	-.431	17.047
17	37.744	-18.819	18.044	-.431	18.047
18	37.744	-19.819	19.044	-.431	19.047
19	37.744	-20.819	20.044	-.431	20.047
20	37.744	-21.819	21.044	-.431	21.047

Estimation terminated at iteration 20 because maximum iterations had been reached. Final solution was not found

Table C.15: Influential social group (section B and C of survey)

Scenario - ^a Social group	^b Ranking (%) ($n=90$)			
	1	2	3	Missing
Strategic (long-term)/water shortage				
- A	27	49	18	7
- B	58	30	3	9
- C	10	12	66	12
Strategic (long-term)/water surplus				
- A	27	53	13	7
- B	60	29	4	7
- C	10	10	69	11
In-season (short-term)/water shortage				
- A	29	48	13	10
- B	52	31	6	11
- C	11	10	64	14
In-season (short-term)/water surplus				
- A	23	47	14	16
- B	53	23	7	17
- C	9	13	58	20

^aGroups interested in water saving and efficiency (A); Groups interested in crop type and quality (B); and family, friends and neighbours (C)

^bMost influential (1) to least influential (3)

Table C.16: Actual experience of scenario (section B and C of survey)

Scenario	Experienced (%) ($n=90$)		
	Yes	No	Missing
Strategic (long-term)/water shortage			
	9	90	1
Strategic (long-term)/water surplus			
	51	47	2
In-season (short-term)/water shortage			
	12	80	8
In-season (short-term)/water surplus			
	53	36	11

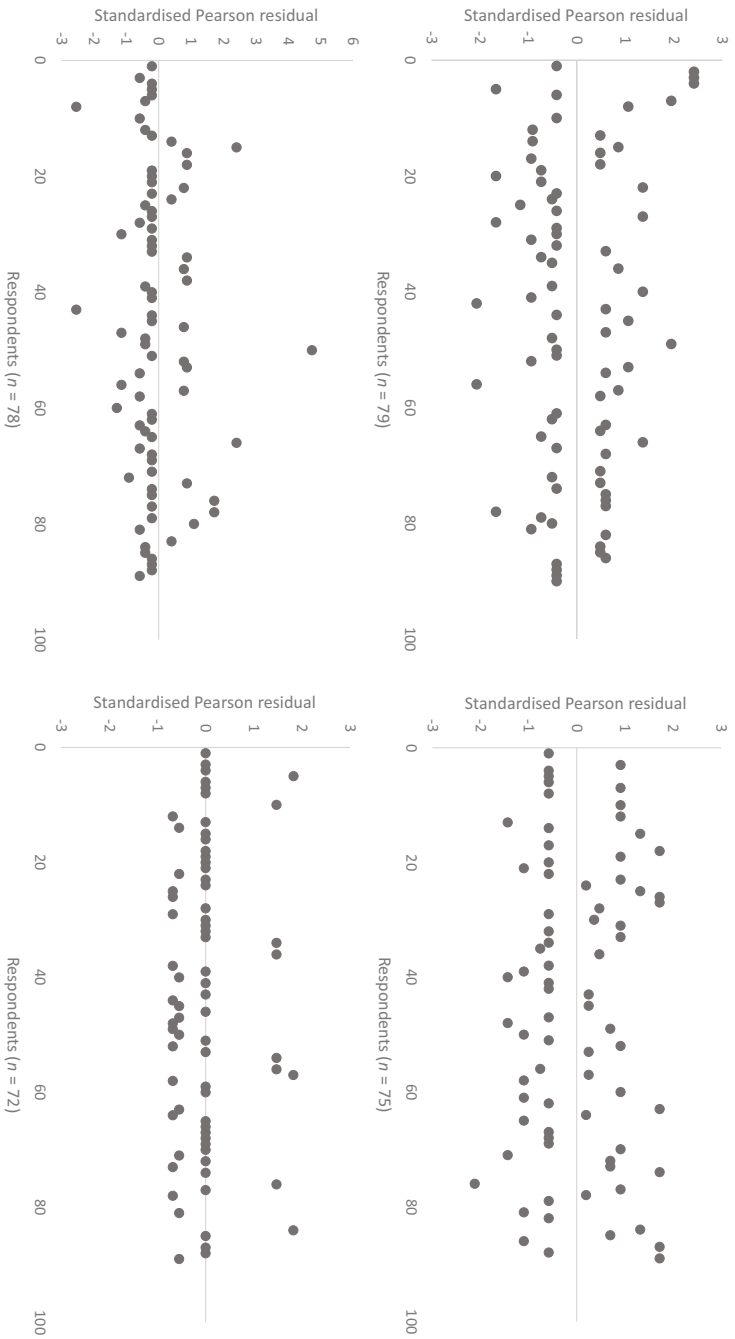


Figure C.1: Scatter plots illustrating standardised Pearson residuals with regards to respondents' preferred behavioural intentions during strategic (long-term) and in-season (short-term) water shortage and surplus scenarios. (Top left) A scatter plot illustrating standardised Pearson residuals with regards to respondents' preferred strategic (long-term)/water shortage behavioural intention. (Top right) A scatter plot illustrating standardised Pearson residuals with regards to respondents' preferred strategic (long-term)/water surplus behavioural intention. (Bottom left) A scatter plot illustrating standardised Pearson residuals with regards to respondents' preferred in-season (short-term)/water shortage behavioural intention. (Bottom right) A scatter plot illustrating standardised Pearson residuals with regards to respondents' preferred in-season (short-term)/water surplus behavioural intention

Appendix D

Summary statistics of farm types

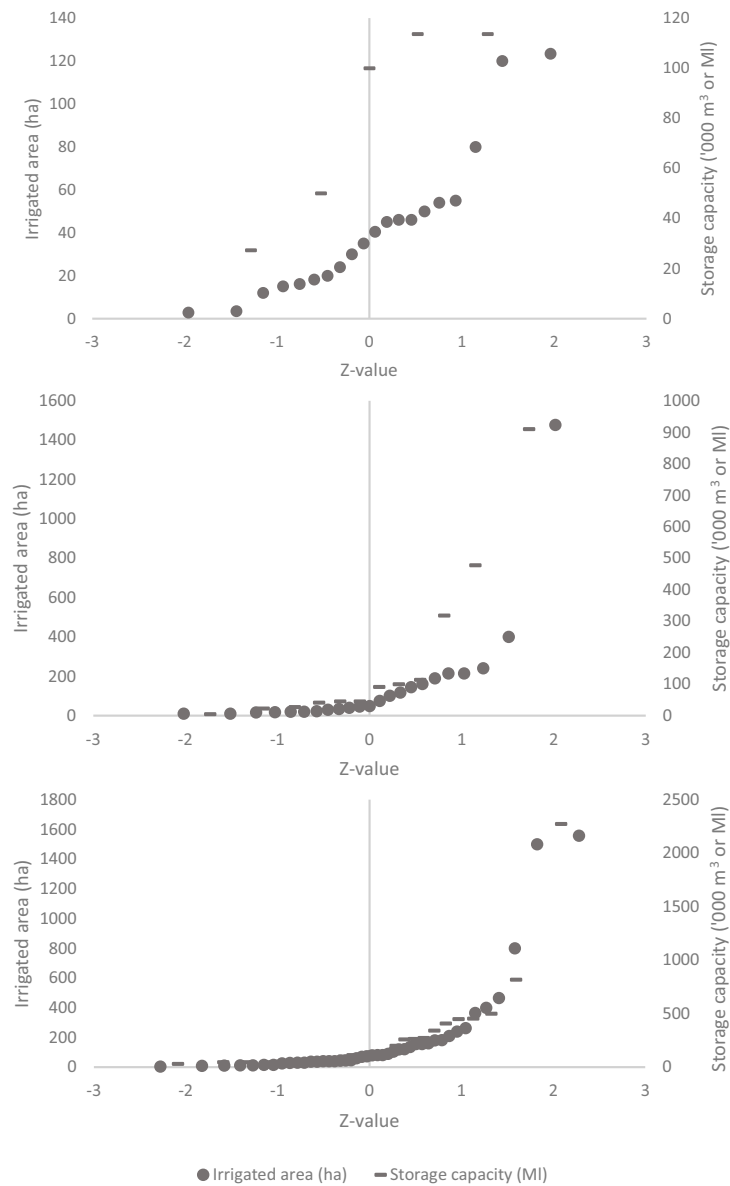


Figure D.1: Normal probability plots illustrating distribution with regards to irrigated area and storage capacity for each farm type. (Top) A normal probability plot illustrating that neither variables were normally distributed for those who preferred LWUO which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for both variables. (Middle) A normal probability plot illustrating that neither variables were normally distributed for those who had no preference which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for both variables. (Bottom) A normal probability plot illustrating that neither variables were normally distributed for those who preferred HWUO which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for both variables

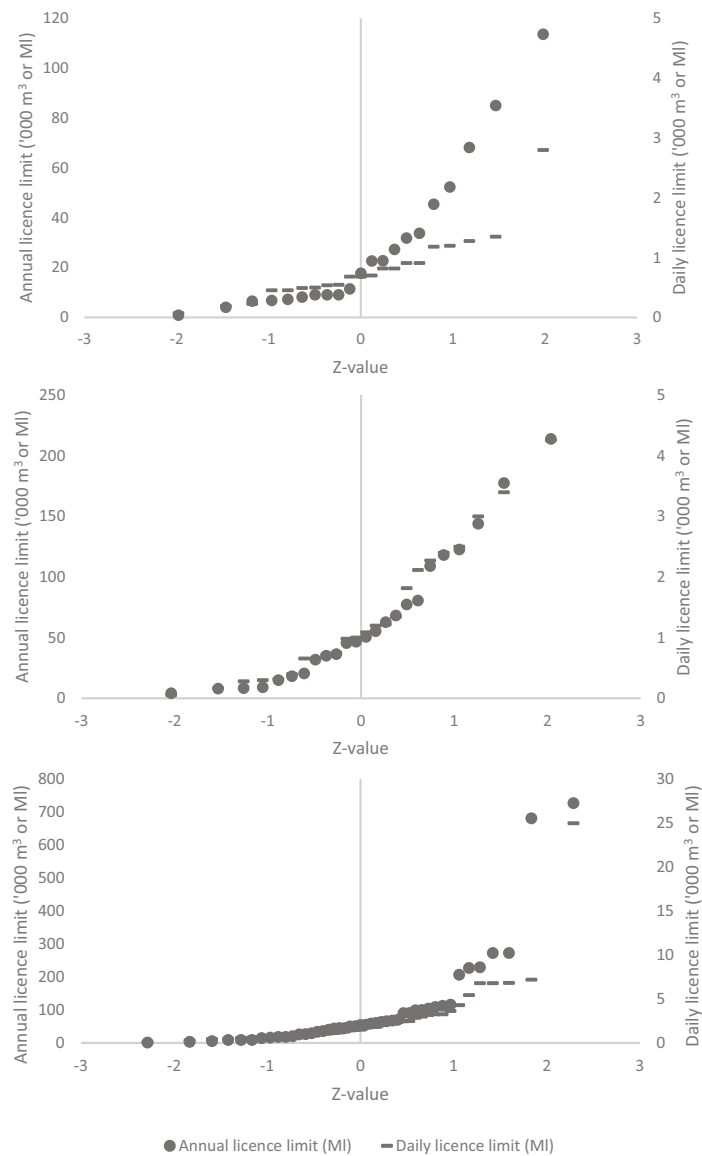


Figure D.2: Normal probability plots illustrating distribution with regards to annual and daily licence limits for each farm type. (Top) A normal probability plot illustrating that neither variables were normally distributed for those who preferred LWUO which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for both variables. (Middle) A normal probability plot illustrating that neither variables were normally distributed for those who had no preference which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for both variables. (Bottom) A normal probability plot illustrating that neither variables were normally distributed for those who preferred HWUO which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for both variables

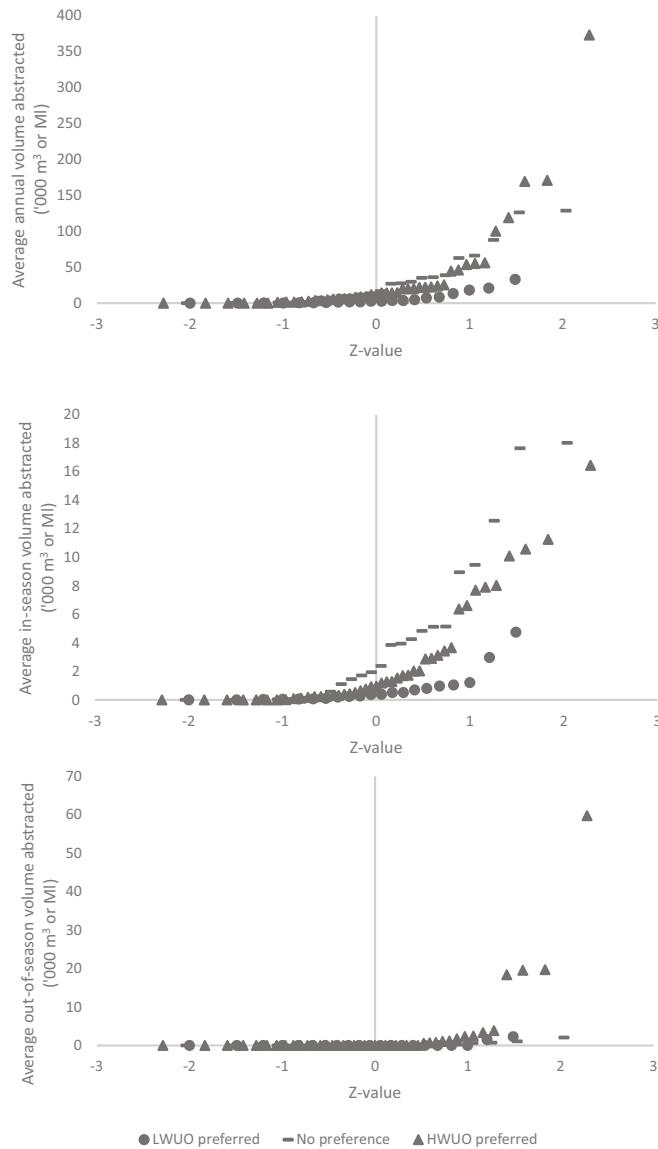


Figure D.3: Normal probability plots illustrating distribution with regards to annual, in-season, and out-of-season abstractions for each farm type. (Top) A normal probability plot illustrating that none of the variables were normally distributed for those who preferred LWUO which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for all variables. (Middle) A normal probability plot illustrating that none of the variables were normally distributed for those who had no preference which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for all variables. (Bottom) A normal probability plot illustrating that none of the variables were normally distributed for those who preferred HWUO which supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for all variables

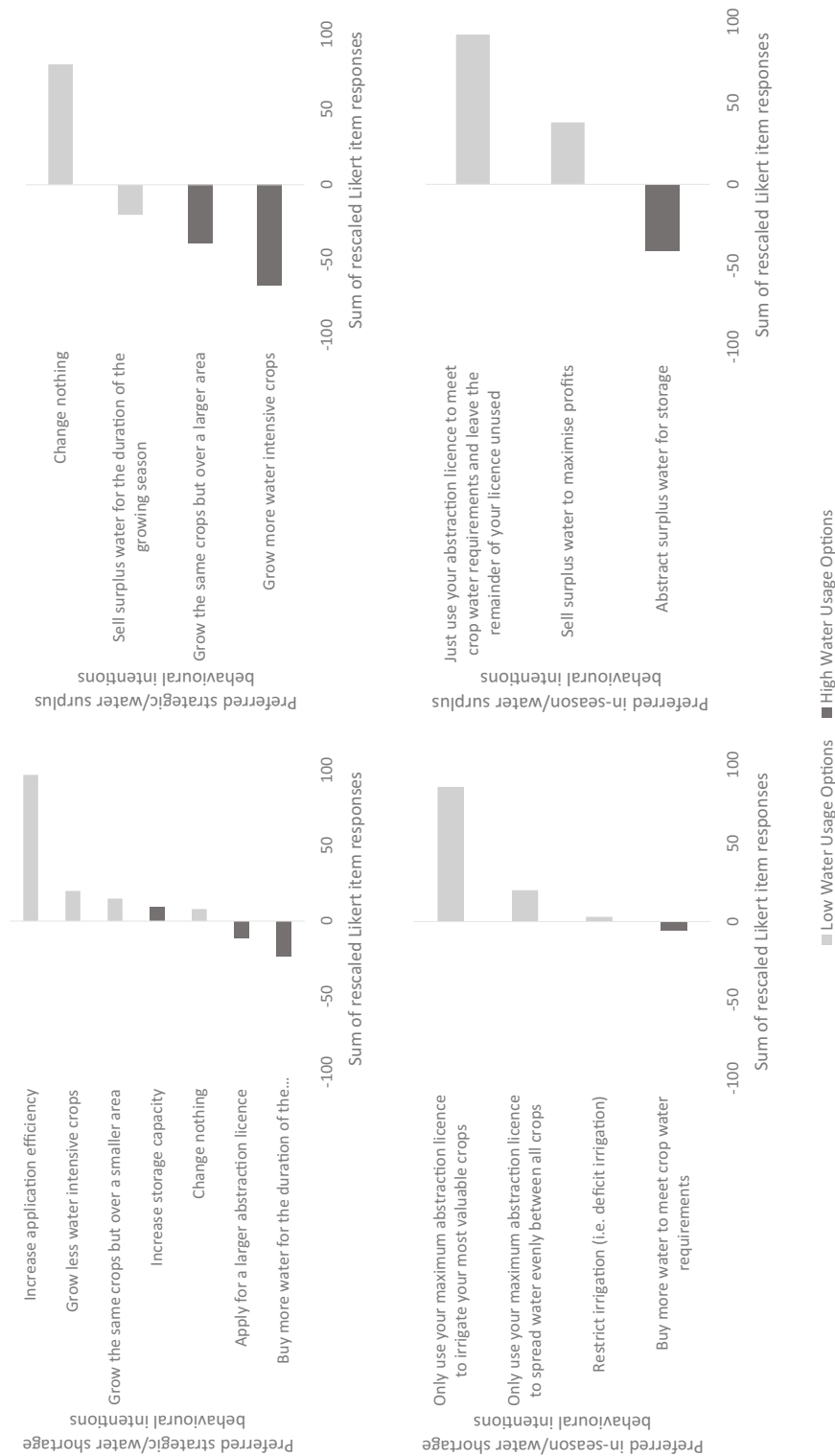


Figure D.4: Bar charts illustrating the preferred behavioural intentions of those who preferred LWUO during strategic (long-term) and in-season (short-term) water shortage and surplus scenarios. (Top left) A bar chart illustrating the preferred strategic (long-term)/water shortage behavioural intentions of those who preferred LWUO. (Top right) A bar chart illustrating the preferred strategic (long-term)/water surplus behavioural intentions of those who preferred LWUO. (Bottom left) A bar chart illustrating the preferred in-season (short-term)/water shortage behavioural intentions of those who preferred LWUO. (Bottom right) A bar chart illustrating the preferred in-season (short-term)/water surplus behavioural intentions of those who preferred LWUO

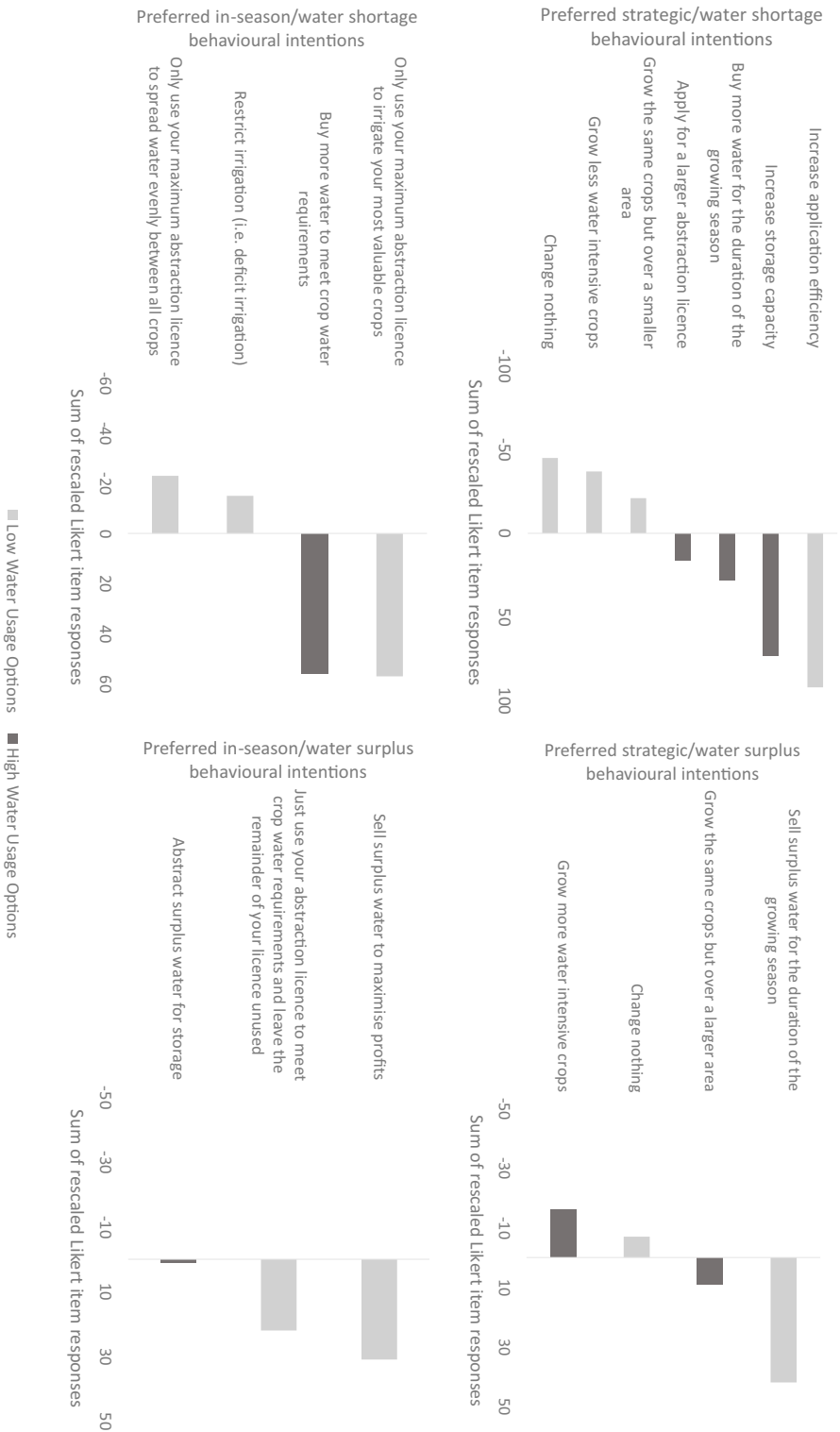


Figure D.5: Bar charts illustrating the preferred behavioural intentions of those who had no preference during strategic (long-term) and in-season (short-term) water shortage and surplus scenarios. (Top right) A bar chart illustrating the preferred strategic (long-term)/water shortage behavioural intentions of those who had no preference. (Top left) A bar chart illustrating the preferred in-season (short-term)/water surplus behavioural intentions of those who had no preference. (Bottom left) A bar chart illustrating the preferred in-season (short-term)/water shortage behavioural intentions of those who had no preference. (Bottom right) A bar chart illustrating the preferred in-season (short-term)/water surplus behavioural intentions of those who had no preference.

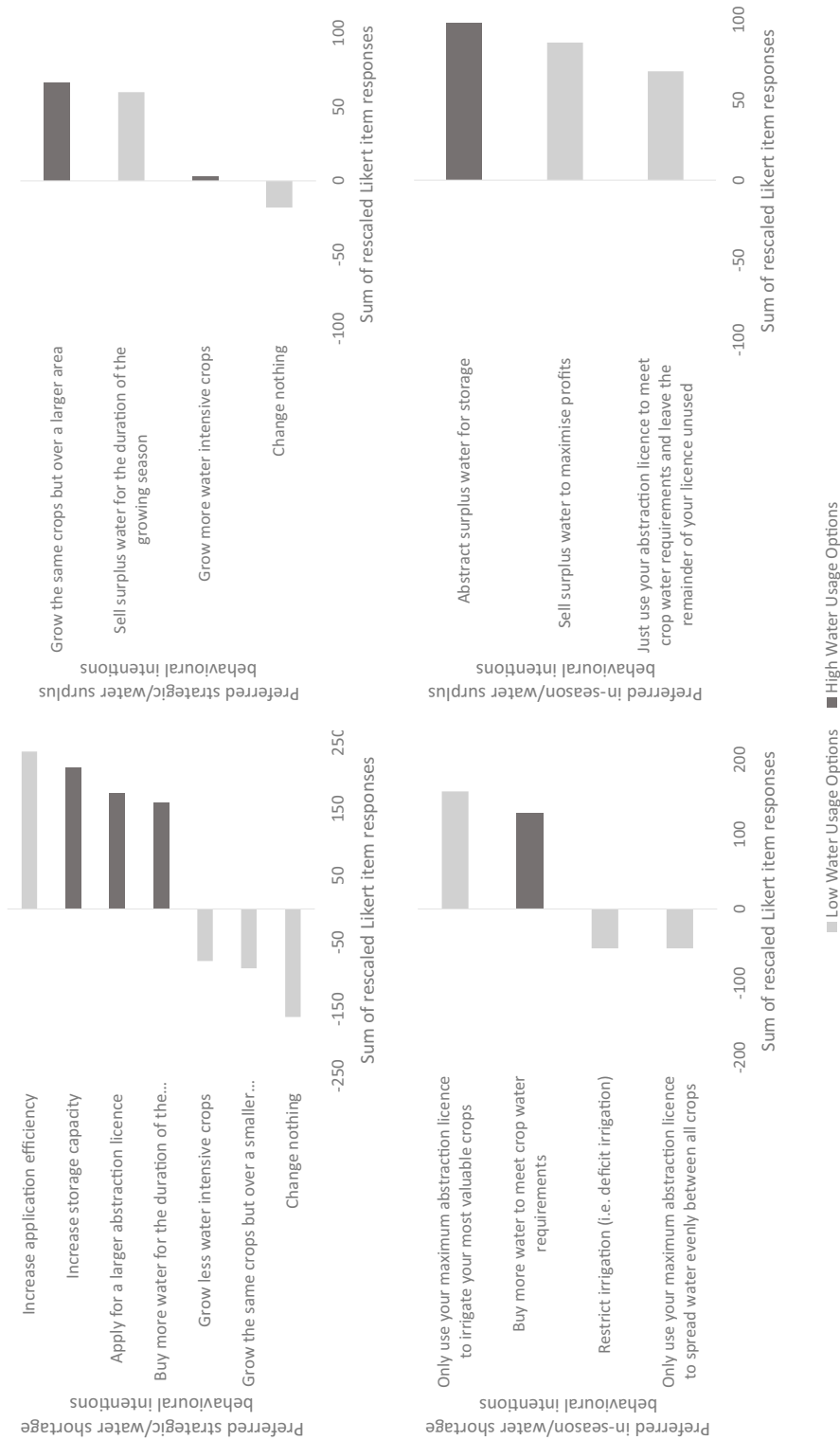


Figure D.6: Bar charts illustrating the preferred behavioural intentions of those who preferred HWUO during strategic (long-term) and in-season (short-term) water shortage and surplus scenarios. (Top left) A bar chart illustrating the preferred strategic (long-term)/water shortage behavioural intentions of those who preferred HWUO. (Top right) A bar chart illustrating the preferred strategic (long-term)/water surplus behavioural intentions of those who preferred HWUO. (Bottom left) A bar chart illustrating the preferred in-season (short-term)/water shortage behavioural intentions of those who preferred HWUO. (Bottom right) A bar chart illustrating the preferred in-season (short-term)/water surplus behavioural intentions of those who preferred HWUO

Appendix E

Model description

E.1 Irrigation Water Need (IWN)

E.1.1 Calculating irrigation water need (IWN)

In addition to soil, air, and sunlight crops need water to grow. Where precipitation is not enough to meet crop water needs alone irrigation is required and in this form is usually referred to as supplemental irrigation. IWN in this study was calculated using the method proposed by Brouwer and Heibloem (1986) as:

$$IWN = ET_c - Pe \quad (\text{E.1})$$

where ET_c equals crop evapotranspiration (i.e. crop water need); and Pe equals effective precipitation (i.e. precipitation which is retained within the root zone).

E.1.2 Calculating crop evapotranspiration (ET_c)

Crop evapotranspiration (ET_c), under optimal conditions, can be affected by: climate (i.e. radiation; air temperature; humidity; and wind speed); crop type (i.e. as different crops, and even varieties of the same crop, have different daily and seasonal crop water demands); and the growth stage of the crop (i.e. as a fully grown crop will require more water than the same crop which has just been planted) (Brouwer and Heibloem, 1986). ET_c in this study was calculated using the method proposed by Allen et al. (1998) as:

$$ET_c = ET_o K_c \quad (\text{E.2})$$

where ET_o equals reference evapotranspiration; and K_c equals a crop coefficient.

E.1.3 Calculating reference evapotranspiration (ET_o)

There are multiple methods for calculating reference evapotranspiration (ET_o) which can broadly be categorised into: water-budget; mass-transfer; combination; radiation; and temperature-based methods (Xu and Singh, 2002). The Food and Agricultural Organisation of the United Nations (FAO) recommend the use of the Penman-Monteith method, a combination-based method, as it has shown to provide relatively accurate results in a variety of locations and climates using a hypothetical grass reference crop with an assumed crop height of 0.12 m, a fixed surface resistance of 70 s m^{-1} , and an albedo of 0.23 (Allen et al., 1998). However, despite the accuracy of this method the main criticism concerns the large data requirements, which for many areas are not recorded, and include: radiation; air temperature; air humidity; and wind speed data.

As an alternative method, when data is limited, the FAO propose the use of the 1985 Hargreaves method, a temperature-based method, which has shown reasonable results with global validity and only requires minimum and maximum air temperature (Allen et al., 1998). In addition, the 1985 Hargreaves method has commonly been used for providing ET_o predictions for weekly or longer periods for use in irrigation scheduling at the farm and regional level (Hargreaves and Allen, 2003). Eight years of measured lysimeter evapotranspiration of a cool season grass (*Alta fescue*), with a crop height of between 0.08 and 0.15 m, at Davis, California, was used as the reference crop (Hargreaves and Samani, 1985). ET_o in this study was calculated using the method proposed by Hargreaves and Samani (1985) as:

$$ET_o = CR_a \left(\frac{T_{max} + T_{min}}{2} + 17.8 \right) (T_{max} - T_{min})^E \quad (\text{E.3})$$

where C and E are parameters of the equation with suggested values of 0.0023 and 0.50 respectively although several studies have indicated that local calibration of these parameters is necessary in order to improve precision (Luo et al., 2014); R_a equals extraterrestrial radiation; and T_{max} and T_{min} equal maximum and minimum temperature ($^{\circ}\text{C}$) respectively.

E.1.4 Calculating extraterrestrial radiation (R_a)

The solar radiation received at the top of the earth's atmosphere on a horizontal surface is called the extraterrestrial (solar) radiation (R_a). If the sun is directly overhead, the angle of incidence is zero and R_a equals $0.0820 \text{ MJ m}^{-2} \text{ min}^{-1}$. Therefore, as seasons change the position of the sun and length of day also change thus R_a is a function of latitude, date, and time (Allen et al., 1998). R_a in this study was calculated using the

method proposed by Allen et al. (1998) as:

$$R_a = \frac{24(60)}{\pi} G_{sc} d_r (\omega_s \sin(\varphi) \sin(\delta) + \cos(\varphi) \cos(\delta) \sin(\omega_s)) \quad (\text{E.4})$$

$$d_r = 1 + 0.033 \cos\left(\frac{2\pi}{365} J\right) \quad (\text{E.5})$$

$$\omega_s = \arccos(-\tan(j) \tan(d)) \quad (\text{E.6})$$

$$j = \frac{\pi}{180} \left(\text{degree} + \frac{\text{minute}}{60} \right) \quad (\text{E.7})$$

$$d = 0.409 \sin\left(\frac{2\pi}{365} J - 1.39\right) \quad (\text{E.8})$$

where G_{sc} equals the solar constant $0.0820 \text{ MJ m}^{-2} \text{ min}^{-1}$; d_r equals the inverse relative distance between the earth and sun (see equation E.5); J equals the day of the year (between 1 and 365 or 366 in a leap year); ω_s equals the sunset hour angle (see equation E.6); j equals latitude in radians (see equation E.7); and d equals solar declination (see equation E.8). Furthermore, to convert R_a into equivalent mm day^{-1} it can be multiplied by 0.408 (a conversion factor equal to the inverse of the latent heat of vaporisation).

Therefore, monthly and annual ET_o values were calculated for the study area from 1973 to 2012 using: regional average minimum and maximum temperature data obtained from the UK Met Office (MetOffice, 2016); the central point of the study area to determine latitude (i.e. $52^{\circ}21' \text{ N}$); and the 15th day of each month to determine day of the year. Furthermore, Figure E.1 (top) illustrates the minimum, average, and maximum ET_o values from 1962 to 1996 presented by Hess (1996) for the months April to September at Silsoe College, Bedfordshire (located within the study area) which were used to calibrate the model. Initial results of the uncalibrated model using the suggested values of 0.0023 and 0.50 for parameters C and E , and temperature data for the same period, resulted in an overestimation of ET_o values. Therefore the parameters were reduced systematically until ET_o values were similar to those presented by Hess (1996) using values 0.0020 and 0.30 for parameters C and E respectively. In addition, Figure E.1 (bottom) illustrates a strong, positive, linear correlation between the average ET_o values presented by Hess (1996) and those calibrated for this study for the period 1962 to 1996 ($r = .99$) which was also statistically significant ($p = <.001$).

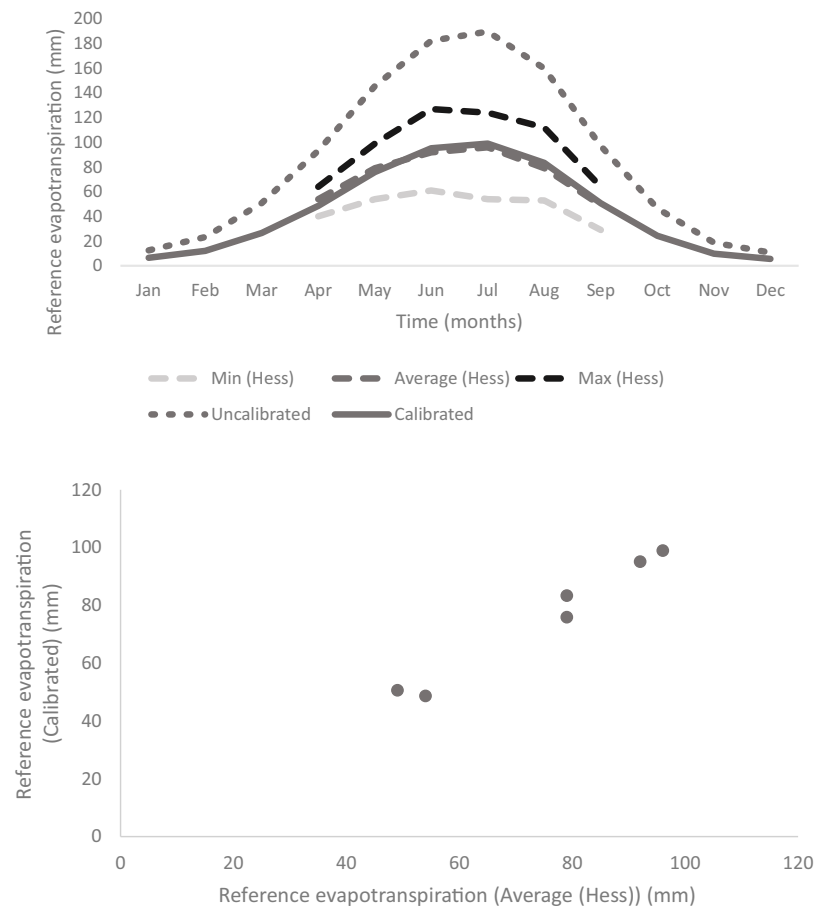


Figure E.1: A line graph and scatter plot indicating the strength of association between monthly reference evapotranspiration calibrated in this study and those presented by Hess (1996) (1962 to 1996). (Top) A line graph indicating calibrated and uncalibrated monthly reference evapotranspiration (ET_o) in relation to those presented by Hess (1996) (1962 to 1996). (Bottom) A scatter plot indicating the strength of association between average ET_o presented by Hess (1996) and calibrated ET_o for this study (1962 to 1996) (Pearson's correlation coefficient $r = .99$, $p = <.001$)

E.1.5 Calculating crop coefficients (K_c)

The relationship between ET_c and ET_o is given by the crop coefficient (K_c). To determine K_c values it is necessary to know: the total growing period for each crop; the various growth stages of each crop; and the K_c values for each crop for each of the growing stages (Brouwer and Heibloem, 1986). Two crops were selected which represented the two most irrigated crop categories in the study area: main crop potatoes (*Solanum tuberosum*) which accounted for 41 % of irrigated crops and vegetables (carrots) (*Daucus carota*) which represented 33 % of irrigated crops. Carrots were selected to represent vegetables similar to the study presented by Knox et al. (1997) which examined IWN at Silsoe College, Bedfordshire. The total growing period, in days, for

Table E.1: Input data for deriving crop coefficient (K_c) values

Measure	Main crop potatoes	Vegetables (i.e. carrots)
Planting date	1 st Apr.	1 st May
Harvest date	15 th Sep.	20 th Sep.
Total growing days	168	143
- Initial stage	40	31
- Crop development stage	61	76
- Mid-season stage	31	18
- Late-season stage	36	18
K_c values		
- Initial stage	0.45	0.45
- Crop development stage	0.75	0.75
- Mid-season stage	1.10	1.00
- Late-season stage	0.85	1.00

each crop is from planting, or transplanting, to harvest. Once the total growing period is known the total number of days is divided into four growing stages: the initial stage (i.e. from planting, or transplanting until the crop covers approximately 10 to 20 % of the ground); the crop development stage (i.e. from the initial stage until full ground cover); the mid-season stage (i.e. from the crop development stage until the crop reaches maturity); and the late-season stage (i.e. from mid-season until harvest). Although K_c values provided by Brouwer and Heibloem (1986) were used in this study local data regarding: planting and harvest dates; length of crop growth stages and K_c values at full cover (i.e. mid-season stage) were derived from Knox et al. (1997) and are presented in Table E.1.

However, the number of days for each crop growing stage does not always equate to the number of days in each month. Therefore, when this occurred K_c in this study was calculated using the method proposed by Brouwer and Heibloem (1986) as:

$$K_c = \left(\frac{n^{stage1}}{n^{month}} * K_c^{stage1} \right) + \left(\frac{n^{stage2}}{n^{month}} * K_c^{stage2} \right) \quad (E.9)$$

where n^{month} equals the number of days in a given month; n^{stage1} and n^{stage2} equal the number of growing days in the given month associated with each growing stage; and K_c^{stage1} and K_c^{stage2} equal the corresponding K_c values for relevant growing stages.

E.1.6 Calculating effective precipitation (Pe)

Only precipitation water which is retained within the root zone can be utilised by crops (i.e. commonly referred to as effective precipitation (Pe)) whilst the remainder

is lost through: evaporation; deep percolation; or surface run-off. The main factors which determine the amount of precipitation which is retained in the root zone include: climate; soil texture; soil structure; and depth of the root zone. Although more complex methods can be used to calculate Pe the method used in this study was that proposed by Brouwer and Heibloem (1986) as:

$$Pe = 0.8P - 25 \quad \text{if } P > 75 \quad (\text{E.10})$$

$$Pe = 0.6P - 10 \quad \text{if } P < 75 \quad (\text{E.11})$$

where P equals precipitation (mm month^{-1}).

Therefore, monthly and annual IWN were calculated for main crop potatoes and vegetables (i.e. carrots) from 1973 to 2012. Furthermore, 1983 and 1988 were selected as the dry and wet years used as the climate scenarios for this study based on total annual IWN (i.e. combined IWN of the two crops) which was equalled or exceeded 20 % and 80 % of years respectively which is a method commonly adopted for these types of studies (Knox et al., 1997) (see Figure E.2).

Moreover, Table E.2 presents IWN for each crop, and the total, during both the dry and wet years. Interestingly, although total IWN for both crops is greatest during the dry year (i.e. 138 Ml and 119 Ml respectively) compared to the wet year (i.e. 64 Ml and 48 Ml respectively) irrigation is still required for a longer period of time during the wet year for both crops. In particular, main crop potatoes required supplemental irrigation from June to August during the dry year but from April to September, excluding July, during the wet year. Vegetables (i.e. carrots) also required supplemental irrigation from June to August during the dry year but from June to September, excluding July, during the wet year.

In addition, Knox et al (1997) reported IWN of 235 mm and 164 mm for main crop potatoes and vegetables (i.e. carrots) respectively, during a design dry year based on data obtained between 1973 and 1992, on medium available water capacity soil whilst the method presented in this study calculated IWN of 251 mm and 195 mm, on average, using temperature and precipitation data for the same period. However, the estimates presented by Knox et al. (1997) were based on an irrigation water requirement model using: different climate data; a two layer soil water balance model to estimate soil water storage; and the Penman-Monteith method to calculate ET_o . Therefore, the dry and wet years used as the two climate scenarios for this study represent reasonable IWN estimates.

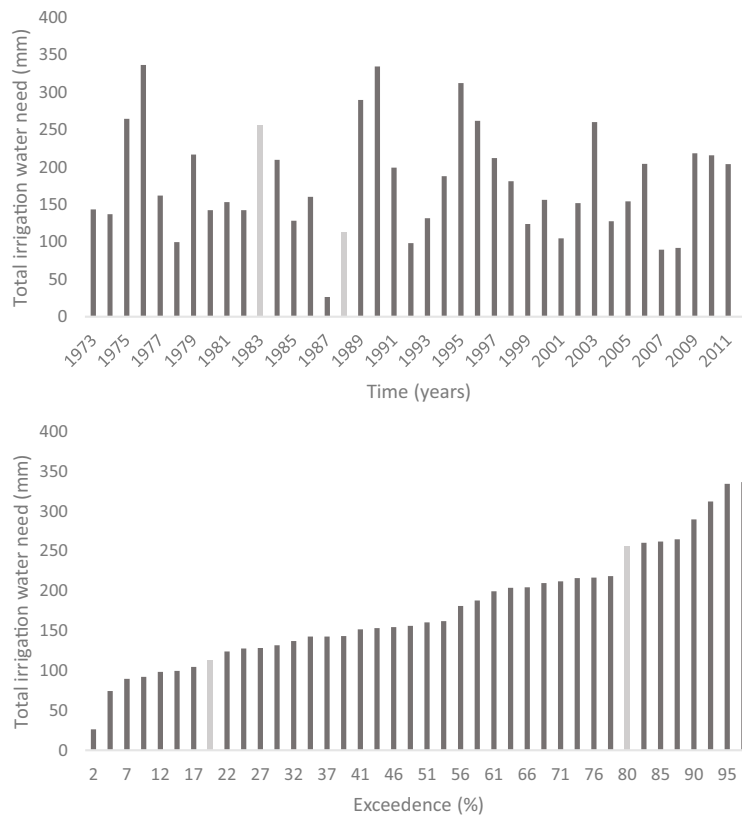


Figure E.2: Bar charts illustrating total annual irrigation water need (IWN) of main crop potatoes and vegetables (i.e. carrots) for the study area from 1973 to 2012 highlighting dry and wet years (i.e. 1983 and 1988 respectively). (Top) A bar chart illustrating total annual irrigation water need (IWN) of main crop potatoes and vegetables (i.e. carrots) for the study area from 1973 to 2012 highlighting dry and wet years (i.e. 1983 and 1988 respectively). (Bottom) A bar chart illustrating the percentage of years which exceeded particular levels of total annual IWN of main crop potatoes and vegetables (i.e. carrots) for the study area, for the period 1973 to 2012, highlighting dry and wet years (i.e. 1983 and 1988 respectively)

Table E.2: Irrigation water needs (IWN) during dry and wet years (i.e. 1983 and 1988 respectively)

Month	Pot _{-dry} (mm)	Veg _{-dry} (mm)	Total _{-dry} (mm)	Pot _{-wet} (mm)	Veg _{-wet} (mm)	Total _{-wet} (mm)
- Jan	0	0	0	0	0	0
- Feb	0	0	0	0	0	0
- Mar	0	0	0	0	0	0
- Apr	0	0	0	3	0	3
- May	0	0	0	9	0	9
- Jun	33	33	66	18	18	37
- Jul	60	43	103	0	0	0
- Aug	45	42	87	28	25	53
- Sep	0	0	0	7	4	11
- Oct	0	0	0	0	0	0
- Nov	0	0	0	0	0	0
- Dec	0	0	0	0	0	0
- Annual	138	119	256	64	48	113

E.2 Input data

Table E.3: Model input data with regards to farm type parameters

Variable	^a LWUO preferred	No preference	^b HWUO preferred
Farm type	1	2	3
Number of farmers	192	226	418
Median irrigated area (m ²)	380,000	490,000	760,000
- 75th percentile (m ²)	510,000	1,760,000	1,650,000
- 25th percentile (m ²)	180,000	210,000	350,000
Main crop potato cover (%)	.63	.49	.57
Vegetable cover (%)	.37	.51	.43
Storage availability (%)	.24	.5	.6
Storage only licence (%)	.0	.16	.22
Median storage capacity (m ³)	100,000	68,000	136,000
- 75th percentile (m ³)	114,000	165,000	307,000
- 25th percentile (m ³)	50,000	38,000	61,000
Median annual licence limit (m ³)	17,727	48,595	54,500
- 75th percentile (m ³)	33,773	87,730	100,000
- 25th percentile (m ³)	8,190	19,884	25,846

^a Low water usage options (LWUO)

^b High water usage options (HWUO)

Table E.4: Model input data with regards to Assessment Point (AP) names, and low (Q95) and high (Q30) flow conditions (m^3/s)

Assessment Point (AP) name	Low flow condition (Q95) (m^3/s)	High flow condition (Q30) (m^3/s)
C1AP01	0.24	0.89
C1AP02	0.09	0.32
C1AP04	0.25	0.94
C1AP06	0.14	0.54
C1AP07	0.14	0.53
C1AP08	0.35	1.33
C1AP09	0.56	2.11
C1AP10	0.26	0.99
C1AP00	0.16	1.27
C2AP00	0.68	5.43
C3AP02	0.01	0.32
C3AP03	0.25	1.05
C3AP04	0.26	1.31
C3AP07	0.11	0.36
C3AP08	0.02	0.33
C3AP09	0.44	1.47
C3AP11	0.10	0.82
C3AP12	0.09	0.54
C3AP13	0.50	2.31
C3AP17	2.38	19.06
C4AP01	2.90	20.54
C4AP03	2.45	17.37
C4AP05	2.37	14.59
C4AP06	1.09	3.33
C4AP12	0.46	5.53
C4AP14	0.55	2.59

Table E.5: Model input data during a dry year (1983) with regards to Irrigation Water Needs (IWN) of main crop potatoes and vegetables (i.e. carrots) (mm), and river discharge for each Assessment Point (AP) area (m³/s)

Variable	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec
IWN _{potatoes}	0	0	0	0	0	33	60	45	0	0	0	0
IWN _{vegetables}	0	0	0	0	0	33	43	42	0	0	0	0
C1AP01 _{discharge}	1.14	1.25	1.10	1.20	1.36	1.14	0.75	0.52	0.55	0.54	0.48	0.54
C1AP02 _{discharge}	0.41	0.45	0.39	0.43	0.49	0.41	0.27	0.19	0.20	0.19	0.17	0.19
C1AP04 _{discharge}	1.21	1.32	1.16	1.27	1.44	1.21	0.79	0.55	0.59	0.57	0.51	0.57
C1AP06 _{discharge}	0.69	0.75	0.66	0.72	0.82	0.69	0.45	0.31	0.33	0.32	0.29	0.33
C1AP07 _{discharge}	0.67	0.74	0.65	0.71	0.80	0.68	0.44	0.31	0.33	0.32	0.29	0.32
C1AP08 _{discharge}	1.69	1.86	1.63	1.78	2.02	1.70	1.11	0.77	0.82	0.80	0.72	0.81
C1AP09 _{discharge}	2.70	2.96	2.60	2.85	3.22	2.71	1.77	1.24	1.31	1.27	1.14	1.28
C1AP10 _{discharge}	1.26	1.38	1.22	1.33	1.50	1.27	0.83	0.58	0.61	0.60	0.53	0.60
C1AP00 _{discharge}	1.86	2.17	1.40	2.40	2.32	1.40	0.55	0.33	0.46	0.49	0.58	0.91
C2AP00 _{discharge}	7.92	9.27	5.98	10.26	9.91	5.97	2.35	1.39	1.96	2.07	2.49	3.87
C3AP02 _{discharge}	0.41	0.71	0.36	0.92	0.62	0.40	0.20	0.13	0.10	0.08	0.07	0.09
C3AP03 _{discharge}	1.52	1.77	1.29	2.23	1.68	1.30	0.81	0.70	0.58	0.58	0.64	0.77
C3AP04 _{discharge}	1.95	2.22	1.61	3.11	2.38	1.88	0.99	0.64	0.58	0.54	0.62	0.91
C3AP07 _{discharge}	0.54	0.60	0.49	0.64	0.69	0.58	0.41	0.33	0.30	0.27	0.30	0.34
C3AP08 _{discharge}	0.60	0.75	0.48	0.93	0.86	0.67	0.36	0.21	0.13	0.11	0.11	0.20
C3AP09 _{discharge}	1.81	2.43	1.67	2.83	3.48	1.88	1.29	1.04	0.97	0.92	0.96	1.16
C3AP11 _{discharge}	1.00	1.49	0.80	1.97	1.82	0.80	0.46	0.33	0.32	0.31	0.34	0.45
C3AP12 _{discharge}	0.72	0.91	0.60	1.14	1.07	0.54	0.36	0.25	0.23	0.24	0.27	0.38
C3AP13 _{discharge}	3.06	3.56	2.50	3.50	3.13	1.69	0.88	0.52	0.61	0.68	0.91	1.60
C3AP17 _{discharge}	27.82	32.56	21.01	36.02	34.79	20.95	8.24	4.90	6.87	7.26	8.75	13.60
C4AP01 _{discharge}	20.89	22.49	15.50	38.69	40.52	18.89	6.82	3.90	5.64	5.18	5.31	13.78
C4AP03 _{discharge}	17.67	19.03	13.12	32.73	34.27	15.98	5.77	3.30	4.77	4.38	4.49	11.66
C4AP05 _{discharge}	14.92	15.21	11.91	24.42	27.80	11.49	4.60	2.99	3.72	3.81	4.17	9.66
C4AP06 _{discharge}	3.96	4.18	3.44	6.52	5.27	4.15	2.10	1.70	1.99	1.90	2.05	2.76
C4AP12 _{discharge}	7.65	6.59	5.76	11.99	14.38	5.07	1.98	1.12	1.34	1.45	1.60	4.63
C4AP14 _{discharge}	3.02	2.96	2.48	4.74	4.42	2.17	1.11	0.81	0.90	1.06	1.15	2.41

Table E.6: Model input data during a wet year (1988) with regards to Irrigation Water Needs (IWN) of main crop potatoes and vegetables (i.e. carrots) (mm), and river discharge for each Assessment Point (AP) area (m³/s)

Variable	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec
IWN _{potatoes}	0	0	0	3	9	18	0	28	7	0	0	0
IWN _{vegetables}	0	0	0	0	0	18	0	25	4	0	0	0
C1AP01 discharge	1.73	1.88	1.85	1.49	1.20	0.86	0.77	0.57	0.57	0.55	0.52	0.62
C1AP02 discharge	0.62	0.67	0.66	0.54	0.43	0.31	0.28	0.21	0.20	0.20	0.19	0.22
C1AP04 discharge	1.83	1.99	1.95	1.58	1.27	0.91	0.81	0.61	0.60	0.58	0.55	0.66
C1AP06 discharge	1.04	1.13	1.11	0.90	0.72	0.52	0.46	0.34	0.34	0.33	0.31	0.37
C1AP07 discharge	1.02	1.11	1.09	0.88	0.71	0.51	0.45	0.34	0.34	0.33	0.31	0.37
C1AP08 discharge	2.57	2.79	2.74	2.22	1.78	1.28	1.14	0.85	0.85	0.82	0.77	0.92
C1AP09 discharge	4.09	4.45	4.37	3.54	2.84	2.05	1.82	1.36	1.35	1.31	1.23	1.47
C1AP10 discharge	1.91	2.08	2.04	1.65	1.33	0.96	0.85	0.63	0.63	0.61	0.58	0.69
C1AP00 discharge	3.80	2.89	2.79	1.76	1.08	0.78	0.90	0.50	0.56	0.63	0.63	1.06
C2AP00 discharge	16.20	12.33	11.91	7.52	4.62	3.31	3.83	2.13	2.37	2.71	2.70	4.53
C3AP02 discharge	0.46	0.47	0.39	0.32	0.25	0.16	0.11	0.09	0.07	0.16	0.27	0.34
C3AP03 discharge	3.59	2.40	2.30	1.54	1.22	1.01	0.99	0.68	0.64	0.74	0.69	0.78
C3AP04 discharge	4.64	3.51	2.75	1.90	1.45	1.06	0.89	0.60	0.66	0.65	0.62	0.98
C3AP07 discharge	0.78	0.66	0.69	0.55	0.45	0.38	0.37	0.27	0.31	0.28	0.28	0.31
C3AP08 discharge	1.37	0.91	1.05	0.64	0.46	0.33	0.24	0.15	0.10	0.10	0.10	0.17
C3AP09 discharge	4.50	2.91	3.32	2.33	1.85	1.45	1.39	1.03	0.99	1.10	1.02	1.16
C3AP11 discharge	3.19	1.83	2.19	1.27	0.88	0.61	0.57	0.37	0.35	0.64	0.44	0.66
C3AP12 discharge	1.98	1.15	1.51	0.90	0.66	0.46	0.44	0.29	0.26	0.36	0.34	0.42
C3AP13 discharge	5.48	5.08	5.51	3.40	2.25	1.48	1.69	1.02	1.04	2.11	1.65	2.44
C3AP17 discharge	56.86	43.30	41.81	26.42	16.21	11.63	13.43	7.49	8.33	9.50	9.48	15.91
C4AP01 discharge	60.76	42.31	33.21	15.67	11.18	6.33	11.38	4.67	7.30	7.91	6.92	14.73
C4AP03 discharge	51.40	35.79	28.09	13.26	9.46	5.35	9.63	3.95	6.17	6.69	5.85	12.46
C4AP05 discharge	42.08	26.26	21.06	9.13	6.69	4.44	7.93	3.53	5.31	5.82	4.72	9.77
C4AP06 discharge	9.28	6.15	5.61	3.97	3.62	2.60	2.91	1.93	2.28	2.64	2.34	2.76
C4AP12 discharge	15.92	9.28	6.75	1.84	1.17	0.84	1.88	0.69	0.92	0.91	1.13	2.03
C4AP14 discharge	7.04	3.88	3.09	1.51	1.33	1.06	1.54	0.77	1.11	1.54	1.21	1.70

E.3 Submodels

Submodels presented below relate to those briefly described in Figure 6.2.

E.3.1 run-monthly-parameters

The first submodel is an observer procedure which sets the global environment state variables including: IWN of main crop potatoes; IWN of vegetables (i.e. carrots); and AP river discharges. However, these variables vary depending on the climate scenario being simulated (i.e. a dry or wet year). Nonetheless, the input data are converted to m^3 per month, and therefore the global environment state variables were set as:

$$IWN_{pot}(m^3) = IWN_{pot}(mm)/1000 \quad (E.12)$$

$$IWN_{veg}(m^3) = IWN_{veg}(mm)/1000 \quad (E.13)$$

$$AP_{discharge}(m^3) = AP_{discharge}(m^3/s) * 60 * 60 * 24 * 30 \quad (E.14)$$

where IWN_{pot} and IWN_{veg} equal IWN of main crop potatoes and vegetables (i.e. carrots) respectively; and $AP_{discharge}$ equals the river discharge of an AP area.

E.3.2 calculate-farm-water-requirements

The second submodel is a farmer procedure which sets IWN of main crop potatoes, vegetables (i.e. carrots), and the total IWN of the two combined for each farmer (i.e. farm water requirements). These were calculated as:

$$FarmIWN_{pot}(m^3) = IWN_{pot}(m^3) * (IA(m^2) * CC_{pot}(\%)) \quad (E.15)$$

$$FarmIWN_{veg}(m^3) = IWN_{veg}(m^3) * (IA(m^2) * CC_{veg}(\%)) \quad (E.16)$$

$$FarmIWN_{total}(m^3) = FarmIWN_{pot}(m^3) + FarmIWN_{veg}(m^3) \quad (E.17)$$

where $FarmIWN_{pot}$ and $FarmIWN_{veg}$ equal farm IWN of main crop potatoes and vegetables (i.e. carrots) respectively; IA equals irrigated area of the farm; CC_{pot} and CC_{veg} equal the percentage crop cover of main crop potatoes and vegetables respectively; and $FarmIWN_{total}$ equals the total IWN of the farm.

E.3.3 calculate-ap-water-availability

The third submodel involves both grid and farmer procedures. First, water availability, and low (Q95) and high (Q30) flow conditions converted to m^3 , were set for each AP area (i.e. landscape state variables) as:

$$AP_{water-available}(m^3) = AP_{discharge}(m^3) \quad (E.18)$$

$$AP_{low-flow}(m^3) = AP_{low-flow}(m^3/s) * 60 * 60 * 24 * 30 \quad (E.19)$$

$$AP_{high-flow}(m^3) = AP_{high-flow}(m^3/s) * 60 * 60 * 24 * 30 \quad (E.20)$$

where $AP_{water-available}$ equals water availability in an AP area; $AP_{low-flow}$ and $AP_{high-flow}$ equal the low and high flow conditions of an AP area respectively.

Secondly, each farmer then calculates how much water is available in the AP in which they are located. Under the current water allocation system, water availability was set as equation E.18, as no low flow conditions are enforced by the regulatory authority, and therefore all water was available. In reality, some licences have HoF conditions attached to them but as this data was unavailable, HoF were not included in the model. However, under the proposed water allocation systems, low flow conditions are enforced and therefore water availability was set as:

$$AP_{water-available}(m^3) = AP_{water-available}(m^3) - AP_{low-flow}(m^3) \quad (E.21)$$

E.3.4 calculate-farm-water-availability

The fourth submodel is a farmer procedure which calculates farmers' monthly licensed water available, and storage water available (i.e. their two sources of water). However, monthly licensed water available depends on the policy scenario being simulated as seasonal restrictions on licences are only enforced under the current system. Therefore, monthly licensed water available for each licence type was set as:

$$Licence_{direct}(m^3) = Licence_{annual}(m^3)/7 \quad (E.22)$$

$$Licence_{storage}(m^3) = Licence_{annual}(m^3)/5 \quad (E.23)$$

$$Licence_{directandstorage}(m^3) = Licence_{annual}(m^3)/12 \quad (E.24)$$

where $Licence_{direct}$ equals monthly licensed water available for farmers with direct licences (i.e. 7 months of the year during in-season); $Licence_{annual}$ equals farmers annual licence limit derived during model initialisation; $Licence_{storage}$ equals monthly licensed water available for farmers with storage only licences (i.e. 5 months of the year during out-of-season); and $Licence_{directandstorage}$ equals monthly licensed water available for farmers with direct and storage licences (i.e. 12 months of the year).

Therefore, under the current water allocation system, if water was available within an AP area, farmers' monthly licensed water availability was set depending on their licence type. If water was unavailable in the AP area, then no farmers' would have monthly licensed water available. However, under the basic and enhanced water allocation systems, seasonal restrictions on licences are removed, and therefore, in the model, all licences became direct and storage, as farmers could abstract for 12 months of the year, as long as water was available within the AP area. Furthermore, under the enhanced water allocation system, monthly licensed water availability was based on a share of available water within an AP area. Therefore, during the first month of the year a percentage was calculated for each farmer as:

$$Licence_{share}(\%) = Licence_{directandstorage}(m^3) / AP_{water-available-proposed}(m^3) \quad (E.25)$$

where $Licence_{share}(\%)$ equals the percentage used in subsequent months to derive monthly licensed water availability. Therefore, in subsequent months, monthly licensed water availability was set as:

$$Licence_{share}(m^3) = Licence_{share}(\%) * AP_{water-available-proposed}(m^3) \quad (E.26)$$

where $Licence_{share}$ equals monthly water available for farmers under the proposed enhanced water allocation system.

Overall, licensed water available depended on the policy scenario being simulated but was set as one of the above for each farmer and is referred to simply as $Licence_{available}$ for the remainder of the submodels. Furthermore, storage water available was set as:

$$Storage_{available}(m^3) = Storage_{annual}(m^3) \quad (E.27)$$

where $Storage_{available}$ equals farmers' storage water available; and $Storage_{annual}$ equals farmers storage capacity derived during model initialisation.

E.3.5 calculate-farm-water-abstractions

The fifth submodel is a more complex farmer procedure which calculates farmers' monthly licensed abstractions, and storage water balance (i.e. how much they abstract to refill and how much they abstract to satisfy IWN) (see Figure 6.3). Furthermore, the procedures involved have been discussed in Section 6.2.1 and are not repeated here. Nonetheless, this section provides the main calculations involved. If a farmer had monthly licensed water available, and it was greater than their IWN, then licence abstracted for IWN was set as:

$$Licence_{abstracted\ for\ IWN}(m^3) = FarmIWN_{total}(m^3) \quad (E.28)$$

however, if a farmer had monthly licensed water available, and it was less than their IWN, then licence abstracted for IWN was set as:

$$Licence_{abstracted\ for\ IWN}(m^3) = Licence_{available}(m^3) \quad (E.29)$$

where $Licence_{abstracted\ for\ IWN}$ equals the volume of monthly licensed water available abstracted to satisfy IWN.

After licensed water available has been abstracted for farm IWN, farmers calculate whether they have surplus water available for storage. This is set as:

$$Licence_{available\ for\ storage}(m^3) = Licence_{available}(m^3) - Licence_{abstracted\ for\ IWN}(m^3) \quad (E.30)$$

where $Licence_{available\ for\ storage}$ equals the remainder of the monthly licensed volume available which was not used to satisfy IWN. Farmers also calculate whether they require to use their storage to satisfy farm IWN, if their monthly licensed volume available was not enough. If a farmer has storage water available, and it was greater than the shortage required to satisfy farm IWN, then:

$$Storage_{abstracted\ for\ IWN}(m^3) = FarmIWN_{total}(m^3) - Licence_{abstracted\ for\ IWN}(m^3) \quad (E.31)$$

however, if a farmer has storage water available, and it was less than the shortage required to satisfy farm IWN, then:

$$Storage_{abstracted\ for\ IWN}(m^3) = Storage_{available}(m^3) \quad (E.32)$$

where $Storage_{abstracted\ for\ IWN}$ equals the volume of storage water available abstracted to satisfy IWN.

Farmers then record how much of their storage water they have remaining, and the how much they require to replenish to full capacity. These are set as:

$$Storage_{limitnextmonth}(m^3) = Storage_{limitnextmonth}(m^3) - Storage_{abstractedforIWN}(m^3) \quad (E.33)$$

and;

$$Storage_{deficitnextmonth}(m^3) = Storage_{annual}(m^3) - Storage_{limitnextmonth}(m^3) \quad (E.34)$$

where $Storage_{limitnextmonth}$ equals how much storage water a farmer has remaining; and $Storage_{deficitnextmonth}$ equals how much water a farmer requires to replenish their farm storage reservoir. Therefore, if a storage water deficit exists, and licence available for storage was less than the deficit, then:

$$Licence_{abstractedforstorage}(m^3) = Licence_{availableforstorage}(m^3) \quad (E.35)$$

or where licence available for storage was greater than the deficit, then:

$$Licence_{abstractedforstorage}(m^3) = Storage_{deficitnextmonth}(m^3) \quad (E.36)$$

where $Licence_{abstractedforstorage}$ equals the volume of monthly licensed water available which was abstracted to replenish a farm storage reservoir. In addition, if licensed water is abstracted for storage then the storage limit next month is updated accordingly:

$$Storage_{limitnextmonth}(m^3) = Storage_{limitnextmonth}(m^3) + Licence_{abstractedforstorage}(m^3) \quad (E.37)$$

Furthermore, farmers' also record how much of their annual and monthly licensed water they have available for the following month. These are set as:

$$Licence_{annuallimitnextmonth}(m^3) = Licence_{annuallimitnextmonth}(m^3) - Licence_{abstractedforIWN}(m^3) - Licence_{abstractedforstorage}(m^3) \quad (E.38)$$

and if licence annual limit next month is greater than their licence monthly limit next month, then:

$$Licence_{monthlylimitnextmonth}(m^3) = Licence_{annuallimitnextmonth}(m^3) \quad (E.39)$$

however, if it less, then:

$$Licence_{monthlylimitnextmonth}(m^3) = Licence_{monthlylimitnextmonth}(m^3) \quad (E.40)$$

where $Licence_{annuallimitnextmonth}$ equals annual licensed volume remaining which was set to annual licence limit during model initialisation; and $Licence_{annuallimitnextmonth}$ equals the monthly licensed volume remaining depending on the volume of their annual licence remaining. Lastly, farmers calculate the total volume of licensed water they abstracted:

$$Licence_{abstractedtotal}(m^3) = Licence_{abstractedforIWN}(m^3) + Licence_{abstractedforstorage}(m^3) \quad (E.41)$$

E.3.6 perform-preferred-behavioural-strategies

This final submodel is another farmer procedure where farmers' perform, if required, their in-season (short-term) water shortage and surplus behaviours as discussed in Section 6.2.1. These are dependent on farm type and whether behaviours are triggered in the model. However, first, farmers calculate how much water they have in total to satisfy their farm IWN:

$$FarmIWN_{totalavailable}(m^3) = Licence_{abstractedforIWN}(m^3) + Storage_{abstractedforIWN}(m^3) \quad (E.42)$$

where $FarmIWN_{totalavailable}$ equals the total water that the farmer has at their disposal. This is then divided equally between the two crops, if they require it:

$$PotIWN_{totalavailable}(m^3) = (100/FarmIWN_{total}(m^3)) * FarmIWN_{pot}(m^3) \quad (E.43)$$

$$PotIWN_{available}(m^3) = (PotIWN_{totalavailable}(m^3)/100) * FarmIWN_{totalavailable}(m^3) \quad (E.44)$$

$$VegIWN_{totalavailable}(m^3) = (100/FarmIWN_{total}(m^3)) * FarmIWN_{veg}(m^3) \quad (E.45)$$

$$VegIWN_{available}(m^3) = (VegIWN_{totalavailable}(m^3)/100) * FarmIWN_{totalavailable}(m^3) \quad (E.46)$$

where $PotIWN_{totalavailable}$ and $VegIWN_{totalavailable}$ are temporary variables which equal the volume of farm IWN available which is required by each crop; $PotIWN_{available}$ and $VegIWN_{available}$ equal the actual volume of water which is available for each crop.

Farmers then determine whether they have a shortage or surplus of water. If the total volume of water a farmer has at their disposal is less than their farm IWN, then they have a shortage:

$$Shortage(m^3) = FarmIWN_{total}(m^3) - FarmIWN_{totalavailable}(m^3) \quad (E.47)$$

If, however, the total volume of licensed water abstracted was less than their total licensed water available, then they have a surplus:

$$Surplus(m^3) = Licence_{available}(m^3) - Licence_{abstractedtotal}(m^3) \quad (E.48)$$

As all farmers preferred to only use their maximum abstraction licence to irrigate their most valuable crops, then they all perform the same strategy. In this model, their most valuable crop was considered the crop which covered the largest percentage of their irrigated area. Therefore, if a shortage existed, and main crop potatoes were the predominant crop, and their IWN was greater than the farm IWN available, then:

$$PotIWN_{applied}(m^3) = FarmIWN_{totalavailable}(m^3) \quad (E.49)$$

and;

$$VegIWN_{applied}(m^3) = 0 \quad (E.50)$$

where $PotIWN_{applied}$ equals the total volume of water available to satisfy IWN of main crop potatoes; and $VegIWN_{applied}$ equals the total volume of water available to satisfy IWN of vegetables (i.e. carrots). However, if main crop potatoes IWN was less than the farm IWN available, then:

$$PotIWN_{applied}(m^3) = FarmIWN_{pot}(m^3) \quad (E.51)$$

and;

$$VegIWN_{applied}(m^3) = FarmIWN_{totalavailable}(m^3) - PotIWN_{applied}(m^3) \quad (E.52)$$

Conversely, when vegetables (i.e. carrots) were the predominant crop, the same calculations were performed, and not repeated here, except the role of main crop potatoes and vegetables (i.e. carrots) were switched. Furthermore, if a water shortage did not occur then farmers divided their total water available equally between the two crops based on the percentage of their crop cover as previously discussed.

If a water surplus existed, then those who preferred LWUO preferred to just use their abstraction licence to meet crop water requirements and leave the remainder of their licence unused. Those who had no preference preferred to sell surplus water to maximise profits. Whilst those who preferred HWUO preferred to abstract surplus water for storage. However, as the latter strategy was performed automatically by all farm types, if farmers had storage available, those who preferred HWUO reverted to their second preference, to sell surplus water to maximise profits. Therefore, although nothing was to be reported for those who preferred LWUO, the quantity of water available for sale was recorded for those who had no preference and those who preferred HWUO:

$$Surplus_{available\ for\ sale}(m^3) = Surplus(m^3) \quad (E.53)$$

E.4 Results

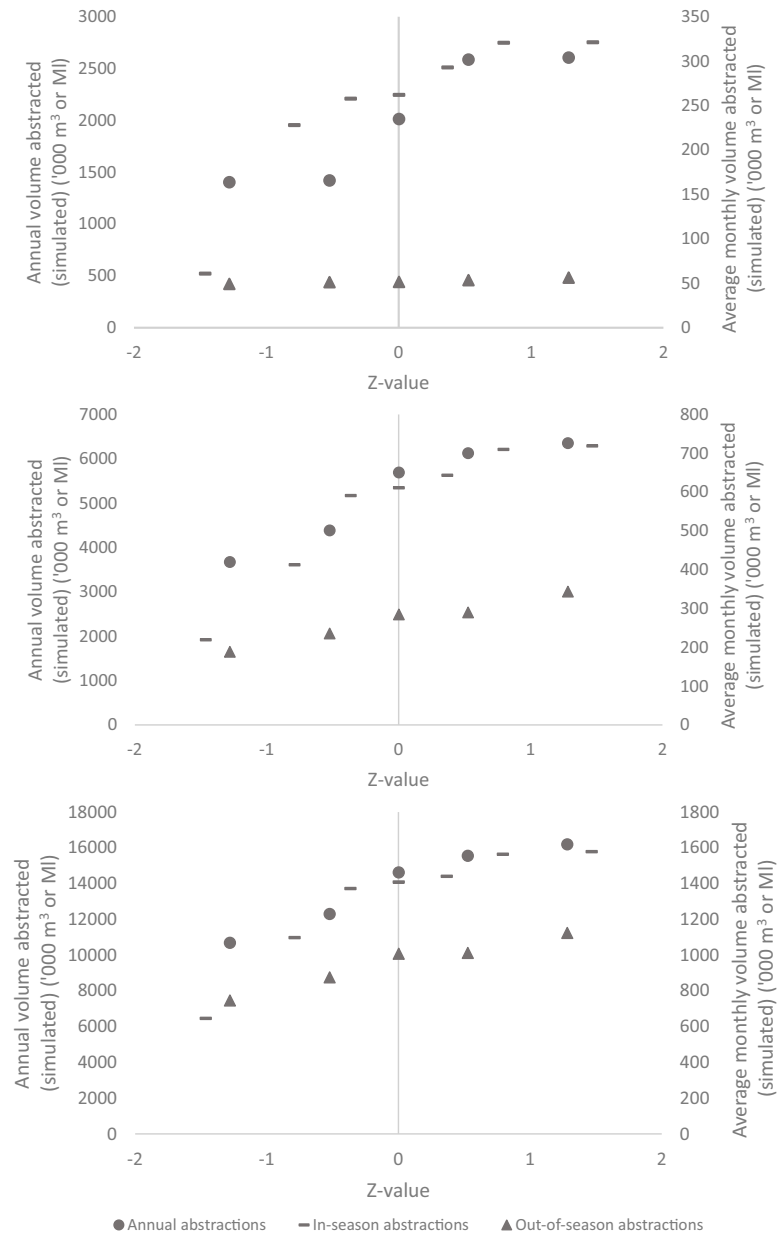


Figure E.3: Normal probability plots illustrating distribution with regards to annual, in-season, and out-of-season abstractions for each simulated farm type. (Top) A normal probability plot illustrating that annual, and out-of-season, abstractions were normally distributed, whilst in-season abstractions were not for those who preferred LWUO. This supports the results of the Shapiro-Wilk tests of normality as $p = >.05$ for both annual and out-of-season abstractions, whilst $p = <.05$ for in-season abstractions. (Middle) A normal probability plot illustrating that all of the variables were normally distributed for those who had no preference. This supports the results of the Shapiro-Wilk tests of normality as $p = >.05$ for all of variables. (Bottom) A normal probability plot illustrating that all of the variables were normally distributed for those who preferred HWUO. This supports the results of the Shapiro-Wilk tests of normality as $p = >.05$ for all of variables

Table E.7: System level patterns of abstraction behaviour based on the preferred behavioural intentions of each farm type under current (CS), basic (BS), and enhanced (ES) water allocation systems during the dry year (1983) and wet year (1988) climate scenarios

Scenario	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec	Annual
Licence volume abstracted (Ml)													
- CS-Dry	1,925	1,622	1,242	789	677	3,441	3,531	3,512	1,244	1,197	1,838	1,746	22,765
- CS-Wet	1,929	1,632	1,244	1,198	1,936	3,118	923	3,345	2,204	902	1,490	1,264	21,185
- BS-Dry	1,543	1,454	1,157	978	840	2,858	2,862	2,862	1,529	1,473	1,389	1,285	20,231
- BS-Wet	1,560	1,471	1,171	1,408	2,004	2,744	1,146	2,845	2,252	1,120	985	900	19,605
- ES-Dry	1,537	1,663	884	1,377	1,068	2,301	852	421	315	319	381	704	11,821
- ES-Wet	1,540	1,117	901	818	768	540	323	294	339	226	208	364	7,438
Surplus water available (Ml)													
- CS-Dry	40	342	722	2,756	2,869	104	14	34	2,301	2,348	126	219	11,876
- CS-Wet	36	334	721	2,346	1,608	426	2,621	199	1,341	2,642	476	702	13,453
- BS-Dry	1,319	1,408	1,705	1,884	2,022	4	0	0	1,333	1,389	1,473	1,577	14,114
- BS-Wet	1,317	1,406	1,706	1,468	873	133	1,731	31	625	1,757	1,892	1,977	14,917
- ES-Dry	1,331	1,701	1,335	2,770	2,949	2	0	0	272	276	330	609	11,576
- ES-Wet	1,336	1,050	1,172	465	67	0	285	0	0	198	182	320	5,074
Shortage water required (Ml)													
- CS-Dry	0	0	0	0	0	5,471	12,771	15,572	0	0	0	0	30,814
- CS-Wet	0	0	0	0	163	2,288	0	4,356	425	0	0	0	7,232
- BS-Dry	0	0	0	0	0	6,357	13,543	13,297	0	0	0	0	33,196
- BS-Wet	0	0	0	7	334	2,962	0	5,125	615	0	0	0	9,044
- ES-Dry	0	0	0	0	0	6,620	14,765	15,032	0	0	0	0	36,417
- ES-Wet	0	0	0	68	797	3,988	0	6,692	1,534	0	0	0	13,078
Farm irrigation water need satisfied (%)													
- CS-Dry	-	-	-	-	-	66	50	41	-	-	-	-	52
- CS-Wet	-	-	-	100	93	74	-	66	85	-	-	-	84
- BS-Dry	-	-	-	-	-	61	47	37	-	-	-	-	48
- BS-Wet	-	-	-	99	86	67	-	61	79	-	-	-	78
- ES-Dry	-	-	-	-	-	59	42	29	-	-	-	-	44
- ES-Wet	-	-	-	91	67	56	-	49	47	-	-	-	62

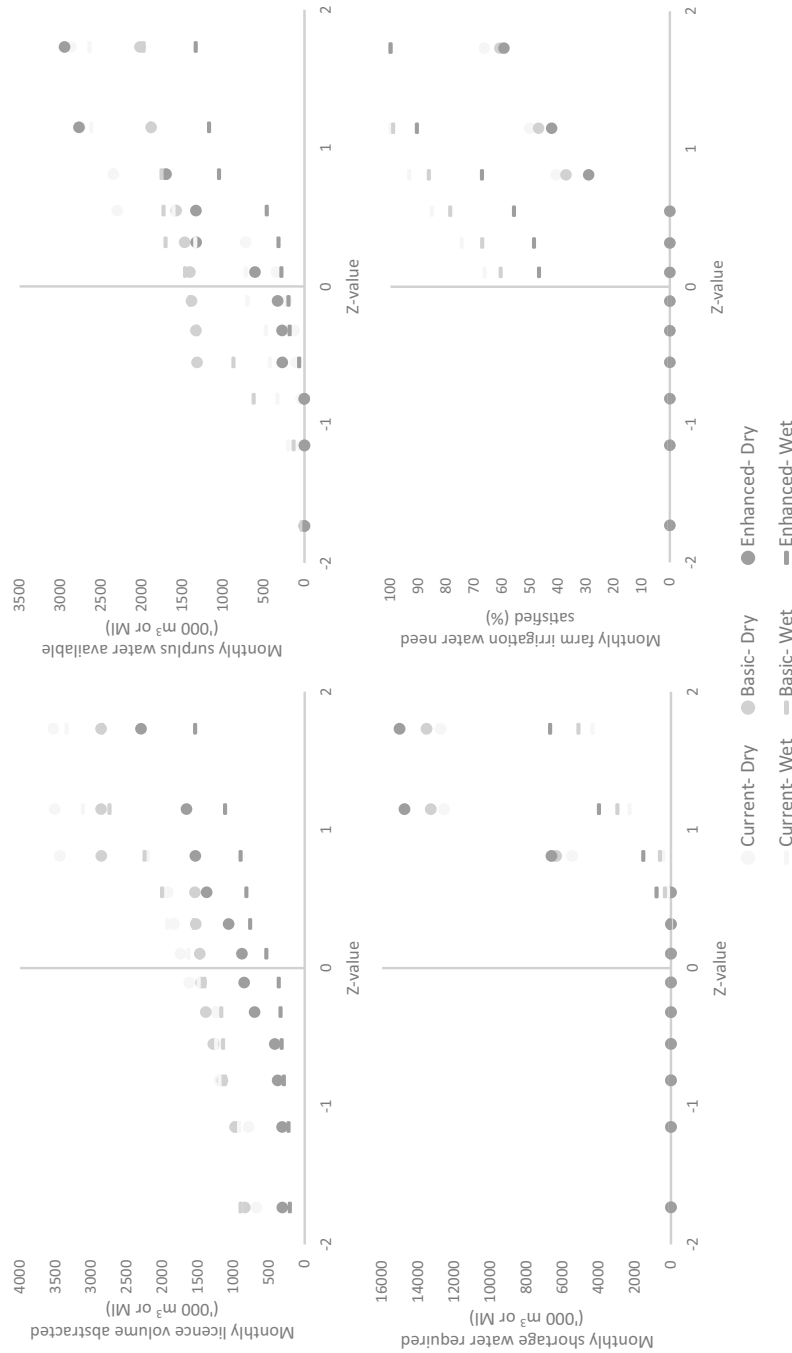


Figure E.4: Normal probability plots illustrating distribution with regards to monthly licence volume abstracted, surplus water available, shortage water required, and farm irrigation water need satisfied under each policy scenario for both dry and wet year climate scenarios. (Top left) A normal probability plot illustrating that licence volume abstracted was not normally distributed under any of the policy or climate scenarios. This supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ under the current and proposed basic water allocation systems during the dry year climate scenario, whilst $p = >.05$, but only marginally, under the remaining policy and climate scenarios. (Top right) A normal probability plot illustrating that surplus water available was not normally distributed under any of the policy or climate scenarios. This supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ under all policy scenarios during the dry year climate scenario and under the proposed enhanced water allocation system during the wet year climate scenario, whilst $p = >.05$, but only marginally, under the remaining policy or climate scenarios. (Bottom left) A normal probability plot illustrating that shortage water required was not normally distributed under any of the policy or climate scenarios. This supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for all policy and climate scenarios. (Bottom right) A normal probability plot illustrating that farm irrigation water need satisfied was not normally distributed under any of the policy or climate scenarios. This supports the results of the Shapiro-Wilk tests of normality as $p = <.05$ for all policy and climate scenarios

